2012 Annual Report

Horse Creek Stewardship Program

E213008501 2000

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Introduction

This is the tenth annual report summarizing the status of the Horse Creek Stewardship Program (HCSP). After a series of legal challenges to the required permits, the Mosaic Company (Mosaic) and the Peace River Manasota Regional Water Supply Authority (PRMRWSA) executed a settlement agreement to ensure that mining would not have negative impacts on Horse Creek, a major tributary of the Peace River, as a result of proposed mining activities by Mosaic in eastern Manatee and western Hardee Counties, Florida. A principal component of the agreement was the creation of the HCSP. The overall goals of the HCSP are to ensure that Mosaic's mining activities do not interfere with the ability of the PRMRWSA to withdraw water from the Peace River for potable use nor adversely affect Horse Creek, the Peace River, or Charlotte Harbor. The program, which is funded and managed by Mosaic, has two purposes: 1) in order to detect any adverse conditions or significant trends that may occur as a result of mining, the HCSP provides a protocol for the collection of information on the physical, chemical, and biological characteristics of Horse Creek during Mosaic's mining activities in the watershed, and 2) if detrimental changes or trends caused by Mosaic's activities are found, the HCSP provides mechanisms for corrective action. The program is limited to the investigation of the potential impacts of Mosaic mining activities on the Horse Creek Basin and is not intended to investigate the potential impacts of other land uses or mining activities by other entities.

This program has three basic components: 1) monitoring and reporting on stream quality, 2) investigating adverse conditions or significant trends that are identified through monitoring, and 3) implementing corrective action for adverse changes to Horse Creek caused by Mosaic's mining activities. The HCSP is unique in that it does not rely solely upon the exceedance of a standard or threshold to bring about further investigation and corrective action, where appropriate. The presence of a significant temporal trend alone will be sufficient to initiate such steps. This program offers additional protection to Horse Creek; this protection is not usually present in the vast majority of regulatory scenarios.

Monitoring for the HCSP began in April 2003, and this report, which is the tenth in a series of Annual Reports, presents the results of the first ten years of monitoring, including historical data since 1990. Approximately 12,000 acres of land in the Upper Horse Creek Basin had been mined at the time the HCSP was initiated; about 10,000 acres of the total 12,000 acres mined are located upstream of all HCSP monitoring stations on land controlled by Mosaic, with the remaining mined area on other parties' lands lying upstream of all but the northernmost monitoring location.

Recent Mining and Reclamation

A total of 76 acres was mined in the Horse Creek Basin at the Mosaic Fort Green Mine in 2012. Some additional phosphate mining may or may not have been conducted by other companies in the Horse Creek drainage basin in 2012, but Mosaic is not aware of the extent or timing of that mining. In 2012, there was a total of 523 acres planted in the Horse Creek Basin (254 in the West Fork Horse Creek Basin and 269 in the Horse Creek Basin). Tailing/grading activities also occurred throughout the year in the West Fork Horse Creek Basin totaling 122.5 acres. There were 600 acres of land reclaimed to final contour.

Monitoring Program Components

Four locations on Horse Creek were monitored for physical, chemical, and biological parameters; two of these sites are also long-term US Geological Survey (USGS) gauging stations. Water quantity data were collected continuously from the USGS gauging stations. Rainfall data were collected daily from three

Mosaic rain gauges located in the Horse Creek Basin. Water quality data were collected during monthly sampling events, continuously from one Horse Creek location, and during biological sampling events. Biological (fish and benthic macroinvertebrates) sampling events are scheduled to occur three times each year.

Water Quantity Results

Although low and median Horse Creek discharge in 2012 was average for the region, rainfall in 2012 was below the long-term average annual rainfall of 52.72 inches (1908-2012). For 2012, temporal patterns of average daily stream flow and stage were similar across all stations, with the majority of high flows and stages occurring during May to August, during the rainy season. The dry season (January to May 2012) was extremely dry, resulting in periods of little to no streamflow and no NPDES discharge. Summer rains began in late-May/early-June, with lag in streamflow response until mid-June. At the beginning of the wet season, the lag effect between rainfall and streamflow is more apparent as the surrounding wetlands or small creeks either need to fill or resume flow. However, later in the wet season when the wetlands and small creeks are full, the lag is much shorter. Higher streamflow continued through the end of October/first of November.

In September and October, NPDES contributed up to 75 percent of the streamflow at HCSW-1 compared to rainfall; in late October 2012, NPDES discharge accounted for almost all of the streamflow at HCSW-1. NPDES discharge from August to November was also a lagged response to rain that occurred from late-May to early October 2012; NPDES discharge from the Horse Creek outfalls usually does not occur until sufficient water storage accumulates in the circulation system, resulting in a lag. In general, the lower than average rainfall resulted in lower than average streamflow in Horse Creek, with some lags. There is no evidence that mining and reclamation activities in the basin caused any significant decrease in total streamflow in 2012. In a previous study (Robbins and Durbin 2011) that compared streamflow during dry years in reference and potentially impacted streams before and during phosphate mining, there was no evidence that phosphate mining practices caused lower monthly flows in the potentially impacted streams (including Horse Creek) than what would be expected given the conditions in a reference stream (Charlie Creek).

Water Quality Results

Water quality parameters in 2012 were almost always within the desirable range relative to trigger levels and water quality standards at the station closest to mining (HCSW-1, Table 18). Trigger levels were exceeded only once at HCSW-1 in 2012 with the alkalinity exceedance in November. At HCSW-2, trigger levels were exceeded for dissolved oxygen during half of the year (Table 18). Based upon historical conditions in Horse Creek (Durbin and Raymond 2006), the reported values for dissolved oxygen at HCSW-2 are the result of natural conditions (proximity to hypoxic segment of stream – Horse Creek Prairie) and are not related to mining activities. The chlorophyll a trigger level was exceeded during low-flow periods at HCSW-2 in February, April, May, and December, and the pH declined below the acceptable trigger level range in October. HCSW-3 exceeded trigger levels for dissolved oxygen (June-September), calcium (April), sulfate (February-April and June), and TDS (January-April and June). HCSW-4 exceeded trigger levels for specific conductivity (June), dissolved oxygen (July and September), calcium (April-June), iron (July-October), alkalinity (May), sulfate (March-April and June), and TDS (January-June). Dissolved oxygen triggers were exceeded during summer wet months of 2012, when high temperatures reduce the oxygen carrying capacity of the stream. Sulfate, calcium, TDS, and other ions were exceeded in the dry season, when low rainfall and streamflow likely led to increased groundwater inputs from baseflow and agricultural runoff. Dissolved iron concentrations consistently exceeded the trigger value set at HCSW-4, but Mosaic and the PRMRWSA agree that the trigger value at that station has been set too low given historical and upstream concentrations of dissolved iron. Based on impact assessments already completed, none of the observed exceedances pose a significant adverse ecological impact to Horse Creek that would be attributable to mining.

Several of the trends detected during the statistical analysis either have an estimated slope that 1) was not in the direction of an adverse trend (color, ammonia, and dissolved iron) or 2) was very small compared to limits in laboratory prediction or the observed differences between primary and duplicate field samples (pH, ammonia, orthophosphate). For the trends with higher estimated rates of change (specific conductivity and various dissolved ions), the potential trends are discussed in Appendix I. Orthophosphate was also discussed further in this impact assessment as a follow-up to the trend impact assessment of the 2010 Annual Report (Robbins et al. 2014). Appendix I shows that the apparent trend in orthophosphate from 2003-2012 is caused by a data bias, and that extending the period of record into pre-2003 data eliminates this trend. Specific conductivity and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful.

Significant differences between stations were evident for several parameters. Overall, HCSW-2 was the most dissimilar from the other three stations, especially in pH, dissolved oxygen, and some dissolved ions. Some nutrients (nitrate + nitrite) and dissolved ions (specific conductivity, calcium, chloride, sulfate) had higher concentrations downstream in Horse Creek, probably because of increased groundwater seepage and agricultural runoff in the lower Horse Creek basin. Differences in topography, geology, and land use that could account for these trends in Horse Creek are examined in the Horse Creek Stewardship Program Historical Report (Durbin and Raymond 2006).

Water quality parameters were compared with water quantity variables recorded during the same month: average daily streamflow for the month, average daily NPDES discharge for the month, and total monthly rainfall. In general, pH, dissolved oxygen, and most dissolved ions are higher when the overall quantity of water in the Horse Creek system is low. (In this report, specific conductivity, calcium, alkalinity, sulfate, and TDS showed the opposite pattern with NPDES discharge at HCSW-1; possible reasons behind this are discussed as part of the trend analysis in Appendix I.) Conversely, turbidity, color, iron, and nitrogen are high when the water quantity is also high. When water quantity in Horse Creek is low, the stream may be pooled or slow-moving, leading to algal blooms that may increase pH and chlorophyll a. In addition, the majority of water in the stream during low quantity periods may be from groundwater (seepage or agricultural runoff); groundwater has a higher concentration of dissolved ions than surface water. When water quantity is high, an increased amount of sediment and organic debris is washed into the stream, leading to increases in turbidity, color, iron, and nitrogen.

Benthic Macroinvertebrate Results

Benthic invertebrate habitat assessment scores were “Optimal” to “Sub-optimal” and SCI scores were “Healthy” or “Exceptional” at all stations in 2012; these scores are typical of southwestern Florida streams, including those used to develop the Habitat Assessment and SCI indices. Benthic invertebrate taxa diversity and SCI metrics in Horse Creek exhibit both seasonal and year-to-year variation. Overall, taxa diversity indices and SCI metrics show few monotonic trends over time and are very similar between sampling events and stations. However, HCSW-2 has slightly lower diversity and significantly lower SCI metrics than other stations. Habitat conditions at HCSW-2 are consistently poor, with lower streamflow, dissolved oxygen, and pH than other Horse Creek stations. The source of the poor conditions are related to the lower than average streamflow and rainfall of the previous few years and the presence of Horse Creek Prairie, the large marsh located upstream of the biological sampling station.

Fish Results

During 2012, 25 species of fish were collected from the four Horse Creek sampling stations. In 2012, one new fish species was collected at HCSW-4, the Orinoco sailfin catfish. Fewer fish species were collected at HCSW-1 during two sampling events in 2012 than the other stations because of the unique characteristics of that sampling location (Table 21, Figures 93 - 96). In addition, water levels and streamflow were fairly high during biological sampling at all stations in October which led to HCSW-2 not being sampled; higher water levels did not allow for some habitats to be reached by our sampling equipment. Abnormally cold winters in 2009 – 2010 and 2010 – 2011 may have led to decreased fish diversity at some stations in 2010 and 2011, with evidence of recovery and recruitment in 2012. Over the period of record, fish richness and diversity was lowest at HCSW-2, with no significant annual trends. Fish communities were similar for all years when stations were combined and for all stations when years were combined. Catch per effort is variable over time and dependent on sampling technique, a station's physical characteristics, water levels, and available recruitment sources. No trends were evident in the abundance of fish from exotic and native fish groups.

Conclusions

Although this report covers only the tenth year of an ongoing monitoring program, some general conclusions can be drawn. Expected relationships between rainfall, runoff and streamflow were observed in the 2003 to 2012 water quantity data. Program trigger levels were exceeded for several parameters in 2012 and several parameters had statistically significant trends from 2003 to 2012, but the exceedances and trends are not of immediate concern (Appendix I). The benthic macroinvertebrate and fish communities found in Horse Creek in 2003 to 2012 were typical of those found in a Southwest Florida stream.

Recommendations

During the TAG meeting for the 2011 and 2012 draft Annual Reports (October 29, 2014), the following recommendations were made:

- Cardno will add a graphic to the 2011 and all subsequent reports showing more historical flow in Horse Creek extending beyond the HCSP time period (back to at least 1978 to match the double mass curve).
- Cardno will add more discussion of the double mass curve to the 2011 and all subsequent reports.
- Cardno will add a Trend Summary Table to the 2011 Annual report and all subsequent reports instead of only including in Appendix I.
- In the 2013 Annual Report, Cardno will add an appendix that lists any erroneous data and remove problematic data from Appendix C graphs. Any future erroneous data will be added to the appendix as the program continues.
- For the 2013 Annual Report TAG Meeting, Cardno will include a discussion on the DO Saturation standard and implications in the main body of the report.
- In the 2014 annual report, Cardno will add a discussion on NNC standards and implications in the main body of the report.
- In the 2014 annual report, Cardno will add a discussion on legacy CF Industries operations in the Horse Creek Basin.

1 Introduction

As a result of proposed mining operations by Mosaic Fertilizer, LLC (Mosaic) in eastern Manatee and western Hardee Counties, Florida, and a series of legal challenges to the permits required for such mining, Mosaic and the Peace River Manasota Regional Water Supply Authority (PRMRWSA) executed a settlement agreement structured to ensure that mining would not have negative impacts on Horse Creek, a major tributary of the Peace River. A principal component of that agreement was the creation of the Horse Creek Stewardship Program (HCSP), which is funded and managed by Mosaic. The program document, as referenced in the settlement agreement, is provided as Appendix A.

There are two purposes for the HCSP. First, it provides a protocol for the collection of information on the physical, chemical, and biological characteristics of Horse Creek during Mosaic's mining activities in the watershed. This information would then allow the ability to detect any adverse conditions or significant trends that may occur as a result of mining. Second, it provides mechanisms for corrective action with regard to detrimental changes or trends caused by Mosaic's activities, if any are found. The program is limited to the investigation of the potential impact of Mosaic mining activities on Horse Creek Basin and is not intended to investigate the potential impacts of other land uses or mining activities by other entities.

The overall goals of the program are to ensure that Mosaic's mining activities do not interfere with the ability of the PRMRWSA to withdraw water from the Peace River for potable use nor adversely affect Horse Creek, the Peace River, or Charlotte Harbor. There are three basic components to the HCSP: 1) monitoring and reporting on stream quality, 2) investigating adverse conditions or significant trends identified through monitoring, and 3) implementing corrective action for adverse stream quality changes attributable to Mosaic's activities. An important aspect of this program is that it does not rely solely upon the exceedance of a standard or threshold to bring about further investigation and, where appropriate, corrective action. The presence of a significant temporal trend alone is sufficient to initiate such steps. This protection mechanism is not present in the vast majority of regulatory scenarios.

In brief, the HCSP provides for the following data collection:

- Continuous recording (via USGS facilities) of stage and discharge at two locations on the main stem of Horse Creek
- Daily recording of rainfall via Mosaic and USGS rain gauges in the upper Horse Creek basin
- Continuous recording of temperature, dissolved oxygen, conductivity, turbidity and pH at the Horse Creek station nearest to Mosaic's active mining operations
- Monthly water quality monitoring of 21 parameters at four stations on the main stem of Horse Creek[1]
- Sampling of fish, benthic macroinvertebrates and field water quality parameters (temperature, dissolved oxygen, conductivity, turbidity and pH) three times annually at four stations on the main stem of Horse Creek

HCSP monitoring began in April 2003. At the time the HCSP was initiated, some 12,000 acres of land in the Upper Horse Creek Basin had been mined, about 10,000 acres of which lie upstream of all HCSP monitoring stations on land controlled by Mosaic, with the remaining mined area on other parties' lands lying upstream of all but the northernmost monitoring location. In 2012, 76 acres were mined in the Horse Creek Basin upstream of the northernmost monitoring location (Figure 1). Water quantity data are

[1] In 2009, the list of parameters was reduced by three (total amines, total fatty acids, and FL-PRO were removed), and an additional station on Brushy Creek (tributary of Horse Creek) was added.

collected essentially continuously, water quality data are collected monthly, and biological data (fish and benthic macroinvertebrates) are collected three times annually (March - April, July - September and October - December). Specific months when biological sampling occurs may change from year to year to avoid very low or very high flows, which would impede representative sampling.

This report, which is the tenth in a series of Annual Reports, presents the results of monitoring conducted from April 2003 through 2012. All data presented in tables and figures was collected as part of the HCSP unless otherwise noted. Additional sources of data since 1990 have also been included in the box plots to provide a short historical perspective. A separate report contains a review and summary of all available historical water quality and biological information for Horse Creek (Durbin and Raymond 2006).

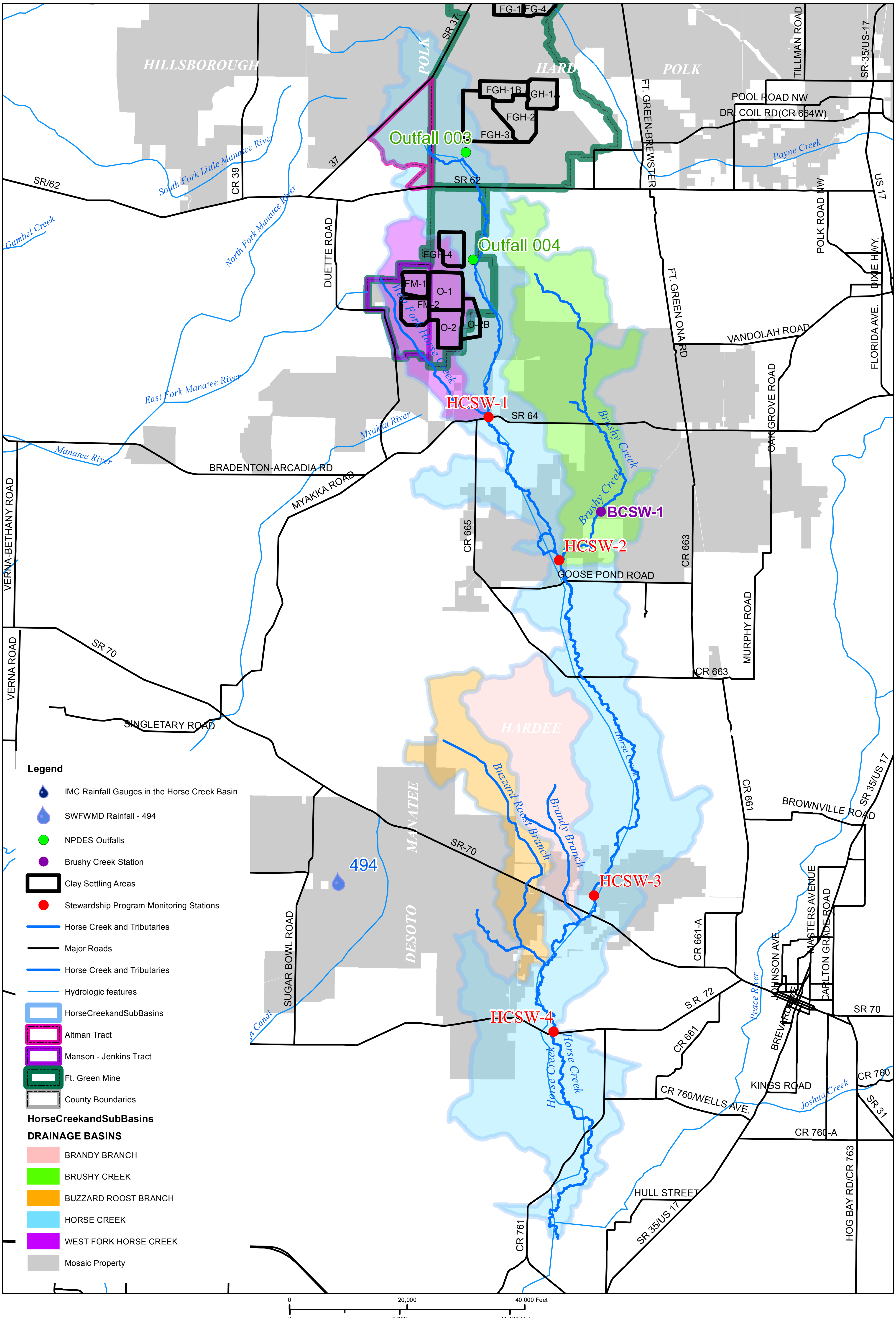

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Figure 1. Overview of drainage basins, HCSP sampling locations, and Mosaic property in the Horse Creek Basin.

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NAD 1983 UTM Zone 17N feet

Date: 12/19/11 Rev. Date: 07/07/14 PM: KMNR GIS Analyst: KMNR Map Document: horsecreek_overview_2011_Figure1.mxd Project Number: 7729-110 PDF Document: horsecreek_overview_2011_Figure1.pdf Plot Size: 11 x 17

2 Description of Horse Creek Basin

The Horse Creek basin is located in five counties of South-Central Florida: Hillsborough, Polk, Manatee, Hardee, and Desoto, with the majority of the watershed spanning portions of western Hardee and DeSoto Counties (Figures 1 and 2). Horse Creek is a major tributary of the Peace River that drains into the southwestern portion of the Peace River Basin and supplies approximately 15 percent of the surface water runoff to the Peace River (Lewelling 1997).

The basin occupies some 241 square miles, and the length of the channel is approximately 43 miles. Horse Creek has an elongated basin with a north-to-south drainage that is influenced by the general topography of the area. Six sub-basins and five tributaries make up the Horse Creek Basin. West Fork Horse Creek and Brushy Creek, two northern tributaries in the Polk Uplands, are generally straight, at least partially channelized, and have relatively rapid flows (Lewelling 1997). The remaining tributaries, occupying the central to southern Horse Creek Basin, include Buzzard Roost Branch and Brandy Branch. These lower reaches are located in the DeSoto Plains/Gulf Coast Lowlands area and are generally meandering, slower streams. Horse Creek ultimately discharges into the Peace River near Fort Ogden (SWFWMD 2000).

The topography of the Horse Creek basin generally follows the north-to-south drainage flows of the creek. Elevation in the basin ranges from 135 feet in the north to 30 feet in the south near the confluence of Horse Creek and the Peace River. The basin is located in the mid-peninsular physiographic zone of Florida, in three subdivisions: Polk Uplands, DeSoto Plains, and Gulf Coast Lowlands. The Polk Uplands underlie the northern portion of the Horse Creek Basin, where the elevation generally exceeds 100 feet NGVD. In this location, the channel of Horse Creek is generally steep and slightly incised, with swiftly moving water. The central Horse Creek basin is located in the DeSoto Plain. Average elevations in this area range from 30 to 100 feet NGVD. Where Horse Creek enters the Peace River, the Gulf Coast Lowlands range in elevation from about 30 to 40 feet NGVD. The Horse Creek channel in the Desoto Plain and Gulf Coast Lowlands is slower and more sinuous than the northern channel (SWFWMD 2000, Lewelling 1997).

The northern Horse Creek Basin is located in the Polk Uplands, with Pomona-Floridana-Popash soils characterized by nearly level, poorly drained, and very poorly drained sandy soils. Some soils in this association have dark colored subsoil at a depth of less than 30 inches over loamy material, and some are sandy to a depth of 20 - 40 inches and are loamy below. The extreme northern basin of Horse Creek contains isolated areas of the Arents-Hydraquents-Neilhurst soils group, parts of which have been strip-mined for phosphate (Robbins et al. 1984).

The central and southern Horse Creek Basin is located in the DeSoto Plain, which is a very flat, submarine plain probably formed under Pleistocene Wicomico seas, 70 to 100 feet above present sea level (Cowherd et al. 1989). The Smyrna-Myakka-Ona and Smyrna-Myakka-Immokalee soil associations characterize this portion of the Horse Creek Basin with flat, poorly drained soils that are sandy throughout (Lewelling 1997). The soil group Bradenton-Felda-Chobee is also located immediately adjacent to the main channel of Horse Creek, from below State Road 64 to just above the mouth of the creek. These soils are characterized by nearly level, poorly drained and very poorly drained soils that are sandy to a depth of 20 to 40 inches and underlain by loamy material or that are loamy throughout and subject to frequent flooding. The dominant soil groups in the Horse Creek basin are generally poorly drained, reducing the infiltration of rainwater to the water table in the surficial aquifer, thereby limiting the amount of water available to support baseflow (SWFWMD 2000).

The climate of Horse Creek Basin is subtropical and humid with an average temperature of about 72 ºF. Summer temperatures average 80 ºF, and winter temperatures average 60 ºF (Hammett, 1990). The average daily temperatures in Hardee County, in the northern Horse Creek Basin, range from is 52 ºF to 91 ºF (Robbins et al. 1984). The average daily temperatures in DeSoto County, in the southern Horse Creek Basin, range from 49 ºF to 92 ºF. Average relative humidity in Horse Creek Basin ranges from 57 percent in the mid-afternoon to 87 percent at dawn. The prevailing wind is from the east-northeast, with the highest average wind speed, 7.8 mph, occurring in March (Cowherd et al. 1989).

The average annual rainfall in the Peace River Basin, which includes Horse Creek, is 52.72 inches, with more than half of that falling during localized thundershowers in the wet season (June – September)[2]. Rain during fall, winter, and spring is usually the result of large, broad frontal systems instead of local storms. November is typically the driest month of the year, averaging 1.75 inches over the historic period from 1908 to 2012. The months of December, January, and April are also characteristically dry, averaging 1.87, 2.16, and 2.39 inches, respectively. Dry conditions coincide with high evaporation rates and generally result in the lowest stream flows, lake stages, and ground-water levels of the year (Hammett, 1990). The wettest month of the year is typically June, averaging 8.44 inches followed by August with 8.43 inches.

Horse Creek flows through a generally rural area. Major land use activities in the basin are primarily agricultural, with extractive mining activities occurring in the northern part of the basin. Agricultural activities include cattle grazing, row crop farming, citrus grove production, sod farming, and conversion of native lands to pasture for both cattle grazing and hay production.

Small rural agricultural communities are located in and near the Horse Creek drainage basin including Fort Green, Ona, and Myakka Head in the northern portion of the basin, Limestone, Lily, and Edgeville in the approximate center of the basin, and Arcadia, Fort Ogden and Nocatee near the southern end of the basin (Post et al. 1999). Generally, the northern Horse Creek basin is covered more by natural vegetation, while the southern basin is covered mostly by pasture and row crops (SWFWMD 2000).

Total acreages in each land cover type and proportions of the various land uses differ between regions of the basin. The percent mining cover has increased between 1988 and 2009, according to SWFWMD landuse maps for those years. The majority of land newly identified as mined in 2009 was agricultural or rangeland in 1988. Mining is the primary land use above State Road 64, but the percentage of land devoted to mining decreases rapidly downstream. Agricultural land use more than doubles in acreage from above County Road 663 (HCSW-2) to above SR 72 (HCSW-4). Rangeland covers about the same percentage of land in the northern part of the basin and in the southern portion. The percent upland forest and wetland cover also remains relatively constant in upstream and downstream sections of the creek.

Water quality sampling on Brushy Creek was newly added to the HCSP in 2009. Landuse in 2009 in the Brushy Creek basin is primarily agricultural (38%), with a relatively small percentage of mining (6%) compared to Horse Creek above State Road 64 or County Road 663. Overall, the Brushy Creek basin has a similar percentage of rangeland (15%), upland forest (13%), and wetland (29%) landuse as does the Horse Creek Basin.

[2] Historical rainfall information came from the following stations and years: 1908-1943 NOAA station 148, 1944-2012 average of NOAA station 148 and 336

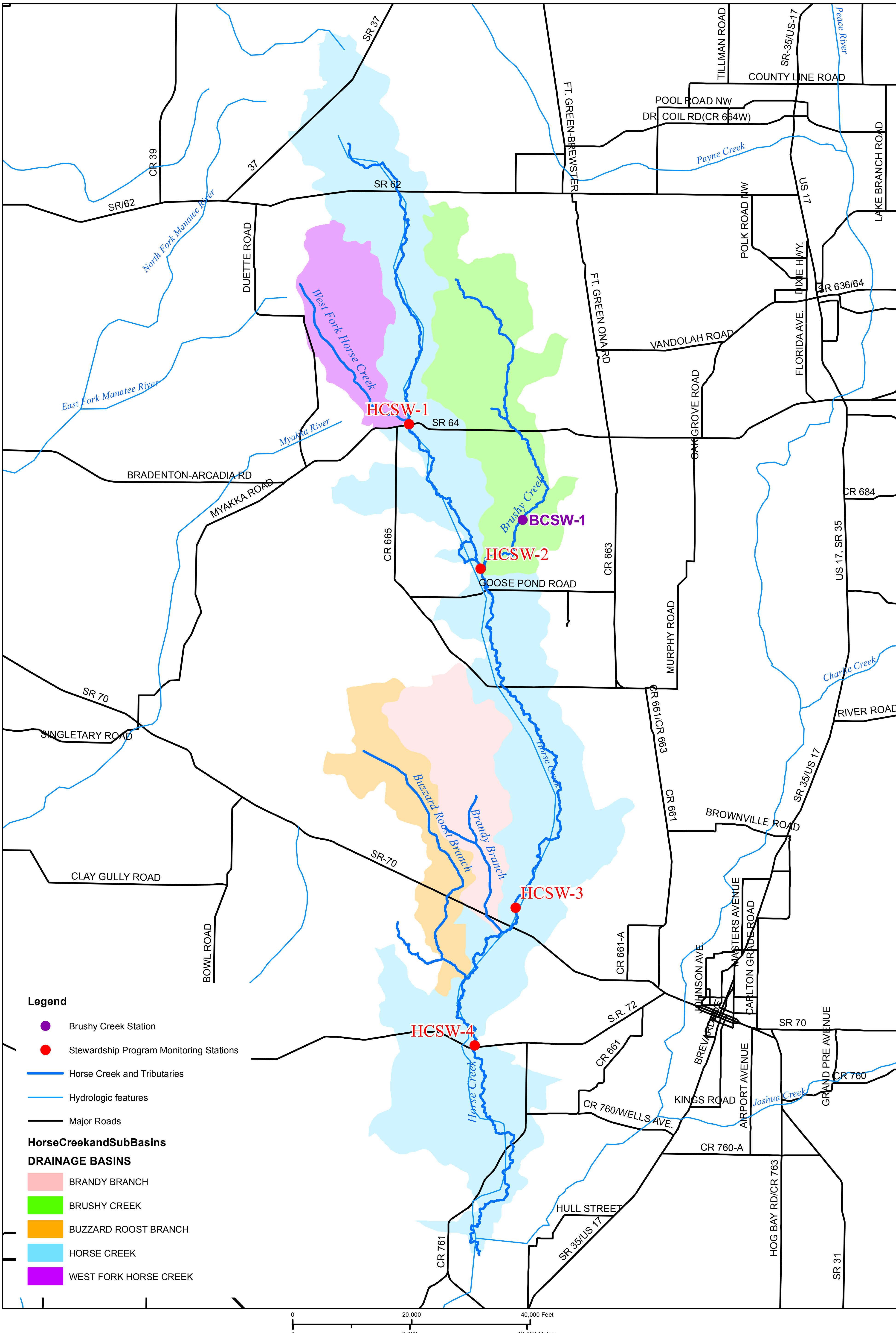

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Figure 2. Aerial photograph of the Horse Creek Basin and HCSP sampling locations

Image:2010

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3 Summary of Mining and Reclamation Activities

3.1 Mining

A total of 76 acres was mined in the Horse Creek Basin at the Mosaic Fort Green Mine in 2012 (Figure 3). A summary of all mining and reclamation activities from 2004 to 2012 is provided below in Table 1. There have been, and will be in the future, mining activities in the Horse Creek Basin outside of those performed by Mosaic. Some additional phosphate mining may or may not have been conducted by other companies in the Horse Creek drainage basin, but Mosaic is not aware of the extent or timing of that mining. Information on pre-mining conditions in the Horse Creek Basin may be found in an Environmental Impact Statement prepared by Environmental Science and Engineering, Inc. (1982) and a Development of Regional Impact statement prepared by Ardaman and Associates and colleagues (1979).

Table 1 lists mining and reclamation data for the Horse Creek Basin over the course of the HCSP (omitting the partial year of 2003). The table lists the acres mined, the acres reclaimed to the final contour (but not necessarily vegetated), and the acres released and reconnected to Horse Creek. Areas contoured have been restored hydrologically, but may not have fully achieved the required vegetative cover. The table does not include the acres revegetated because the same areas could potentially be revegetated more than once if less than ideal climate conditions result in plant loss.

Table 1. Total Acres mined, reclaimed to final contour, and reconnected to Horse Creek by Mosaic in the Horse Creek Basin from 2004 through 2012.

Year	Acres Mined	Acres Reclaimed to Final Contour	Acres Reconnected to Horse Creek
2004	638	30	0
2005	590	205	38
2006	187	0	205
2007	0	106	0
2008	150	245	66
2009	137	711	315
2010	283	270	0
2011	100	114	0
2012	76	600	0

There are three clay settling areas (CSAs) in the Horse Creek Basin at the Fort Green Mine. The FGH-3 clay settling area is located predominantly in Sections 5, 8, and 9, T33S, R23E. Construction of clay settling area FGH-3 was completed in 1999, and it was immediately put into service. The settling area was designed by Ardaman & Associates with a crest elevation of 151 feet NGVD, and a final pool elevation of 146 feet NGVD. The effective area of the CSA is approximately 933 acres. Three decant spillways, two on the west wall and one on the north wall, were designed to return water to the Ft. Green plant. Flow can also be directed to the south, using the 003 outfall, through spillways located in the return water ditch near the southwest corner of FGH-3. Clays are introduced into the settling area approximately midway on the east wall.

The FGH-4 clay settling area is located predominantly in Section 31, T33S, R23E. Construction of the CSA was completed in 2001, and it was put into service shortly thereafter. The settling area was designed by Ardaman & Associates with a crest elevation of 164.0 feet NGVD, and a final pool elevation of 159.0 feet NGVD. The effective area of the CSA is approximately 415 acres. Two decant spillways, one on the north wall, and one on the south wall were designed to return water to the Ft. Green central screening station (the smaller beneficiation plant located on SR39). Decant spillways located in the south return water ditch also have the capability of discharging water to the 004 outfall. Clay slurry is introduced into the settling area at the southwest corner, and at a point approximately midway on the west wall of the dam. The settling area is also used to store mine pit water, which is pumped in at the northwest corner and at approximately the center of the south wall.

The third settling area, Fort Meade-1 (FM-1) is located predominately in Section 1, T34S, R22E. FM-1 was constructed in 2006-2007 and put into service in March 2009. The settling area was designed by Ardaman and Associates with a crest elevation of 164 feet NGVD and a final pool elevation of 159 feet NGVD. The effective area of the CSA is approximately 350 acres. Two decant spillways, both located on the west wall of the dam, are designed to return water to a holding area at the base of the dam, which is then pumped via pipeline back to the Wingate Creek Mine. Water from this dam can also be routed via a series of pumps and surface water conveyance ditches to the 004 outfall; thus discharges from the clay settling area can be routed to either the Myakka or Horse Creek basins. The FGH-3, FGH-4, and FM-1 settling areas have real-time monitoring of the pond level, which is relayed to the PRMRWSA. Any sudden drop in pond level elevations, suggesting a substantial release of wastewater from the settling areas, would be detected promptly, allowing for an expedited response to the situation.

3.2 Reclamation

Reclamation of lands that have been mined is an ongoing process at Mosaic’s Fort Green Mine including lands in the Horse Creek Basin. The reclamation process consists of backfilling of the mined excavations with sand “tailings” produced as a by-product of the phosphate production process or shaping existing deposits of overburden material to bring the ground surface up to rough grade. Overburden material is spread over the backfilled areas and the areas are brought to the required final contours. Planting of both upland and wetland communities is done with appropriate species. Reclaimed areas are monitored and supplemental plantings are done as necessary until the revegetation of the land is successful. In 2012, there was a total of 523 acres planted in the Horse Creek Basin (254 in the West Fork Horse Creek Basin and 269 in the Horse Creek Basin). Tailing/grading activities also occurred throughout the year in the West Fork Horse Creek Basin totaling 123 acres. The number of acres reclaimed to the final contour and the acres reconnected to Horse Creek are summarized in Table 1 above.

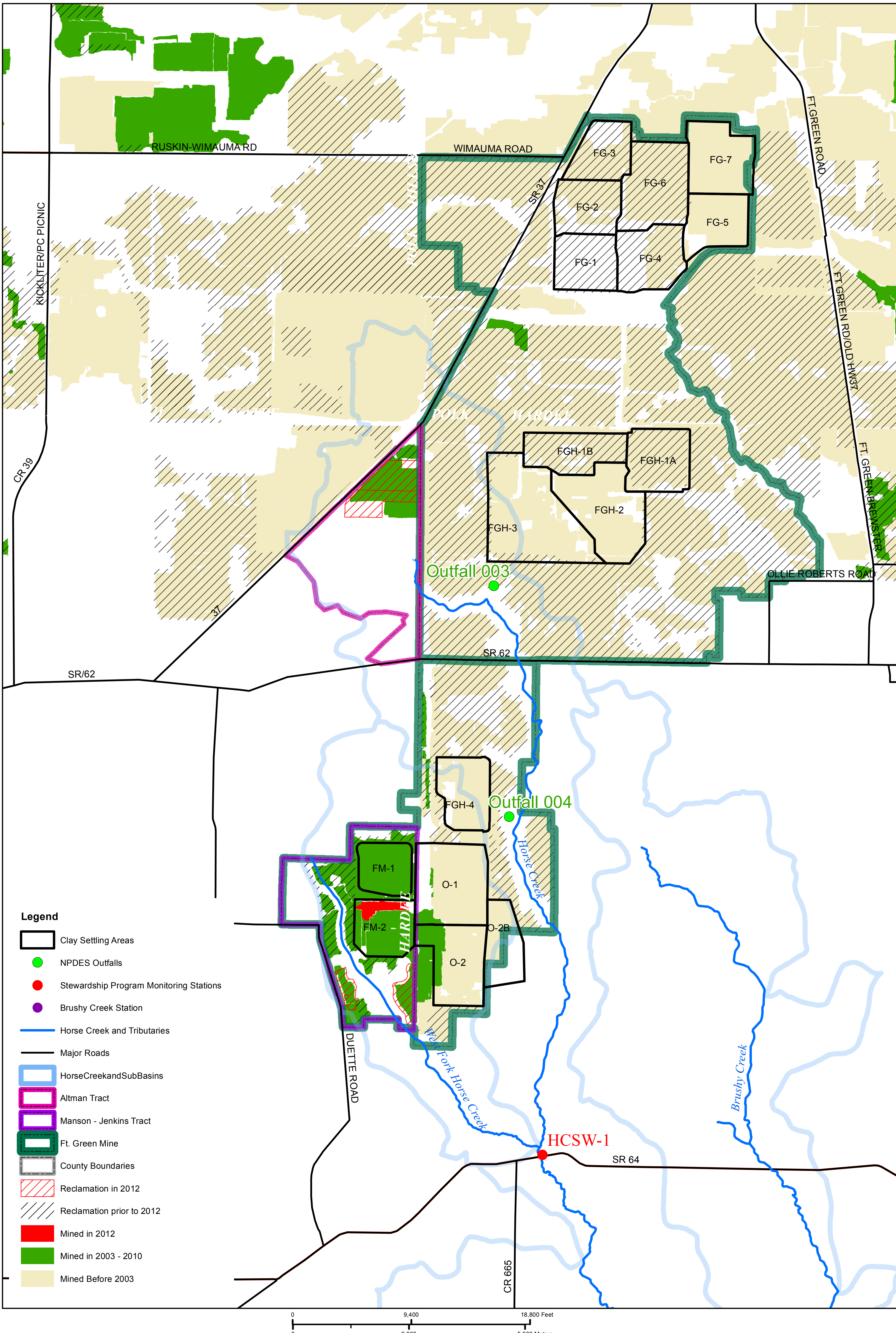

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Figure 3. Mining and reclamation areas in the Horse Creek Basin

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Coordinate System:
NAD 1983 UTM Zone 17N feet

Date: 12/19/11 Rev. Date: 07/07/14 PM: KMNR GIS Analyst: KMNR Map Document: horsecreek_mined_2011_figure3.mxd Project Number: 7729-110 PDF Document: horsecreek_mined_2011_figure3.pdf Plot Size: 11 x 17

4 Methods

4.1 Station Locations and Sampling Schedule

Four Horse Creek locations are monitored for physical, chemical, and biological parameters (Figure 1):

HCSW-1 - Horse Creek at State Road 64 (USGS Station 02297155)
HCSW-2 - Horse Creek at County Road 663A (Goose Pond Road)
HCSW-3 - Horse Creek at State Road 70
HCSW-4 - Horse Creek at State Road 72 (USGS Station 02297310)

As indicated above, HCSW-1 and HCSW-4 are also long-term US Geological Survey (USGS) gauging stations, with essentially continuous stage and discharge records since 1977 and 1950, respectively. Water quality sampling has been conducted monthly beginning in April 2003, while biological sampling events have been conducted typically three times per year (Table 2).

In September 2009, based on recommendations of the PRMRWSA and the Technical Advisory Group (TAG), Mosaic began sampling water quality at an additional station on Brushy Creek (BCSW-1 at Post Plant Road). Brushy Creek is a tributary to Horse Creek and flows into Horse Creek between the HCSW-1 and HCSW-2 sampling stations.

This additional station was added for comparison purposes, and will not be evaluated against the HCSP trigger levels and exceedances and Mosaic does not have a NPDES discharge on Brushy Creek. The Brushy Creek location is also not included in the macroinvertebrate or fish sampling components of the program. In September 2009, Mosaic also discontinued water quality analysis for FL-PRO, total fatty acids, and total amines based on recommendations made in previous HCSP annual reports. Mosaic, PRMRWSA, and the TAG agree that the results for these parameters from 2003-2009 show that these substances are present only occasionally at low concentrations, and are not a cause for concern at this time.

Table 2. 2012 Schedule of Water Quality and Biological Sampling Events of the HCSP

Date	Water Quality Sampling Events	Biology Sampling Events
12 January 2012	X (BCSW-1 dry)	
2 February 2012	X (BCSW-1 dry)	
5 March 2012	X (BCSW-1 dry)	
30 March 2012		X
2 April 2012	X	
2 May 2012	X (BCSW-1, HCSW-1, and HCSW-3 all dry)	
5 June 2012	X (BCSW-1 dry, HCSW-2 no flow)	
5 July 2012	X	
2 August 2012	X	
5 September 2012	X	
10 October 2012	X	
26 October 2012		X
6 November 2012	X	
5 December 2012	X (BCSW-1 no flow)	
12 December 2012		X

4.2 Water Quantity

Approved discharge data were obtained from the USGS (http://waterdata.usgs.gov/fl/nwis/nwis) for HCSW-1 and HCSW-4. Staff gauges were installed and stream cross sections were surveyed by Mosaic at HCSW-2 and HCSW-3; stage data were obtained at those stations during monthly water quality sampling. Daily flow and gauge data are not recorded in Brushy Creek, so there is no summary or analysis of water quantity for the sampling location in this report. Discharge data were obtained for Mosaic's National Pollutant Discharge Elimination System (NPDES)-permitted discharges into Horse Creek (Outfalls 003 and 004) for 2003 - 2012 (Figure 1). Daily rainfall data were obtained from Mosaic's rain gauges in the Horse Creek Basin (Figure 1). New Mosaic rainfall gauges (Pine Level 001 and 002) were installed late July 2011; however, due to the limited data set, totals recorded at these locations will not be used in this analysis. The general relationship between rainfall and streamflow was graphically evaluated. All rainfall gauges with an extended period of record are located in the upper portion of the Horse Creek basin (new Pine Level gauges are located parallel with HCSW-3 and HCSW-4 but only have a single year's worth of data), so longitudinal comparisons along the basin are not possible. A separate report (Durbin and Raymond 2006) addresses long-term rainfall patterns in the area.

4.3 Water Quality

A continuous monitoring unit was installed at HCSW-1 to record pH, specific conductivity, dissolved oxygen, and turbidity. Beginning in April 2003, data were recorded hourly, and daily mean, maximum, and minimum were downloaded at least monthly. This data provides for the characterization of natural background fluctuations and allow for the detection of instantaneous conditions or general water quality changes not observed during the collection of monthly grab samples. Low flow or low water conditions resulted in no continuous data for a week in April and from May through the beginning of June 2012.

Water quality samples were obtained monthly, when flow was present, by Mosaic at each of the four monitoring stations beginning in April 2003. The four locations were sampled the same day, working from downstream to upstream. In September 2009, based on recommendations of the PRMRWSA and the TAG, Mosaic began sampling water quality at an additional station on Brushy Creek (BCSW-1 at State Road 64). Brushy Creek is a tributary to Horse Creek and flows into Horse Creek between the HCSW-1 and HCSW-2 sampling stations. All activities affecting sample collection, sample handling, and field-testing activities were thoroughly documented. Field sample collection logs were completed at each station that include the following information: stream level elevations at the time of sampling (from on-site gauges or from the USGS real-time web site); stream size; a qualitative description of the water color, odor, and clarity; weather conditions; field measurements; sample preservation; and any anomalous or unusual conditions. Individual sample containers were labeled with identification codes, date and time of sampling, sample preservation, and the desired analysis. Sample transmittal chain-of-custody records were filled out during sampling listing locations, times, and required analysis.

Field measurements were taken for pH, dissolved oxygen, specific conductivity, and turbidity using meters that were operated and maintained according to manufacturer's instructions. Instruments were calibrated in the field prior to making measurements using the appropriate standards and acceptance limits (Table 3). All calibration activities were documented and records checked for completeness and accuracy. Field measurements by Cardno in association with the three biological sampling events employed a YSI 6920 multiparameter data sonde with the same measuring methods and acceptance limits listed in Table 3. Cardno also employed a Hach 2100P unit for turbidity measurement.

Table 3. HCSP Water Quality Sampling Field Methods and Acceptance Limits Associated with Monthly Sampling by Mosaic Staff.

Analyte	Meter Used	Method	Minimum Detection Limit	Acceptance Limit
pH	Hach Sension 2	150.1	1 su	+/- 0.2 standards units of the calibration
Temperature	Hach Sension 2	170.1		1 degree Centigrade
Specific Conductivity	Hach CO150	120.1	10 uS/cm	+/- 5% of the calibration standard
Dissolved Oxygen	YS1 Model 52	360.1	0.5 mg/l	+/- 0.2 mg/l of the correct Dissolved Oxygen - Temperature value
Turbidity	Hach 2100P	180.1	0.1 NTU	+/- 8% of the calibration standard

Surface water samples were collected in a manner that represented the physical and chemical characteristics of Horse Creek without contamination or bias in the sampling process. Water samples for chemical analysis were generally collected from mid-stream and from mid-depth to the upper portion of the water column unless flows were at either extreme (flood stage or nearly dry at the upper stations). Samples were usually obtained by wading into the stream (taking care not to disturb or stir up bottom sediments) and collecting samples upstream from the sampler. When flooded conditions precluded wading to collect samples (principally at HCSW-3), samples were taken from the top of the water column in the main flow path from the bridge. Samples were collected directly into unpreserved sample containers, which were used to fill the other sample containers. Pre-preserved sample containers (with either sulfuric or nitric acid) were filled with sample water and their pH levels checked. The sample containers were stored on ice prior to transport to laboratories for analysis. Sample containers were either taken directly to the laboratory or laboratory personnel picked them up in the field, using appropriate chain-of-custody procedures. The monthly surface water samples were analyzed for the parameters listed in Table 4. Table 4 also includes the laboratory analysis methods.

In addition to the continuous recorders and monthly water quality sampling, field measurements of temperature, pH, specific conductivity, turbidity and dissolved oxygen were collected during each biological sampling event (Table 2) using a YSI 6920 data sonde. All sampling was conducted according to the Florida Department of Environmental Protection's (FDEP's) Standard Operating Procedures (SOPs) for field sampling. Laboratory analyses were performed by experienced personnel according to National Environmental Laboratory Accreditation Conference (NELAC) protocols.

Results were tabulated to allow for comparisons among stations and sampling events, through time, and to the "trigger values" established for the HCSP (Table 5). In addition, results were compared with applicable Florida surface water quality standards[3] (which in many cases are the same as the trigger values).

[3] While the Florida Numeric Nutrient Criteria (NNC) development has been ongoing since 2009, the adopted criteria did not go into effect until late October 2014; therefore, incorporation of this standard into the HCSP will occur during the 2014 annual report. The trigger level and the NNC evaluate nutrient concentrations over different parameters and time scales. Monthly samples of orthophosphate are compared to the trigger level and identify acute changes in nutrient concentrations that warrant investigation, while the NNC threshold is for total phosphorus and is based on annual geometric mean concentrations and evaluate longer term trends. Also, the nutrient thresholds are only used in conjunction with biological metrics to determine compliance. A site must first pass the floral components (Rapid Periphyton Survey, Linear Vegetation Survey, and annual geometric mean for chlorophyll-a), then either be within the nutrient thresholds or SCI requirements in order to be in compliance according to 62-302.531(2)(c), F.A.C. Therefore, incorporating the NNC thresholds as standalone trigger levels for the HCSP would be inappropriate and would not accurately reflect the NNC.

Table 4. Parameters Analyzed and Laboratory Methods for HCSP 2003 - 2012 Monthly Water Quality Samples.

Parameter	Method	Hold Time	Preservation	Minimum Detection Limit Range	Container
Color	110.2	48 hours	Unpreserved	2-5 PCU	Clear HDPE
Total Kjeldahl Nitrogen	351.2	28 days	Sulfuric Acid, pH < 2	0.008-0.24 mg/l	Clear HDPE
Nitrate-Nitrite Nitrogen	353.2	28 days	Sulfuric Acid, pH < 2	0.0001-1.0 mg/l	Clear HDPE
Total Ammonia Nitrogen	350.1	28 days	Sulfuric Acid, pH < 2	0.0008-0.05 mg/l	Clear HDPE bottle
Orthophosphate	365.1	48 hours	Unpreserved	0.002-0.75 mg/l	Clear HDPE
Chlorophyll *a*	SM	48 hours	Unpreserved	0.25-2.0 µg/l	Opaque plastic
Specific Conductivity	120.1	28 days	Unpreserved	10 µmhos/cm	Clear HDPE
Total Alkalinity	310.1	14 days	Unpreserved	0.24-3.0 mg/l CaCO3	Clear HDPE bottle
Dissolved Calcium*	200.7	28 days	Unpreserved	0.03-0.80 mg/l	Clear HDPE
Dissolved Iron*	200.7	28 days	Unpreserved	0.003-0.10 mg/l	Clear HDPE
Chloride	300.0	28 days	Unpreserved	0.005-30 mg/l	Clear HDPE
Fluoride	300.0	28 days	Unpreserved	0.003-5.0 mg/l	Clear HDPE
Total Radium (Radium 226+228)	903.0	6 months	Nitric Acid, pH < 2	1 pCi/l	Clear HDPE bottle
Sulfate	300.0	28 days	Unpreserved	0.007-100 mg/l	Clear HDPE
Total Dissolved Solids	160.1	7 days	Unpreserved	5-25 mg/l	Clear HDPE

- All water samples were preserved at 4C while awaiting analysis.
- Orthophosphate samples were filtered in the laboratory rather than the field. While Mosaic is cognizant of the FDEP SOP for field sampling, the decision was made to have samples lab filtered (less risk of contamination and the guarantee of lab filtering within hours of lab delivery). Starting in January 2005, samples were field-filtered.
- The analytical method for iron and calcium was changed during the 2003 – 2005 monitoring period.
- Total radium is the arithmetic sum of Radium 226 and Radium 228. Total nitrogen is reported as the arithmetic sum of nitrate+nitrite nitrogen and total Kjeldahl nitrogen. As requested by the PRMRWMA, if either of each pair is undetected, the MDL of the undetected constituent will be used as part of the total. This use of MDL for undetected constituents is contrary to both laboratory and DEP SOPs.
- Petroleum Range Organics, Fatty Amido-amines, and Total Fatty Acid analysis were discontinued in September 2009.

Table 5. Parameters, General Monitoring Protocols, and Corrective Action Trigger Values for the HCSP.

Pollutant Category	Analytical Parameters	Analytical Method	Reporting Units	Monitoring Frequency	Trigger Level	Basis for Initiating Corrective Action Process
General Physio-chemical Indicators	pH	Calibrated Meter	Std. Units	Monthly	<6.0->8.5	Excursions beyond range or statistically significant trend line predicting excursions from trigger level minimum or maximum.
	Dissolved Oxygen	Calibrated Meter	mg/L(1)	Monthly	<5.0	Excursions below trigger level or statistically significant trend line predicting concentrations below trigger level.
	Turbidity	Calibrated Meter	NTU(2)	Monthly	>29	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Color	EPA 110-2	PCU	Monthly	<25	Excursions below trigger level or statistically significant trend line predicting concentrations below trigger level.
Nutrients	Total Nitrogen	EPA 351 + 353	mg/l	Monthly	>3.0	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Ammonia	EPA 350.1	mg/l	Monthly	>0.3	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Ortho Phosphate	EPA 365	mg/l	Monthly	>2.5	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Chlorophyll *a*	EPA 445	mg/l	Monthly	>15	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
Dissolved Minerals	Specific Conductance	Calibrated Meter	µs/cm(3)	Monthly	>1,275	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Alkalinity	EPA 310.1	mg/l	Monthly	>100	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Calcium	EPA 200.7	mg/l	Monthly	>100	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Iron	EPA 200.7	mg/l	Monthly	>0.3 (6); >1.0(7)	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Chloride	EPA 325	mg/l	Monthly	>250	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Fluoride	EPA 300	mg/l	Monthly	>1.5(6); >4(7)	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Radium 226+228	EPA 903	pCi/l(4)	Quarterly	>5	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Sulfate	EPA 375	mg/l	Monthly	>250	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Dissolved Solids	EPA 160	mg/l	Monthly	>500	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
Mining Reagents(5)	Petroleum Range Organics	EPA 8015 (FL-PRO)	mg/l	Monthly	>5.0	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Fatty Acids, Incl.Oleic, Linoleic, and Linolenic Acid	EPA/600/4-91/002	mg/l	Monthly	>NOEL	Statistically significant trend predicting concentrations in excess of the No Observed Effects Level (NOEL to be determined through standard toxicity testing with Mosaic reagents early in monitoring program, NOEL to be expressed as a concentration – e.g., mg/L)
	Fatty Amido-Amines	EPA/600/4-91-002	mg/l	Monthly	>NOEL	Statistically significant upward trend predicting concentrations in excess of No Observed Effects Level (NOEL to be determined through standard toxicity testing with Mosaic reagents early in monitoring program, NOEL expressed as a concentration – e.g., mg/L)
Biological Indices: Macroinvertebrates	Total Taxa Ephemeropteran Taxa Tricopteran Taxa Percent Collector-Filterer Taxa Long-lived Taxa Clinger Taxa Percent Dominant Taxon Percent Tanytarsini Sensitive Taxa Percent Very Tolerant Taxa Shannon-Wiener Diversity(a)	Stream Condition Index (SCI) sampling protocol, taxonomic analysis, calculation of indices according to SOP-002/01 LT 7200 SCI Determination	Units vary based upon metric or index	3 times per year	N/A	Statistically significant declining trend with respect to SCI values, as well as presence, abundance or distribution of native species
Biological Indices: Fish	Total Number of Taxa Abundance Shannon-Wiener Diversity(a) Species Turnover (Morisita Similarity Index(a) Species Accumulation Curves(b)	Various appropriate standard sampling methods, taxonomic analysis, calculation of indices using published formulas	Units vary based upon metric or index	3 times per year	N/A	Statistically significant declining trend with respect to presence, abundance or distribution of native species

Notes:

(1) Milligrams per liter.

(2) Nephelometric turbidity units

(3) Microsiemens per centimeter.

(4) PicoCuries per liter.

(5) If reagents are not detected after two years, sampling frequency will be reduced to quarterly - if subsequent data indicate the presence of reagents, monthly sampling will be resumed. Parameter sampling removed from program in September 2009 as agreed by TAG.

(6) At Station HCSW-4 only, recognizing that existing levels during low-flow conditions exceed the trigger level.

(7)At Stations HCSW-1, HCSW-2, and HCSW-3.

(8) Some metrics have been revised from original HCSP plan document due to revision of DEP SCI Protocol.

References:

(a) Brower, J. E., Zar, J. H., von Ende, C. N. Field and Laboratory Methods for General Ecology. 3rd Edition. Wm. C. Brown Co., .Dubuque, IA. pp. 237; 1990

(b) Gotelli, N.J., and G.R. Graves. 1996. Null Models in Ecology. Smithsonian Institution Press, Washington, DC.

4.4 Benthic Macroinvertebrates

Benthic macroinvertebrate sampling was conducted at HCSW-4 on 30 March 2012, at all stations but HCSW-2 on 26 October 2012, and all stations but HCSW-2 on 12 December 2012. Only HCSW-4 was sampled in March 2012 because the other stations had little to no water or flow. The summer sampling event was delayed until the end of October to ensure that there were at least 90 days of flow because HCSW-1 and HCSW-3 were observed as “dry” in May 2012. No samples were collected at HCSW-2 in 2012 either due to dry/no flow (March and December) or flooded conditions (October). The Brushy Creek location is not included in macroinvertebrate sampling component of the HCSP.

At each Horse Creek station, a Stream Habitat Assessment (DEP-SOP-001/01, Form FD 9000-5) was performed, and a Physical/Chemical Characterization Field Sheet (DEP Form FD 9000-3) was completed. The habitat assessment is comprised of a variety of physical criteria that are independently evaluated on a numerical scale, and the component values are summed to provide a quantitative rating for a stream segment that is presumed to be proportional to the quality of the stream for native macroinvertebrates. The Physical/Chemical form records a variety of other information and also provides for the delineation of various microhabitats in the stream into categories to allow for sampling of such microhabitats in general proportion to their abundance.

Macroinvertebrate sampling was performed in Horse Creek according to the Stream Condition Index (SCI) protocol developed by the DEP (DEP-SOP-002/01 LT 7200) by personnel with training and experience in the SCI protocol and who have successfully passed DEP audits for the protocol. The SCI is a standardized macroinvertebrate sampling methodology that accounts for the various microhabitats available (e.g. leaf packs, snags, aquatic vegetation, roots/undercut banks) within a 100-meter segment of stream. Utilizing this methodology, 20 half meter D-frame dip net sweeps are performed within a 100-meter segment of the stream. The number and quality of benthic macroinvertebrate microhabitats present during the sampling event determines the number of sweeps performed within each microhabitat type. Consistent with DEP protocols, each benthic macroinvertebrate sample was processed and taxonomically analyzed.

Data from each invertebrate sample were used to calculate the various SCI metrics and resulting overall SCI values as per the methodology for the Florida Peninsula (Table 6). The general interpretation for SCI score ranges are provided in Table 7. The calculation methodology for the SCI was revised by DEP in June 2004, and sampling conducted in 2003–2006 uses that methodology. This change requires a departure from the specific metrics listed for benthic macroinvertebrates in the HCSP plan; however, the plan contemplated such changes in methodology and the use of the revised protocol is acceptable with the plan. Between the 2004 and 2005 biological sampling events, individuals conducting biological sampling were trained and audited by the DEP in SCI and Stream Habitat Assessment techniques. Because of this improvement, some SCI results from previous years may not be directly comparable with results from 2005 and beyond.

Table 6. Equations for Calculating SCI Metrics for Peninsular Florida (Range from Zero to Ten).

SCI Metric	Peninsula Score (*)
Total Taxa	10(X-16)/25
Ephemeropteran Taxa	10X/5
Trichopteran Taxa	10X/7
Percent Collector-Filterer Taxa	10(X-1)/39
Long-lived Taxa	10X/4
Clinger Taxa	10X/8
Percent Dominant Taxa	10-(10[(X-10)/44])
Percent Tanytarsini	10[ln(X+1)/3.3]
Sensitive Taxa	10X/9
Percent Very Tolerant Taxa	10-(10[ln(X+1)/4.1])

* In each equation, "X" equals the number representing the count or percentage listed in the corresponding row of the left column. For calculated values greater than ten, the score is set to ten; for values calculated less than zero, the score is set to zero.

Fortunately, the revisions to the SCI protocol in 2004 were implemented before the previous methodology was used to calculate SCI values for the HCSP, so there was no need to retroactively adjust SCI values from 2003 sampling results. Changes made to the calculation protocol were fairly esoteric, essentially based upon a broad array of statistical analyses with invertebrate samples collected across Florida to determine the best correlates with human disturbance to stream habitats (Fore 2004).

In 2007, the FDEP SCI protocol was again revised, this time to include provisions for the taxonomic analysis of two aliquots for every SCI sample collected to account for variation in sample sorting (DEP-SOP-002/01 LT 7200). Under the new protocol, the SCI score for each sample is the arithmetic average of the SCI scores of the two aliquots. Table 6 provides the new list of metrics used in calculating SCI scores, while the parameter table from the HCSP methodology document (copied as Table 5 above) includes the metrics used in the original SCI protocol. In addition to the change in aliquots in 2007, the new protocol also gives a slightly different ecological interpretation of SCI scores (Table 7). Scores from the 2004 SCI (2003 – 2006) and the 2007 SCI (2007 – 2012) may not be directly comparable, given the differences in how they were collected.

The Shannon-Wiener Diversity Index was calculated using Ecological Methodology Software, Version 7.0 (www.exetersoftware.com).

Table 7. Ecological Interpretation of SCI Scores Calculated for Benthic Macroinvertebrate Samples Collected for the HCSP

SCI Category	Range	Typical Description for Range
Category 3 (Exceptional)	68-100	Higher diversity of taxa than for Category 2, particularly for Ephemeroptera and Trichoptera; several more clinger and sensitive taxa than found in Category 2; high proportion for Tanytarsini; few individuals in the dominant taxon; very tolerant individuals make up a very small percentage of the assemblage.
Category 2 (Healthy)	35-67	Diverse assemblage with 30 different species found on average; several different taxa each of Ephemeroptera, Trichoptera, and long-lived and, on average, 5 unique clinger and 6 sensitive taxa routinely found; small increase in dominance by a single taxon relative to Category 1; very tolerant taxa represent a small percentage of individuals, but noticeably increased from Category 1.
Category 1 (Impaired)	0-34	Notable loss of taxonomic diversity; Ephemeroptera, Trichoptera, long-lived, clinger, and sensitive taxa uncommon or rare; half the number of filterers than expected; assemblage dominated by a tolerant taxon, very tolerant individuals represent a large portion or the individuals collected.

4.5 Fish

Fish sampling was conducted at HCSW-4 on 30 March 2012, at all stations but HCSW-2 on 26 October 2012, and all stations but HCSW-2 on 12 December 2012. No fish sampling occurred at HCSW-2 in 2012 either due to dry/no flow (March and December) or flooded conditions (October). Fish sampling did not occur at HCSW-1 or HCSW-3 in March since no SCI was conducted due to not meeting SOP criteria (dry/no flow). The Brushy Creek location is not included in the fish sampling component of the HCSP.

Fish were collected with a 4-foot x 8-foot seine (3 mm mesh size) and by electrofishing with a Smith-Root, Inc. backpack unit (Model LR-24 Electrofisher). Electrofishing was timed (typically 500 seconds), and the number of seine hauls (typically five) was recorded to standardize the sampling efforts among stations and between events.

Some fish (generally those larger than about 10 cm) were identified, weighed, measured, and released in the field, while some large and most small fish (<10 cm) were preserved in the field for analysis in the laboratory. All fish collected were identified in the field or laboratory according to *American Fisheries Society*-accepted taxonomic nomenclature (American Fisheries Society 2004). Total length (mm) and weight (g) were recorded for each individual, with the following exceptions: for samples with very large numbers of fish of the same species (a common occurrence with species like eastern mosquitofish [*Gambusia holbrooki*], least killifish [*Heterandria Formosa*], and sailfin molly [*Poecilia latipinna*]), a randomly selected subset of individuals (approximately 8 to 10) were measured for length and weight, while the remaining individuals were counted and then weighed *en masse*. All fish retained as voucher specimens were submitted to the Ichthyology Collection at the Florida Museum of Natural History in Gainesville.

Taxa richness (number of species) and abundance were determined by station and for each event, and data were compared among stations and across sampling events. The Shannon-Wiener Diversity Index and Morisita's Community Similarity Index were calculated using the Ecological Methodology Software. Species accumulation curves were plotted to estimate the efficacy of the sampling at producing a complete list of the species present in the sampled portions of the stream.

4.6 Initial General Habitat Configuration at Monitoring Stations

The following descriptions and panoramic photos of the four HCSP sampling sites represent the general habitat conditions at the time of initial sampling, April 2003. Several hurricanes in summer 2004, however, substantially altered the landscape and channel of Horse Creek, which have since continued to change through 2012.

The sampling segment at HCSW-1 is a deeply incised, narrow valley with very steep banks of rock-like outcroppings (Figure 4). The substrate is also rocky with little sand accumulation except in deeper holes. There is little woody/herbaceous structure at the water level. There are few undercut banks, but some eroded holes are available for fish and macroinvertebrates in the rocky substrate. Canopy cover in the sampling zone is heavy (>75 percent); thus the area receives a minimal amount of direct sunlight.

At HCSW-2, the sampling segment is essentially an oxbow of the main Horse Creek channel (Figure 4). The substrate is generally sandy. There are numerous holes, snags, and undercut banks and roots present. Canopy cover along the sampling zone is moderate (approximately 25 to 50 percent).

The sampling segment at HCSW-3 is more sinuous than the other three stations, with some shallow, sandy areas and several deep holes (Figure 4). There are numerous snags, undercut banks/roots, and occasional organic debris. Sand is the primary substrate component. During periods of low flow, portions of the sandy bottom are exposed, creating large sand bars. The canopy cover is low (approximately 25 percent); so, the area receives considerable direct sunlight.

At HCSW-4, the sampling segment is less sinuous (Figure 4). Submerged habitats include holes, undercut banks/roots, snags, and small amounts of emergent aquatic vegetation. The substrate is primarily sand, with occasional areas of small gravel. Several sand bars are located in the sampling zone and are exposed during periods of low flow. Canopy cover is moderate (about 50 percent).

4.7 Current Habitat Configuration at Monitoring Stations

At HCSW-1, the channel configuration in the sampling area is essentially fixed by the deeply incised, rock-like banks. Sand and silt deposition was relatively low during the October and December 2012 sampling events. The substrate diversity and availability were consistent in 2012, but water levels and velocities varied, with October having the highest water levels and velocity (Figures 5-7).

At HCSW-2, the size and position of a sand bar on the west side of the stream in the sampling area has changed noticeably, indicating accrual of sediment there (Figures 5-7). In March 2012, the sampling location was completely dry with the exception of a small pool between the 30-40 meter mark (no SCI sampling); the rest of the creek bed consisted of either sand or terrestrial vegetation. During the October event, many habitats were not accessible due to high water levels (no SCI sampling), and the Carolina willow downstream of the 0 meter mark appeared to have been uprooted and washed away. In December, the velocity was less than 0.01 m/sec with the water apparently moving backwards in sections (no SCI sampling). During the 2012 wet season, the water levels overflowed both banks.

At HCSW-3, very few productive habitats were present during the October and December sampling events, and the large area of water hyacinth previously sampled was not observed in 2012 (Figures 5-7). In March, there was a noticeable algae bloom throughout the 100-meter sampling area, with little to no flow observed (no SCI sampling). Sand and silt smothering were slight to moderate for most of the year, while water velocities ranged from 0.01 m/sec (March) to 0.33 m/sec (October). During the wet season, water appeared to overflow the banks, and some bank erosion was present.

At HCSW-4, the stream channel is steep-sided and generally deeper throughout the middle of the sampling area, which continues to complicate sampling efforts. The sand and silt smothering at this station was slight to severe in 2012 due to areas of very shifty sand in the center of the creek. There was more sand and silt deposition in October with higher water levels and velocities observed than in previous years. The water hyacinth returned in high quantities (nearly 50% of the creek) in March, but was not present for the rest of 2012 (Figures 5-7). The stream remained fairly tannic with higher velocities (0.25 to 0.30 m/sec) in October and December. While some habitats were not reachable during October due to higher water levels, snags continue to be very productive habitats and can be sampled throughout the year.

HCSW-1 Horse Creek above SR 64

HCSW-2 Horse Creek above CR 663

HCSW-3 Horse Creek above SR 70

HCSW-4 Horse Creek above SR 72

Figure 4. Panoramic Photographs of the HCSP Sampling Locations, Photos taken on 25 April 2003

Figure 5. Photographs of HCSP Sampling Locations on 30 March 2012.

Figure 6. Photographs of HCSP Sampling Locations on 26 October 2012.

Figure 7. Photographs of HCSP Sampling Locations on 12 December 2012.

5 Results and Discussion

Below we present a summary of water quantity and quality data collected as part of the HCSP in 2012 in Sections 5.1 and 5.2. Results of the 2012 benthic macroinvertebrate and fish sampling are presented in Sections 5.3 and 5.4.

5.1 Water Quantity

5.1.1 Rainfall

Figure 8 includes 2012 total monthly rainfall data from the three Mosaic rain gauges located in the Horse Creek watershed[4] (see Figure 1 for locations). Total and median monthly rainfall in 2012 was slightly different at each gauge, but the heaviest rainfall was observed during May to September at all locations, with some rain events in October (Figure 8). Overall rainfall for 2012 was less than that for 2003 – 2005, greater than totals in 2006 – 2008 and 2010-2011, and similar to totals observed in 2009 (Table 8, Figure 9); it was below the historic range (52.77 in) for the closest long-term station[5]. When one of the rainfall gauges was non-functional, average daily rainfall was calculated from the other functional gauges[6], and total monthly or annual rainfall was calculated from these adjusted daily averages.

Table 8. Annual Total Rainfall in Inches at Mosaic Gauges in the Horse Creek Watershed from 2003 to 2012.

Gauge	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012
Horse Creek North	53.40	53.82	54.52*	31.82*	33.90	40.49	36.63	32.53	24.54*	19.99*
Horse Creek South	59.75	60.74	64.53	34.17	31.97	36.80	43.70	37.47	31.73*	36.06*
Manson Jenkins	30.10*	62.15	31.34*	41.26	32.49	37.48	46.87	41.84	39.85	37.96*
Average of Gauges	57.10	58.90	66.04	37.35	32.79	38.26	42.40	37.28	37.11	44.49[6]

* - Gauge was non-functional during portion of year.

[4] Continuous rainfall data collected by the SWFWMD at HCSW-3 (SWFWMD Station 494) ended in November 2011. Previous HCSP Annual reports used the USGS gauge at HCSW-1, which has been discontinued. At the end of July 2011, two new rainfall gauges (Pine Level 001 and 002) were installed by Mosaic in the lower basin west of stations HCSW-3 and HCSW-4, but they will not be used for general analysis purposes because there is only one complete year of data; these gauges may be used if the three upper basin gauges are all offline at the same time.

[5] Historical rainfall information came from the following stations and years: 1908-1943 NOAA station 148, 1944-2012 average of NOAA station 148 and 336

[6] All three Mosaic upper basin gauges were out of service from August 7, 2012, to September 11, 2012. Average of 2 Pine Level rain gauges were used for this period.

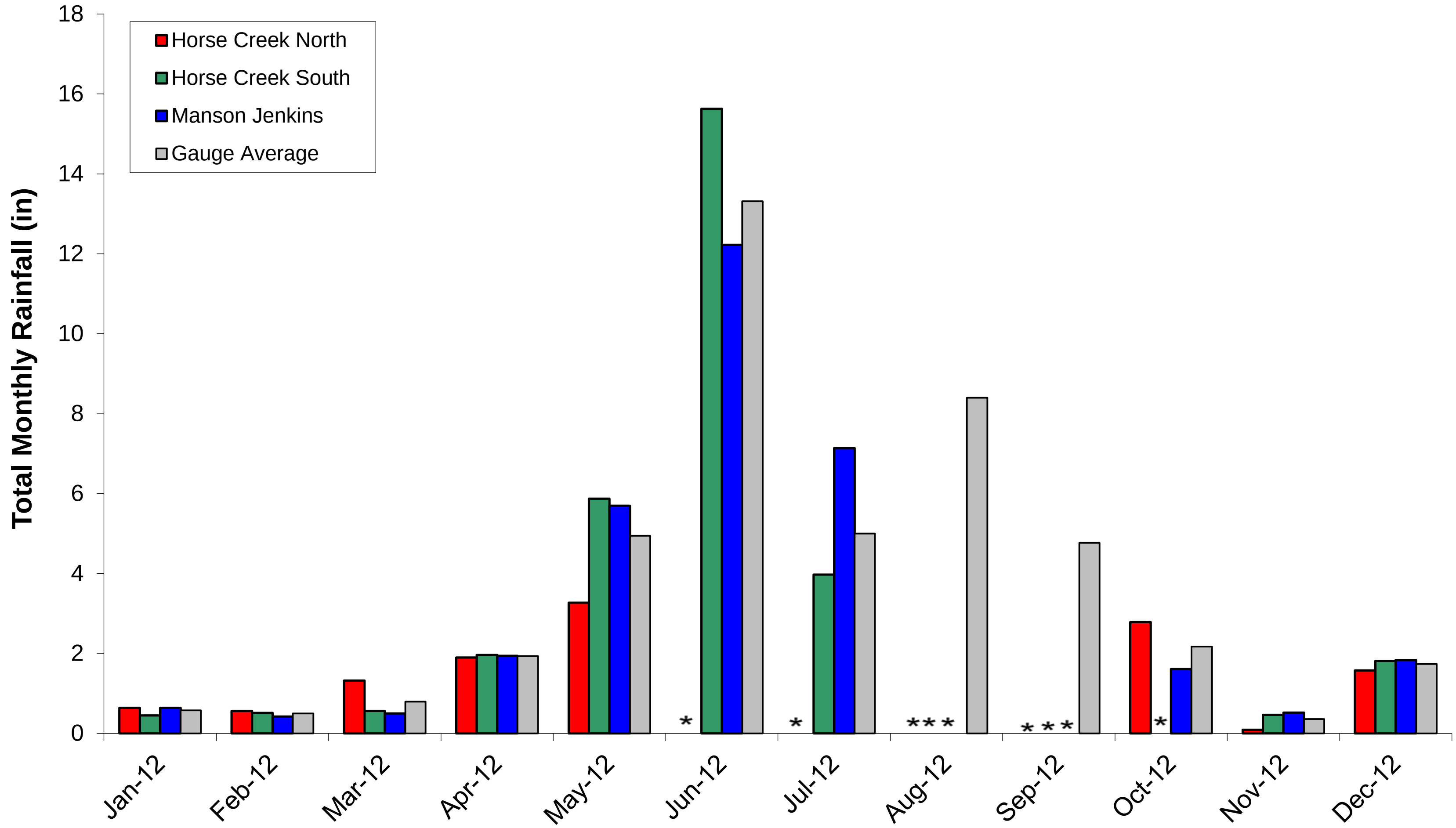

Figure 8. Total Monthly Rainfall From Three Mosaic Gauges in the Horse Creek Watershed in 2012.[7]

[7] * = Gauge non-functional for all or part of month. All three Mosaic upper basin gauges were out of service from August 7, 2012, to September 11, 2012. Average of two Pine Level rain gauges were used for this period.

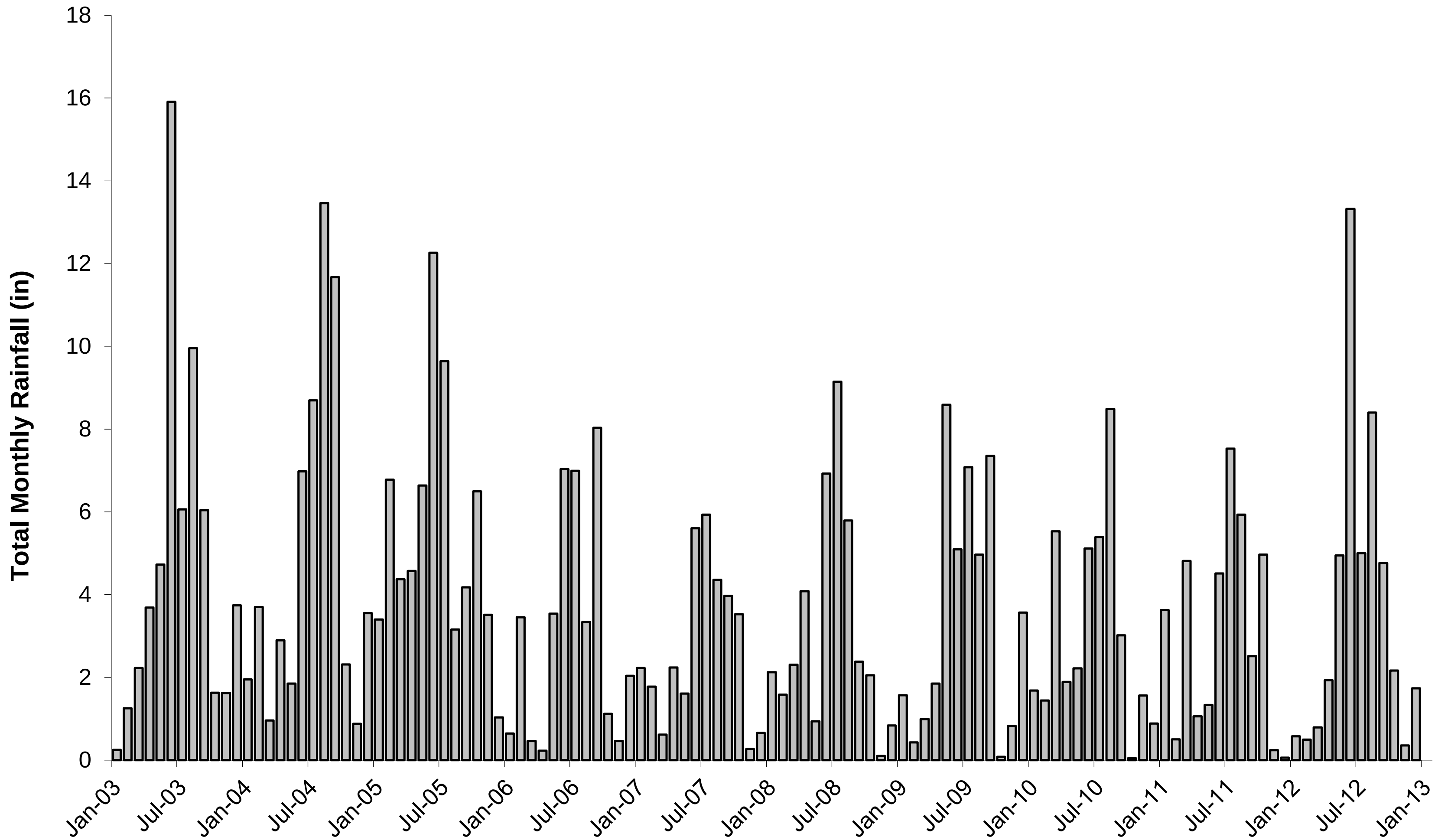

Figure 9. Total Monthly Rainfall from the Average of Three Mosaic Gauges in the Horse Creek Watershed in 2003 - 2012.[8]

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[8] All three Mosaic upper basin gauges were out of service from August 7, 2012, to September 11, 2012. Average of two Pine Level rain gauges were used for this period.

5.1.2 **Stream Stage**

Figure 10 illustrates the relationship between the staff gauge readings made during each Mosaic monthly water-quality sampling event. It also provides the average daily stage as recorded at the USGS gauging stations at HCSW-1 and HCSW-4 (after adjustment to NGVD datum). Patterns of daily stage levels were clearly temporally correlated among the four stations (Figure 10). Stage height (feet NGVD) collected monthly by Mosaic at four sites and continuously by the USGS at two sites was examined using Spearman's rank correlations (Zar 1999) because the gauge heights are not distributed normally (Shapiro-Wilk test for normality, $p < 0.05$). Gauge heights showed a strong and significant correlation between all Mosaic stations and USGS stations (Table 9). Such close correspondence is expected for a fairly small watershed in a low gradient setting like peninsular Florida.

Mean daily stage levels in 2012 were fairly low during the dry season at HCSW-1 and HCSW-4 through May before increasing in June and maintaining higher elevations through the beginning of October (Figure 10). Stage duration curves for 2012 were developed for HCSW-1 and HCSW-4 (Figure 11) to indicate the percentage of time stream stage was above particular elevations. Stage at HCSW-1 varied by just over three feet between the curve's P10 (69.90 feet NGVD) and P90 (66.67 feet NGVD) in 2012, indicating that stream height was relatively constant over time (P10 and P90 are commonly used to bracket the 'typical' fluctuation of a water body, thus omitting the highest and lowest 10 percent of the flows). The small difference in height between the maximum and the P10 show that 2012 rainfall was not enough to raise the stream significantly at HCSW-1. Stream stage at HCSW-4 is more variable than at HCSW-1 between the P10 (19.05 feet NGVD) and P90 (12.24 feet NGVD) (over 6.8 feet), but it also showed a rise in stage beyond the P10 level (~3.4 feet). Stage levels in 2012 were slightly higher than the low levels recorded in 2006 through 2009, but were lower than those recorded in 2003, 2004, and 2005.

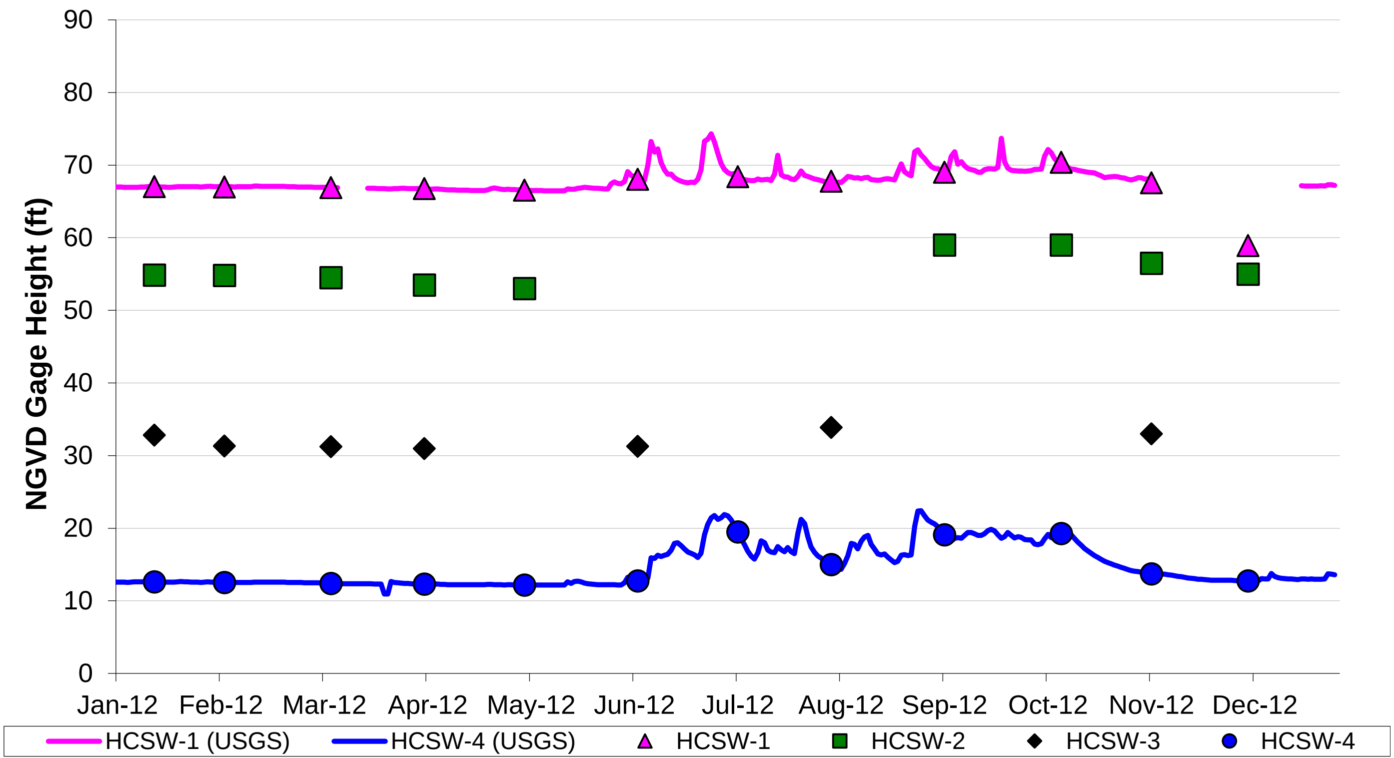

Figure 10. Stream Stage at HCSP Monitoring Stations in 2012. Individual data points are from Mosaic's monthly monitoring; continuous lines are average daily stage from USGS (Stations 02297155 and 02297310).

Table 9. Coefficients of Rank Correlation (rs) for Spearman's Rank Correlations of Monthly Gauge Height (NGVD) for 2003-2012 ($p < 0.0001$).

	HCSW-1 (USGS)	HCSW-4 (USGS)	HCSW-1 (Mosaic)	HCSW-2 (Mosaic)	HCSW-3 (Mosaic)	HCSW-4 (Mosaic)
HCSW-1 (USGS)		0.90	0.99	0.79	0.83	0.89
HCSW-4 (USGS)			0.89	0.85	0.90	0.99
HCSW-1 (Mosaic)				0.77	0.81	0.88
HCSW-2 (Mosaic)					0.87	0.85
HCSW-3 (Mosaic)						0.89
HCSW-4 (Mosaic)						

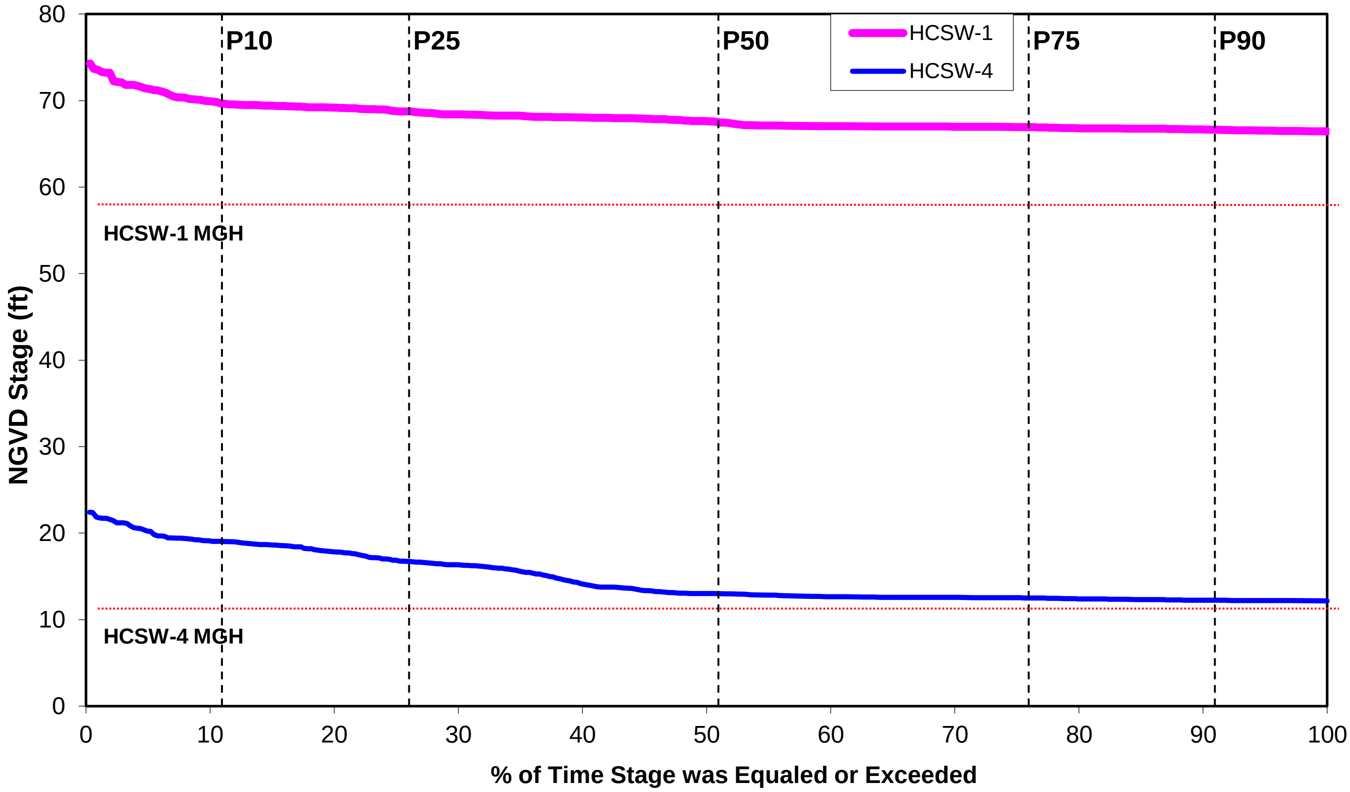

Figure 11. Stage Duration Curves for HCSW-1 and HCSW-4 in 2012, showing percent of year water levels were at or above a given stage. Typical reference points of 10% (P10), 25% (P25), 50% (P50), 75% (P75), and 90% (P90) are indicated on the graph, as well as the minimum gauge heights (MGH) of HCSW-4 (10.96 ft, NGVD) and HCSW-1 (58.12 ft NGVD).

5.1.3 Stream Discharge

The average daily streamflow for 2012, obtained from the USGS continuous recorder data for HCSW-1 and HCSW-4, is presented in Figure 12 and Table 10. In 2012, flows were low from January – May before increasing in June and staying elevated through mid-October before decreasing through the end of the year, similar to historical patterns (Durbin and Raymond 2006). Average daily stream flows exhibited a similar pattern at both HCSW-1 and HCSW-4 (Figure 12); however, streamflow was much higher at HCSW-4 than at HCSW-1 as a logical consequence of HCSW-4's lower position in the basin. At HCSW-1, streamflow in 2012 was similar to previous years, with slightly lower median flows than wet years (2003-2004, 2010) and higher 90th percentile flows than dry years (2006-2008, Table 10). At HCSW-4, streamflow in 2012 was less than 2003 – 2004, but similar to 2006, 2008 – 2011 (Table 10).

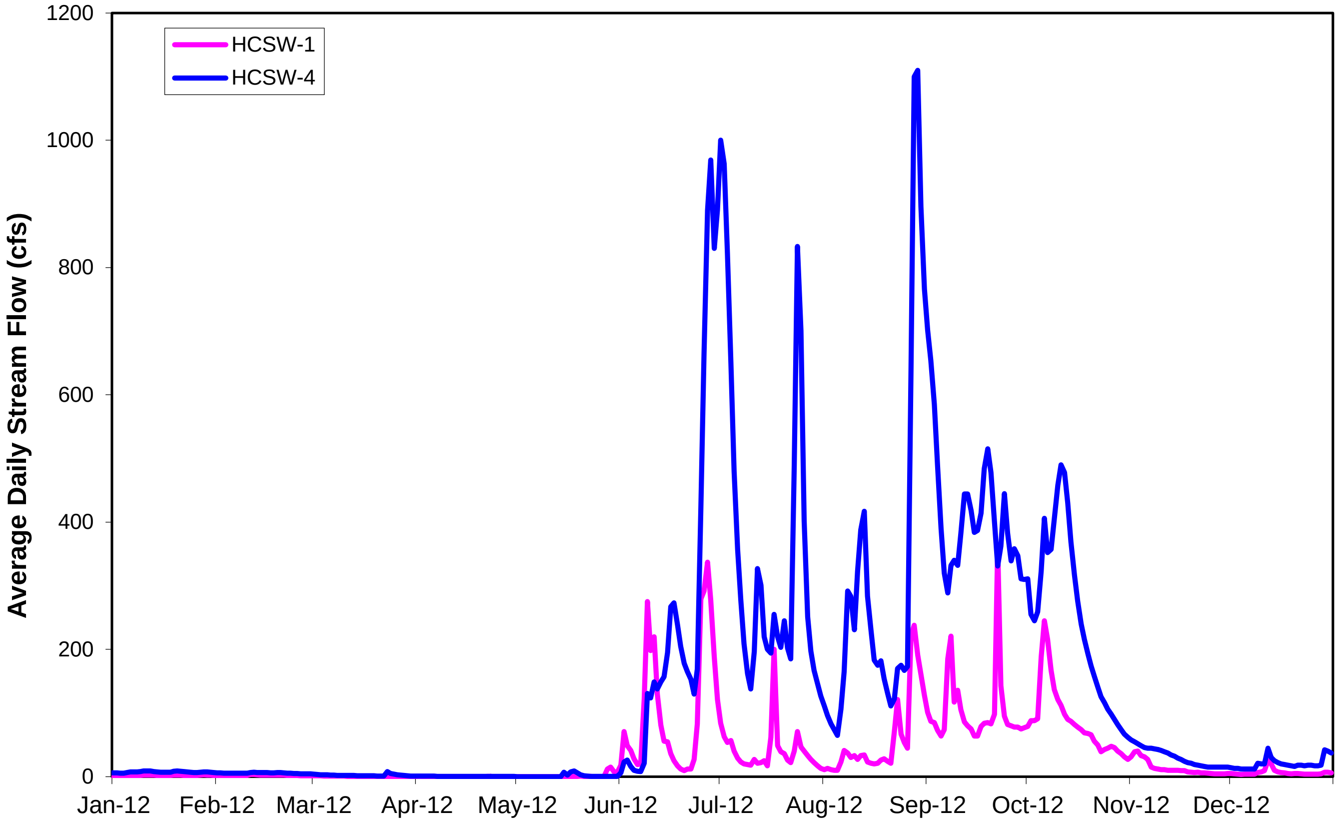

Figure 12. Average Daily Stream Flow for HCSW-1 and HCSW-4 in 2012.

Table 10. Median, 10th Percentile, and 90th Percentile Stream Discharge at HCSW-1 and HCSW-4 in 2003-2012

Station	Year	10th	Median	90th
HCSW-1	2003	2 cfs	20 cfs	127 cfs
	2004	< 1 cfs	7 cfs	166 cfs
	2005	6 cfs	21 cfs	134 cfs
	2006	< 1 cfs	5 cfs	29 cfs
	2007	< 1 cfs	3 cfs	8 cfs
	2008	0 cfs	2 cfs	39 cfs
	2009	< 1 cfs	5 cfs	102 cfs
	2010	< 1 cfs	27 cfs	80 cfs
	2011	< 1 cfs	5 cfs	97 cfs
	2012	< 1 cfs	7 cfs	91 cfs
HCSW-4	2003	21 cfs	84 cfs	1222 cfs
	2004	15 cfs	56 cfs	1184 cfs
	2005	36 cfs	145 cfs	653 cfs
	2006	4 cfs	24 cfs	379 cfs
	2007	4 cfs	14 cfs	43 cfs
	2008	2 cfs	13 cfs	285 cfs
	2009	2 cfs	26 cfs	368 cfs
	2010	19 cfs	93 cfs	379 cfs
	2011	2 cfs	26 cfs	296 cfs
	2012	< 1 cfs	18 cfs	406 cfs

5.1.4 Rainfall-Runoff Relationship

Stream discharge at HCSW-1 and the average daily rainfall for 2012 (average of daily rainfall at three Mosaic rain gauges upstream of Highway 64) are compared in Figure 13. To examine the strength of covariation between daily stream discharge and rainfall, Spearman's rank correlation procedure was used (Zar 1999). Average monthly stream discharge at HCSW-1 was compared to total monthly rainfall at the three Mosaic rain gauges, as well as the average total monthly rainfall of the gauges for the years 2003 - 2012. The correlation between stream discharge at HCSW-1 and rainfall was statistically significant for each rainfall gauge (Table 11). Although these results suggest that stream discharge and rainfall in Horse Creek covary more than would be expected by chance alone, not all of the variation in streamflow is explained by rainfall ($0.49 < r < 0.58$). The lag between rainfall and runoff, as well as other antecedent condition factors, are strongly affecting this relationship; however, there is very little lag between rainfall events and streamflow response after June 2012 (Figure 13). In addition, discharge from the NPDES outfalls may also affect the timing of the rainfall-discharge relationship, although outfall discharge is much more likely to occur in conjunction with periods of increased rainfall. At the beginning of the wet season, the lag effect between rainfall and streamflow is more apparent as the surrounding wetlands or small creeks either need to fill or resume flow. However, later in the wet season when the wetlands and small creeks are full, the lag is much shorter as can be seen in Figure 13. To look at the relationship on a longer timeframe than the HCSP, Figure 14 shows the total monthly rainfall and the monthly average of daily stream discharge at HCSW-1 from 1978 to 2012.

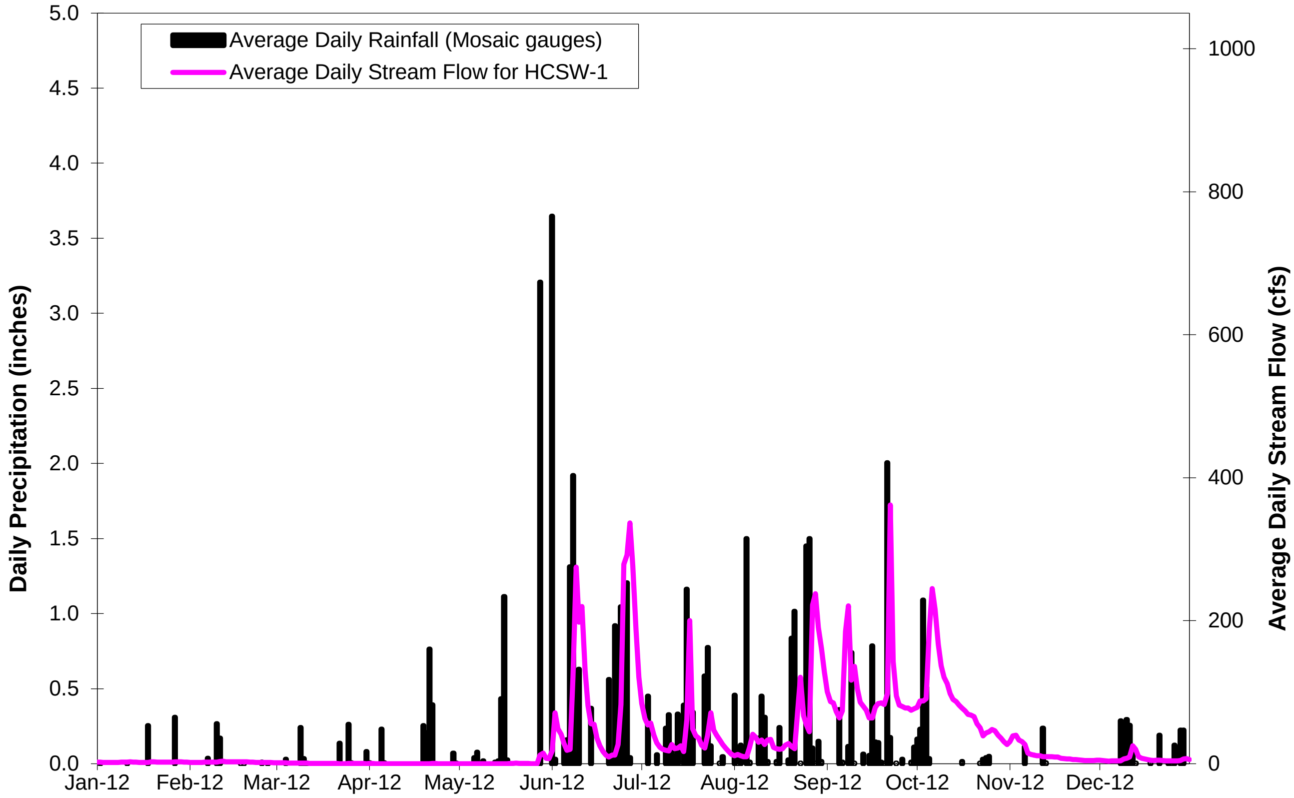

Figure 13. Average Daily Stream Flow at HCSW-1 and Average Daily Rainfall (from 3 Mosaic gauges[9]) in the Horse Creek Watershed in 2012.

Table 11. Coefficients of Rank Correlation (r_s) for Spearman's Rank Correlations of HCSW-1 Monthly Average Stream Discharge and Total Monthly Rainfall at SWFWMD Gauge and Three Mosaic Gauges in 2003 - 2012.

Rainfall Gauge	rs (with HCSW-1 Streamflow)	p value	N (Sample Size)
Horse Creek North	0.50	<0.0001	108
Horse Creek South	0.49	<0.0001	116
Manson Jenkins	0.49	<0.0001	113
Average Rainfall	0.58	<0.0001	117

[9] All three Mosaic upper basin gauges were out of service from August 7, 2012, to September 11, 2012. Average of two Pine Level rain gauges were used for this period.

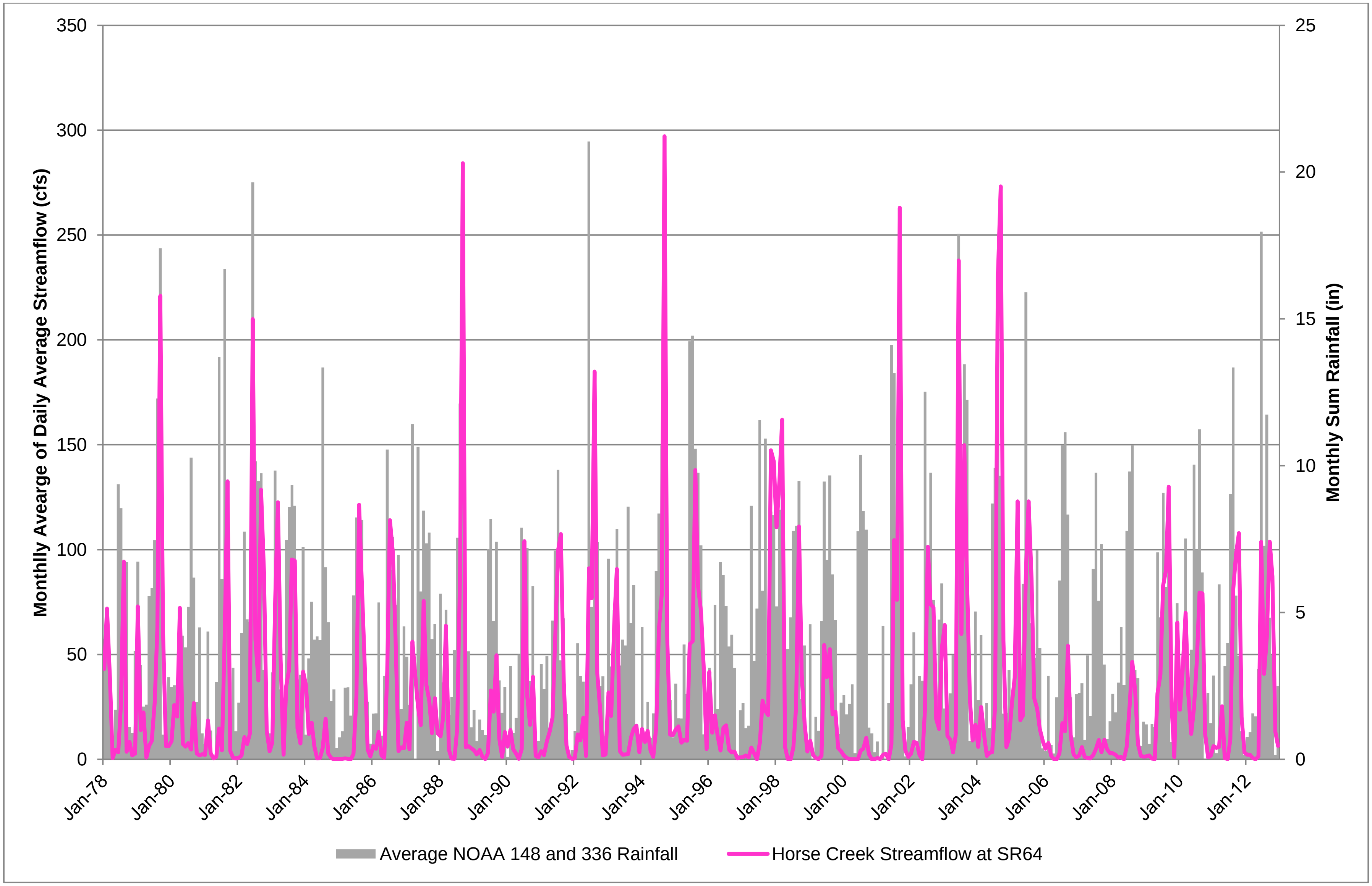

Figure 14. **Monthly Average of Average Daily Streamflow at HCSW-1 and Monthly Sum Rainfall (Average of NOAA 148 and 336 gauges) in the Horse Creek Watershed from 1978-2012.**

To look at the relationship between stream discharge and rainfall over the stream gauge period of record, HCSW-1 discharge was converted from cubic feet per second (cfs) to cumulative discharge in thousands of cfs days. Cumulative historical rainfall was calculated from NOAA gauges 148 and 336, which have a longer period of record than the SWFWMD or Mosaic gauges. Figure 15 illustrates the relationship between cumulative annual discharge at HCSW-1 and annual NOAA rainfall from 1978 to 2012. Changes in the relationship between rainfall and stream discharge can be seen as inflection points in the overall slope. Of the HCSW-1 period of record, we identified three potential inflection points. In 2000, cumulative discharge began to increase slightly relative to rainfall for a few years when rainfall was above average, when compared to the slope of the overall period of record, meaning there was more stream discharge per unit of rainfall. Between 2005 and 2008, which included several very dry years, cumulative discharge had almost no increase, despite changes in cumulative rainfall. Thus, as expected during a very dry period, the relationship changed and less water entered the stream per unit rainfall than happened during wetter periods. After 2008, the slope was again similar to the wet period of 2000-2004 and the overall period of record slope, because rainfall began to return to average conditions and cumulative discharge began to resume previous patterns relative to cumulative rainfall. If mining was having a significant effect on the amount of water that reached Horse Creek compared to rainfall, then one would expect to see one or more large inflection points that correspond to the beginning of mining in the basin or the mining of large tracts and last for many years. However, for the majority of the period of record (which included pre-mining data), the relationship is remarkably constant over time, with only a few minor inflection points that correspond to unusually wet and dry periods in the 2000 decade.

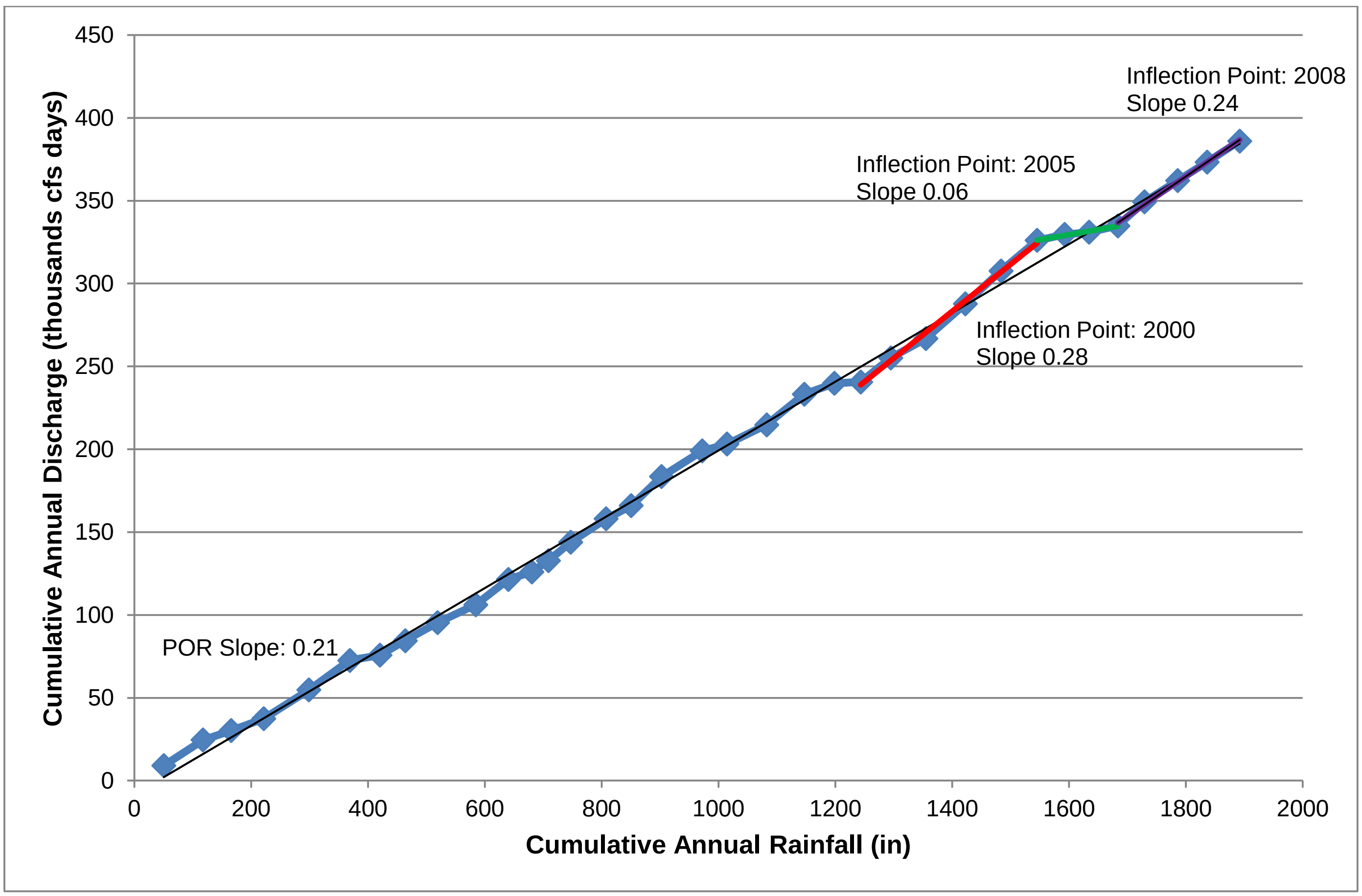

Figure 15. Double Mass Curve of Cumulative Daily Discharge (USGS gauge at SR64) and Rainfall (NOAA gauges 148 and 336) at HCSW-1 in 1978 – 2012.

5.1.5 NPDES Discharges

Industrial wastewater is discharged to Horse Creek through two outfalls located at the Fort Green Mine (Outfalls 003 and 004 on NPDES Permit FL0027600, see Figure 1). Both outfalls are 20-foot wide concrete flumes with continuous flow measurement. A mine wastewater system consists of clay settling areas, mined but not yet reclaimed land, and unmined but disturbed lands. The runoff from all these lands is contained within the industrial wastewater system boundaries. The “loop” of wastewater from the plant to the clay settling areas with the subsequent return of clarified water to the plant for reuse is the backbone of the system. The system has a finite storage capacity and excess wastewater (as a result of rainfall into the system) is discharged from permitted outfalls. This general relationship is illustrated in the rainfall and NPDES discharge data for 2012 (Figure 16). The Horse Creek outfalls, however, are not the only discharge points of the mine, so this data represents only a portion of the mine’s rainfall-discharge relationship (Table 12). The Horse Creek portion of the Fort Green Mine is not a distinct entity on the ground; the mine property is continuous and covers portions of several basins and, as such, conclusions drawn from this data may be misleading. Mosaic has no other discharges to Horse Creek, and no other known industrial wastewater discharges to Horse Creek or any tributary by any other firm are known.

Because they potentially affect stream discharge, the combined 2012 daily discharge of two Mosaic NPDES outfalls (Outfalls 003 and 004) located upstream of HCSW-1 was plotted against the 2012 daily flow for HCSW-1 (Figure 17)[10]. In 2012, the Ft. Green NPDES outfalls discharged during portions of four months (27 August to 5 November 2012) into Horse Creek. Comparing HCSW-1 stream discharge and NPDES discharge in 2003 - 2012 using a Spearman’s rank correlation procedure (Zar 1999) indicates they covary strongly ($r_s = 0.73$, $p < 0.0001$). Thus, an increase in one parameter will correspond to an increase in the other. Just as stream discharge at HCSW-1 was correlated with rainfall (Table 11), so too is NPDES discharge (Table 13, Figure 16), with lagtimes and antecedent conditions affecting this relationship. There is a lag in the start of NPDES discharge relative to rainfall (similar to the lag between rainfall and streamflow), because the NPDES system must fill to the discharge elevation, which can occur further into the wet season. NPDES discharge can also continue after the wet season rains have slowed (Figure 16) until water is once again below the discharge elevation in the circulation system.

Table 12. 2012 Total monthly Mosaic Industrial Wastewater Discharge (NPDES) to Horse Creek (Outfalls 003 and 004) and Payne Creek (Outfall 001, 002, 005, 006) from the Fort Green Mine.

Month	Discharge to Payne Creek (MG)	Discharge to Horse Creek (MG)
January	101.10	0.00
February	47.70	0.00
March	20.70	0.00
April	57.14	0.00
May	153.91	0.00
June	1,346.05	0.00
July	4,567.20	0.00
August	3,889.90	113.61
September	2,403.43	850.20
October	1,710.48	904.60
November	1,328.29	78.29
December	730.68	0.00
Annual Total	**16,666.58**	**1,946.70**

[10] Mosaic gauge may be based on instantaneous rather than continuous flow.

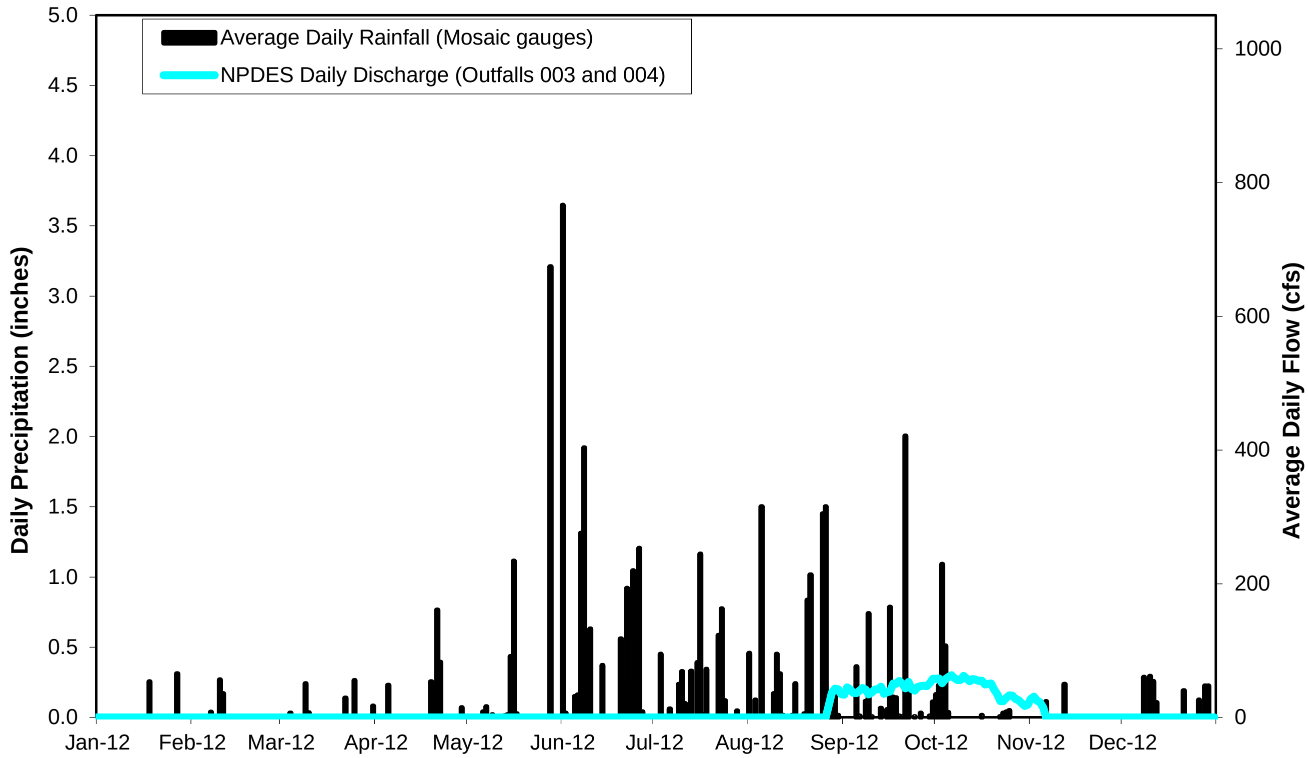

Figure 16. **Combined Mosaic NPDES Discharge and Average Daily Rainfall in the Horse Creek Watershed in 2012.**

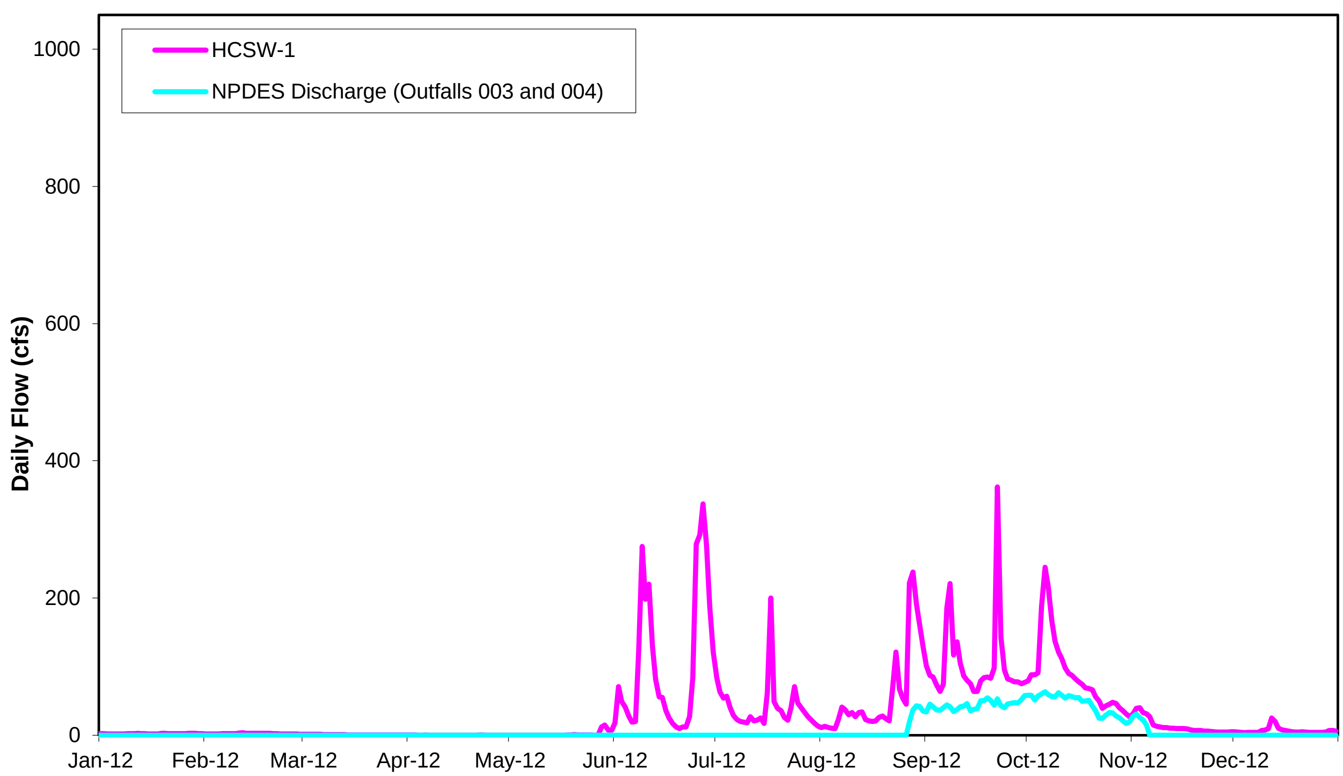

Figure 17. **Daily Flow at HCSW-1 and Combined Mosaic NPDES Discharge for 2012.**

Table 13. Coefficients of Rank Correlation (r_s) for Spearman's Rank Correlations of NPDES Monthly Average Discharge and USGS Daily Discharge, Gage Height, and Monthly Total Rainfall at Three Mosaic Gauges in 2003 – 2012.

Gauge	r_s (with NPDES Outfall)	p value	N (Sample Size)
HCSW-1 (USGS Discharge)	0.73	< 0.0001	117
HCSW-1 (USGS Gauge Ht)	0.73	< 0.0001	116
Horse Creek North (Rain)	0.40	< 0.001	108
Horse Creek South (Rain)	0.32	< 0.001	116
Manson Jenkins (Rain)	0.24	0.01	113
Average Rainfall	0.38	< 0.001	117

5.1.6 Summary of Water Quantity Results

Although low and median Horse Creek discharge in 2012 was average for the region, rainfall in 2012 was below the long-term average annual rainfall of 52.72 inches (1908-2012)[11]. For 2012, temporal patterns of average daily stream flow and stage were similar across all stations, with the majority of high flows and stages occurring during May to August, during the rainy season. The dry season (January to May 2012) was extremely dry, resulting in periods of little to no streamflow and no NPDES discharge. Summer rains began in late-May/early-June, with lag in streamflow response until mid-June. At the beginning of the wet season, the lag effect between rainfall and streamflow is more apparent as the surrounding wetlands or small creeks either need to fill or resume flow. However, later in the wet season when the wetlands and small creeks are full, the lag is much shorter. Higher streamflow continued through the end of October/first of November.

In September and October, NPDES contributed up to 75 percent of the streamflow at HCSW-1 compared to rainfall; in late October 2012, NPDES discharge accounted for almost all of the streamflow at HCSW-1. NPDES discharge from August to November was also a lagged response to rain that occurred from late-May to early October 2012; NPDES discharge from the Horse Creek outfalls usually does not occur until sufficient water storage accumulates in the circulation system, resulting in a lag. In general, the lower than average rainfall resulted in lower than average streamflow in Horse Creek, with some lags. There is no evidence that mining and reclamation activities in the basin caused any significant decrease in total streamflow in 2012. In a previous study (Robbins and Durbin 2011) that compared streamflow during dry years in reference and potentially impacted streams before and during phosphate mining, there was no evidence that phosphate mining practices caused lower monthly flows in the potentially impacted streams (including Horse Creek) than what would be expected given the conditions in a reference stream (Charlie Creek).

[11] Historical rainfall information came from the following stations and years: 1908-1943 NOAA station 148, 1944-2012 average of NOAA station 148 and 336

5.2 Water Quality

The results of field measurements and laboratory analyses of water samples obtained monthly during 2012 at each HCSP monitoring station are presented in this section (see Appendix C for water quality figures from 2003 through present). Continuous recorder data for pH, dissolved oxygen, turbidity, and specific conductivity are also presented, along with the field measurements obtained during benthic macroinvertebrate and fish sampling on 30 March, 26 October, and 12 December 2012. Water quality raw data are included in a database on the attached CD-ROM.

In September 2009, based on recommendations of the PRMRWSA and the TAG, Mosaic began sampling water quality at an additional station on Brushy Creek (BCSW-1 at Post Plant Road). Brushy Creek is a tributary to Horse Creek and flows into Horse Creek between the HCSW-1 and HCSW-2 sampling stations. This additional station was added for comparison purposes, and will not be evaluated against the HCSP trigger levels and exceedances. Mosaic does not have a NPDES discharge on Brushy Creek. While the Brushy Creek data has been included in the graphs of the 2012 water quality data, it was not included in any other plots or analyses.

In September 2009, Mosaic also discontinued water quality analysis for FL-PRO, total fatty acids, and total amines based on recommendations made in previous HCSP annual reports. Mosaic, PRMRWSA, and the TAG agree that the results for these parameters from 2003-2009 show that these substances are present only occasionally at low concentrations, and are not a cause for concern at this time.

Water quality of NPDES discharge is normally obtained periodically when water is discharged from Outfalls 003 and 004. Water was discharged for 71 days in 2012 from Outfall 004, with multiple water quality samples taken. No water was discharged from Outfall 003 in 2012. For all 2012 NPDES discharge water quality results, chlorophyll a was the only water quality parameter above the Horse Creek trigger levels (Table 14). Of two total measurements, one measurement was above the trigger level of 15 mg/m^3.

Table 14. Water quality summary of NPDES discharge into Horse Creek during 2012 at Outfall 004.

	Outfall 004 (August – November)			
Constituent	Avg	Count	Min	Max
pH (su)	7.86	11	7.39	8.44
Conductivity (umhos/cm)	555	4	496	626
Temperature (degrees C)	27.7	4	22.6	32.2
Turbidity (NTU)	8.72	4	5.67	14.1
Dissolved Oxygen (mg/L)	7.34	4	5.68	10.4
TSS (mg/L)	7.71	11	2.00	20.0
Total Phosphorus (mg/L)	0.75	11	0.05	1.49
TKN (mg/L)	1.33	4	0.93	1.90
Nitrate-Nitrite (mg/L)	0.02	4	0.01	0.02
Total Nitrogen (mg/L)	1.32	3	0.94	1.92
Fluoride (mg/L)	0.71	2	0.67	0.67
Sulfate (mg/L)	120	2	104	136
Chlorophyll *a* (mg/m^3)	13.6	2	1.66	**25.6**

5.2.1 Data Analysis

Line graphs are used to display water quality measurements for each parameter during 2012, but the lines connecting each station's measurements are included merely to enhance visual interpretation and not to imply that the values between actual measurements are known (Appendix C contains line graphs for each parameter from 2003 to 2012). For continuous recorder data measured at HCSW-1 in 2012, the daily mean of the water quality parameter is plotted with streamflow from the USGS gauge at HCSW-1. Monthly water quality data for 2003–2012 were compared to other data sources (SWFWMD, FDEP, USGS) since 1990 using median box-and-whisker plots[12]. Graphical representations of HCSP data include undetected values, represented by the respective MDLs for each parameter, except for total nitrogen and total radium. Total nitrogen and total radium are composite parameters without MDLs. Values of these parameters for which one or both components were undetected are circled in red. Undetected results for all parameters were changed to one-half of the MDL for any statistical analyses.

Based on a literature review (Appendix D) on tests for water quality data trend detection, the best monotonic trend detection method for use in the Horse Creek Stewardship Program (HCSP) is the Seasonal Kendall. Because the USGS recommends a minimum of five years of data collection before applying the Seasonal Kendall test (Schertz et al. 1991), the 2008 Annual Report was the first report to include this analysis. The Seasonal Kendall method is a frequently recommended method for detecting trends in water quality data (Lettenmaier 1988, Hirsch et al. 1982, Harcum et al. 1992, Helsel et al. 2005). The Seasonal Kendall was developed by and is now the method of choice for the United States Geological Survey (USGS) (Hirsch et al. 1982, Helsel et al. 2005).

The Seasonal Kendall test determines water quality trends after correcting for seasonality by only comparing values between similar seasons over time (Schertz et al. 1991). The Seasonal Kendall test selects one value for each season (average, median, or subsample) and makes all pair-wise comparisons between time-ordered seasonal values (Harcum et al. 1992). A test statistic (Tau) is computed by comparing the number of times a later value is larger than an earlier value in the data set, and vice-versa (Schertz et al. 1991). Results for the Seasonal Kendall include the magnitude of the statistic Tau, its significance (p), its slope (Sen slope estimator), and the direction of significant trends. The trend (Sen) slope is the median slope of all pairwise comparisons. The direction of this slope (positive or negative) is more resistant to the effects of observations below minimum detection limits and missing data than the magnitude of the slope. The slope is a measure of the monotonic trend; the actual temporal variation may include step trends or trend reversals. If alternate seasons exhibit trends in opposite directions, the Seasonal Kendall test will not detect an overall trend (Lettenmaier 1988).

Trend detection in this report is limited by several factors. With only ten years of data, the power of the test to detect trends of small magnitude will be limited (Harcum et al. 1992, Hirsch et al. 1982). Data for parameters whose method detection limits have changed several times over the HCSP (fluoride, nitrate+nitrite, ammonia) would have to be truncated to the highest detection limit, thereby reducing the available data for the test. As a result, trends were evaluated using alternative data collected by SWFWMD for these parameters. Because HCSW-1 and HCSW-4 were the only stations with USGS flow data that is necessary for interpreting trends in flow-dependent parameters, they were the only stations used in this trend analysis. Any changes over time detected using the Seasonal Kendall test should be further examined to determine if the perceived change is caused by a data bias, if its magnitude is ecologically significant, and if the cause of the trend is related to Mosaic mining activities. If warranted, an impact analysis can be performed on statistically significant trends to determine if trends are caused by Mosaic mining activities and if a corrective action by Mosaic is necessary.

A summary of the Seasonal Kendall Tau results for all parameters is presented in Table 15. The year was split into three seasons, corresponding to wet/dry periods. Season one encompassed the first part of peninsular Florida's dry season, January through April. Season two spanned May to September (the wet season along with

[12] In median box-and-whisker plots, the small center square is the median of the distribution, and the large box is bounded by the 25% (mean – standard error) and 75% (mean + standard error) quartiles of the distribution. The length of the large box is designated H, and the "whiskers" represent the range of values between the box limits and 1.5H above and below the box limits. Outside the whiskers lie outliers and extreme values. Outliers are values that lie between 1.5H and 3H from the box limits, and extreme values lie beyond 3H from the box limits (StatSoft, Inc 2005).

May, which in this region tends to be fairly rainy) and season three represented the second dry season during the calendar year, October through December.

Parameters that were significantly correlated with USGS streamflow were corrected for the effect of annual variation in streamflow using a LOWESS smooth (F=0.5) before the Seasonal Kendall Tau was performed. LOWESS (locally weighted scatterplot smoothing) in the seasonal Kendall Tau describes the relationship between the concentrations of a water quality parameter and streamflow using a weighted linear least squares regression. The residuals of the smooth have the effect of streamflow subtracted, and are called flow-adjusted concentrations (Hirsch et al. 1991). Flow adjusted concentrations are necessary when the variable in question has an inherent relationship with streamflow, which can confound any comparisons made of water quality between stations or times with different instantaneous flow. If the variability of a water quality parameter could be completely explained by streamflow, then during smoothing, all of the data points would fall along a single best-fit line, and all of the residuals (distance between the points and the line) would be zero. For real data, the differences between the data points and the best-fit line show the part of the variability in water quality that is not caused by changes in streamflow, i.e. the flow-adjusted concentrations.

The Sen slope estimate for a parameter was only reported if the trend was statistically significant. For those parameters where SWFWMD data was used for trend analysis (fluoride, nitrate+nitrite, and ammonia), the magnitude of the slope estimate may not be accurate for ammonia because data in 2007 was missing from the SWFWMD dataset. In addition, in October 2011, SWFWMD went from monthly sampling to every other month, making the slope estimates for the third season inconsistent with the analysis that used the HCSP data. For that reason, the analysis used for the 2012 report was the Annual Kendall Tau analysis using annual median values with LOWESS smoothing of annual average streamflow. While there is more statistical power in using the Seasonal Kendall Tau (more data points, monthly flow instead of annual, and division of data by season), it was no longer feasible for these three parameters due to the data restrictions. For those parameters with statistically significant trends, Appendix I contains additional graphics from a more detailed impact analysis of the data than what is discussed under the relevant parameter headings in the report text below.

Table 15. Summary of Seasonal Kendall-tau with LOWESS (F=0.5) for HCSW-1 and HCSW-4 from 2003-2012 Unless Noted.

Parameter	HCSW-1				HCSW-4			
	tau	p-value	slope	2012 Median	tau	p-value	slope	2012 Median
pH	0.42	**0.004**	**0.05**	7.62	0.03	0.79	N/A	7.39
Dissolved Oxygen	-0.07	0.68	N/A	7.69	-0.17	0.256	N/A	7.18
Turbidity	0.08	0.61	N/A	4.32	-0.67	0.68	N/A	2.92
Color, total	0.32	**0.03**	**5.25**	120	0.50	**0.001**	**10.6**	80.0
Nitrogen, total	0.07	0.68	N/A	0.99	0.11	0.47	N/A	1.55
Nitrogen, total Kjeldahl	0.04	0.84	N/A	0.86	0.19	0.22	N/A	0.96
Nitrogen, ammonia*	-0.56	**0.05**	**-0.0003**	0.01	0.17	0.60	N/A	0.05
Nitrogen, nitrate-nitrite*	0.11	0.72	N/A	0.05	0.02	1.00	N/A	0.36
Orthophosphate	0.36	**0.01**	**0.02**	0.42	0.16	0.30	N/A	0.46
Chlorophyll a[1]	-0.13	0.40	N/A	0.72	-0.10	0.50	N/A	1.80
Specific Conductance	0.51	**0.004**	**10.6**	320	0.16	0.30	N/A	189
Calcium, dissolved	0.51	**0.0004**	**1.05**	24.0	0.20	0.18	N/A	78.0
Iron, dissolved	-0.35	**0.02**	**-0.02**	0.15	-0.35	**0.02**	**-0.01**	0.08
Alkalinity	0.42	**0.004**	**2.96**	58.4	0.53	**0.0003**	**1.66**	73.0
Chloride	0.26	0.08	N/A	18.9	0.20	0.18	N/A	31.2
Fluoride*	0.07	0.86	N/A	0.55	-0.11	0.72	N/A	0.52
Sulfate	0.32	**0.03**	**2.27**	62.5	0.13	0.41	N/A	189
Total Dissolved Solids	0.35	**0.02**	**6.64**	216	0.23	0.12	N/A	478
Radium, total[1]	-0.81	0.60	N/A	0.80	-0.09	0.56	N/A	1.60

*SWFWMD data was used from April 2003-December 2012 (some parameters were missing 2007 data). Annual Mann Kendall with LOWESS was used for analysis of 2003-2012 data since sampling was reduced to every other month starting October 2011.

[1]Data was not correlated with streamflow for either station; LOWESS was not used.

Differences in water quality between stations from 2003 to 2012 for each water quality parameter were evaluated using ANOVA and Duncan's post hoc test (Table 16). This analysis will help to identify potential differences among stations that can be examined in more detail as the HCSP continues. A summary of the ANOVA results for all parameters is presented in Table 16. Parameters whose MDLs have changed over the course of the program were omitted because of limited comparable data between sampling events and stations (i.e., fluoride, nitrate+nitrite, ammonia).

Table 16. Summary of results from ANOVA for differences between stations from 2003-2012.

Parameter	F	p-value
pH	35.32	**< 0.001**
Dissolved Oxygen	131.09	**< 0.001**
Turbidity	2.02	0.11
Color, total	6.58	**< 0.001**
Total Nitrogen	8.68	**< 0.001**
Total Kjeldahl Nitrogen	16.94	**< 0.001**
Orthophosphate	23.7	**< 0.001**
Chlorophyll A	28.55	**< 0.001**
Specific Conductance	37.69	**< 0.001**
Calcium, dissolved	61.92	**< 0.001**
Iron, dissolved	0.54	0.66
Alkalinity	35.86	**< 0.001**
Chloride	23.55	**< 0.001**
Sulfate	47.25	**< 0.001**
Total Dissolved Solids	41.08	**< 0.001**
Radium, Total	3.12	**< 0.05**

Water quality parameters were compared with water quantity variables recorded during the same month from 2003 to 2012: average daily streamflow for the month, average daily NPDES discharge for the month, and total monthly rainfall. Because these three water quantity variables are correlated to each other (Table 13), a statistically significant correlation between NPDES discharge and water quality does not prove a causal relationship between water quality and mining discharge. The results of this correlation analysis are presented in Table 17. Each of these correlations is discussed further in each water quality section. Parameters whose MDLs have changed over the course of the program were omitted because of limited comparable data between sampling events and stations (fluoride, nitrate+nitrite, ammonia).

Table 17. Spearman's rank correlation between water quality and water quantity at HCSW-1 and HCSW-4, as represented by average daily streamflow, average daily NPDES discharge, and total rainfall for the same month from 2003 - 2012.

Parameter	HCSW-1			HCSW-4		
	NPDES	Rainfall	Streamflow	NPDES	Rainfall	Streamflow
pH	-0.08	**-0.36***	**-0.36***	**-0.35***	**-0.35***	**-0.62***
Dissolved Oxygen	**-0.41***	**-0.57***	**-0.47***	**-0.53***	**-0.55***	**-0.71***
Turbidity	**0.51***	**0.32***	**0.54***	**0.48***	**0.31***	**0.60***
True Color	**0.44***	**0.41***	**0.49***	**0.58***	**0.29***	**0.71***
Total Nitrogen	**0.33***	**0.51***	**0.48***	0.17	0.18	**0.37***
TKN	**0.36***	**0.53***	**0.50***	**0.42***	**0.36***	**0.59***
Orthophosphate	0.01	-0.16	**-0.21***	0.10	0.08	0.04
Chlorophyll a	0.03	-0.07	0.03	0.09	0.13	0.10
Specific Conductance	**0.20***	**-0.23***	-0.10	**-0.52***	**-0.34***	**-0.79***
Calcium, dissolved	**0.21***	**-0.27***	-0.14	**-0.58***	**-0.32***	**-0.81***
Iron, dissolved	**0.35***	**0.62***	**0.62***	**0.53***	**0.43***	**0.79***
Alkalinity	**0.23***	**-0.20***	-0.02	**-0.40***	**-0.55***	**-0.76***
Chloride	**-0.58***	**-0.38***	**-0.69***	**-0.58***	**-0.33***	**-0.81***
Sulfate	**0.29***	-0.14	-0.03	**-0.54***	**-0.31***	**-0.75***
TDS	**0.34***	-0.09	0.04	**-0.45***	**-0.24***	**-0.71***
Radium, Total	**-0.35***	0.13	-0.16	**-0.38***	0.15	-0.14

* - Statistically significant at $p < 0.05$

5.2.2 Physio-Chemical Parameters

pH

Levels of pH, dissolved oxygen, turbidity, and specific conductivity were obtained in the field during each monthly water-quality sampling event. Values of pH were within the range of established trigger levels during all of 2012 sampling events at all stations with the exception of the October 2012 measurement at HCSW-2 (Figure 18). Monthly sampling in October 2012 occurred a few days after both a rainfall and streamflow pulse which most likely flushed lower pH water from wetlands upstream of the sampling locations. In the case of HCSW-2, it received waters from the Horse Creek Prairie which could have had a lower pH than the other sampling locations causing the exceedance. Values obtained during biological sampling events were fairly consistent with pH levels determined during the monthly water quality sampling events (Figure 18). The pH levels at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources (FDEP, SWFWMD, USGS) (Figures 19 and 20). Continuous pH data obtained daily at HCSW-1 in 2012 was within a range similar to that obtained during monthly water quality sampling (Figure 21).

HCSW-4 exhibited no monotonic trends from 2003 to 2012 for pH (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15). There was a slightly increasing monotonic trend for pH at HCSW-1 (Seasonal Kendall Tau with LOWESS, Tau = 0.42, $p < 0.05$, Sen slope = 0.05 SU per year, Figures 19 and 20). The slope for this potential trend is very small compared to limits in laboratory prediction or the observed differences between primary and duplicate field samples, and does not represent an adverse trend that would need additional analysis or corrective action at this time (Table 2 of Appendix I).

Levels of pH were significantly different among stations in 2003 – 2012 (ANOVA F = 35.32, $p < 0.001$, Table 16). Station HCSW-2, which had significantly lower pH than other stations followed by HCSW-3 (Duncan's multiple range-test, $p < 0.05$), lies just downstream of a large swamp complex that has the potential to add substantial organic acids from plant decomposition and decrease the pH (Reid and Wood 1976). Brushy Creek also contributes to HCSW-2, and similarly has a relatively low pH compared to the Horse Creek stations. Levels of pH were significantly correlated with streamflow (r_s= -0.36) and rainfall (r_s= -0.34) at HCSW-1 and with streamflow (r_s= -0.62), rainfall (r_s= -0.35), and NPDES discharge (r_s= -0.35) at HCSW-4 (Spearman's rank correlation, $p < 0.05$, Table 17).

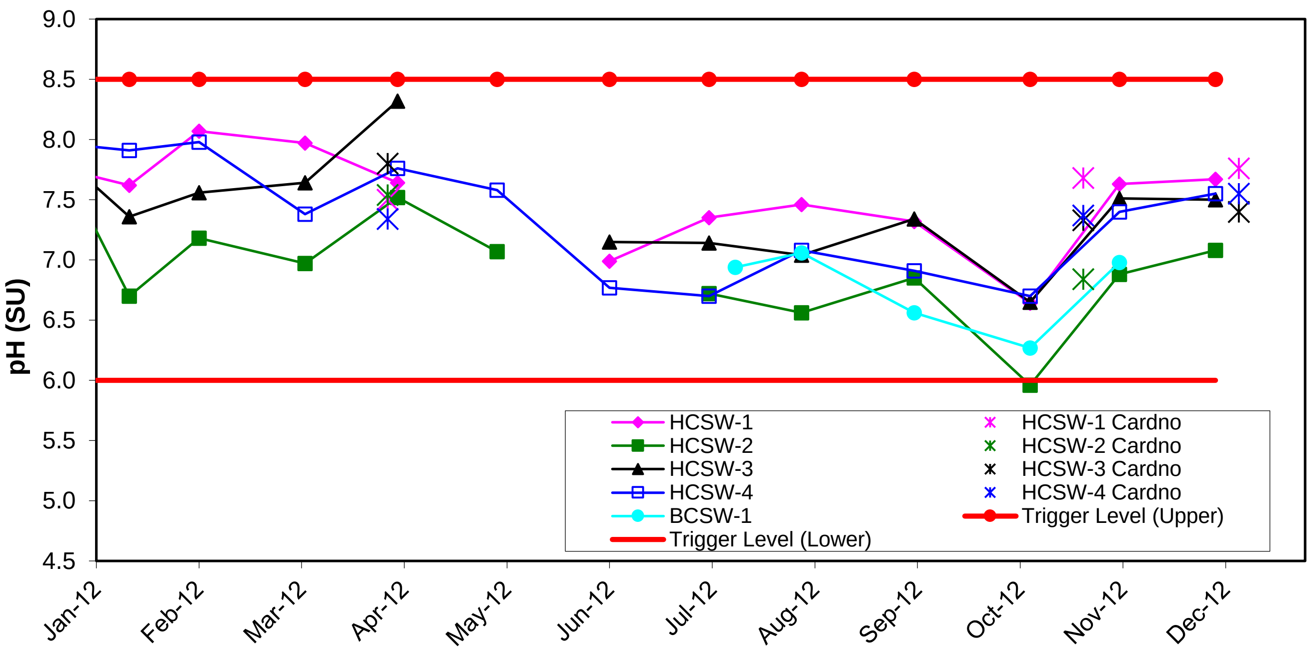

Figure 18. Values of pH Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events in 2012.

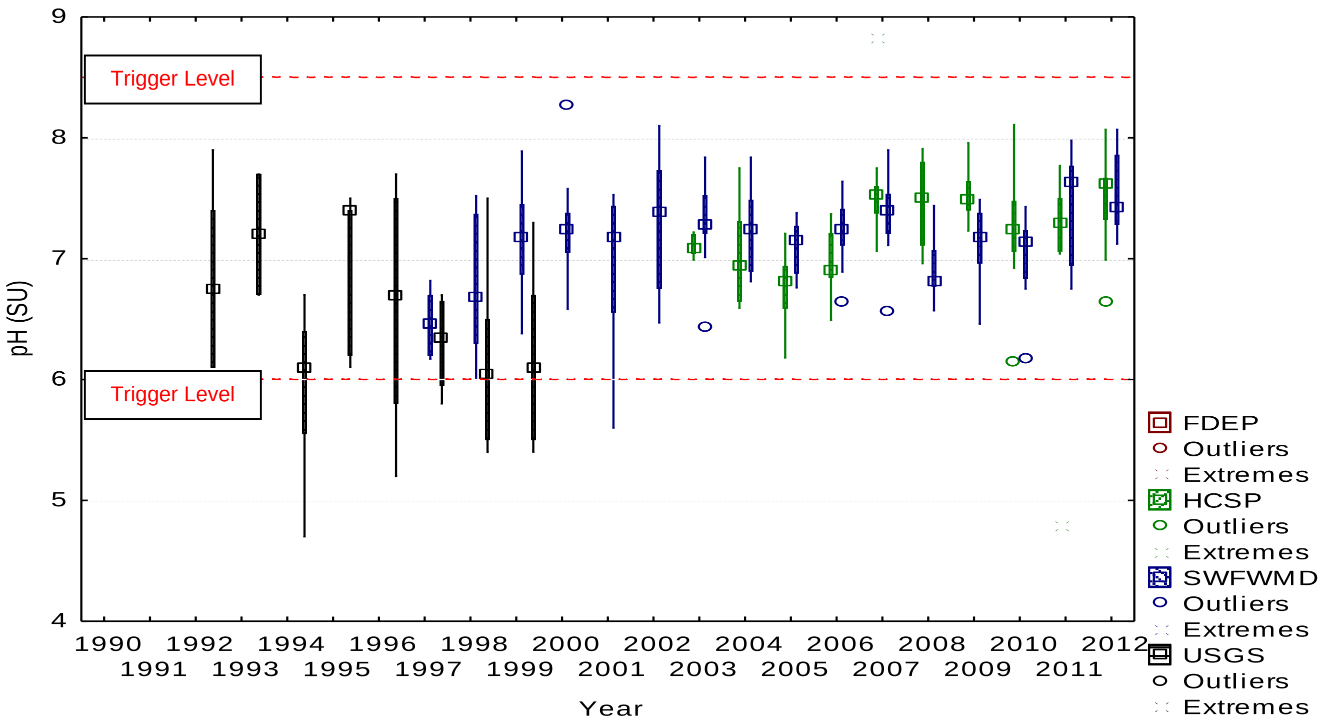

Figure 19. **HCSW-1 Values of pH Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

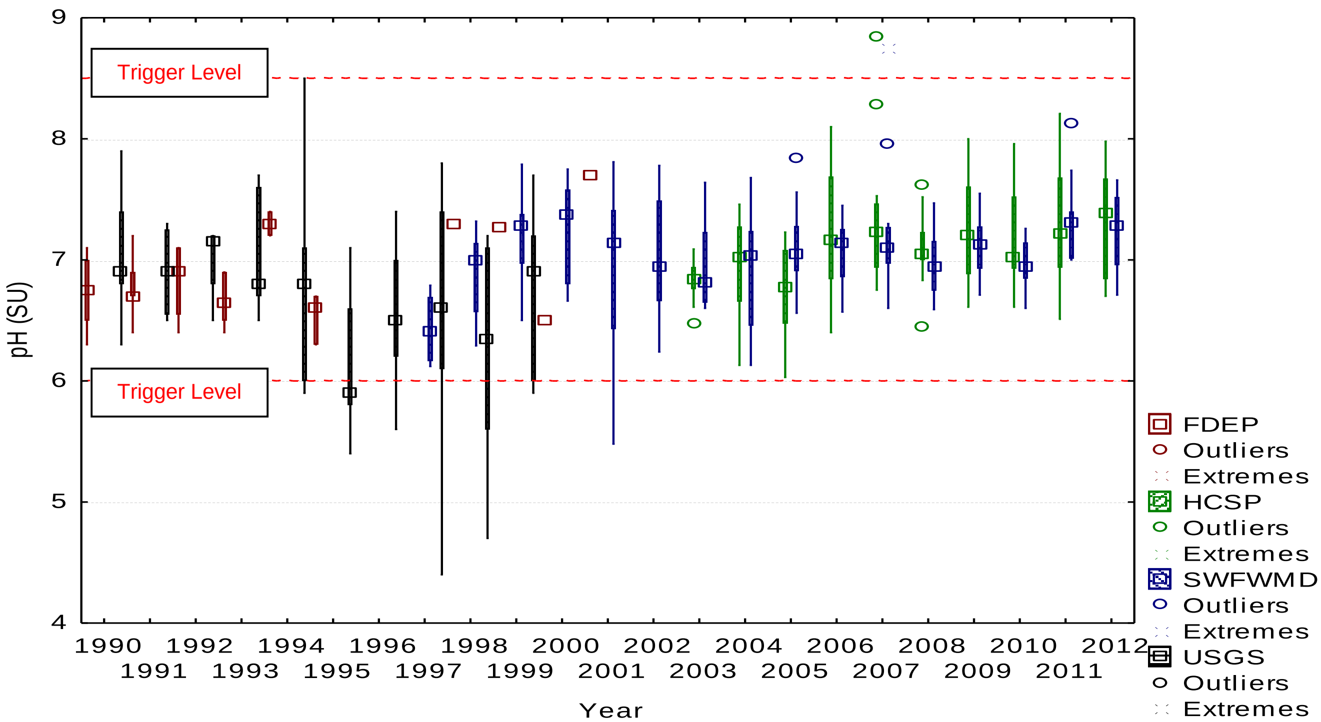

Figure 20. **HCSW-4 Values of pH Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

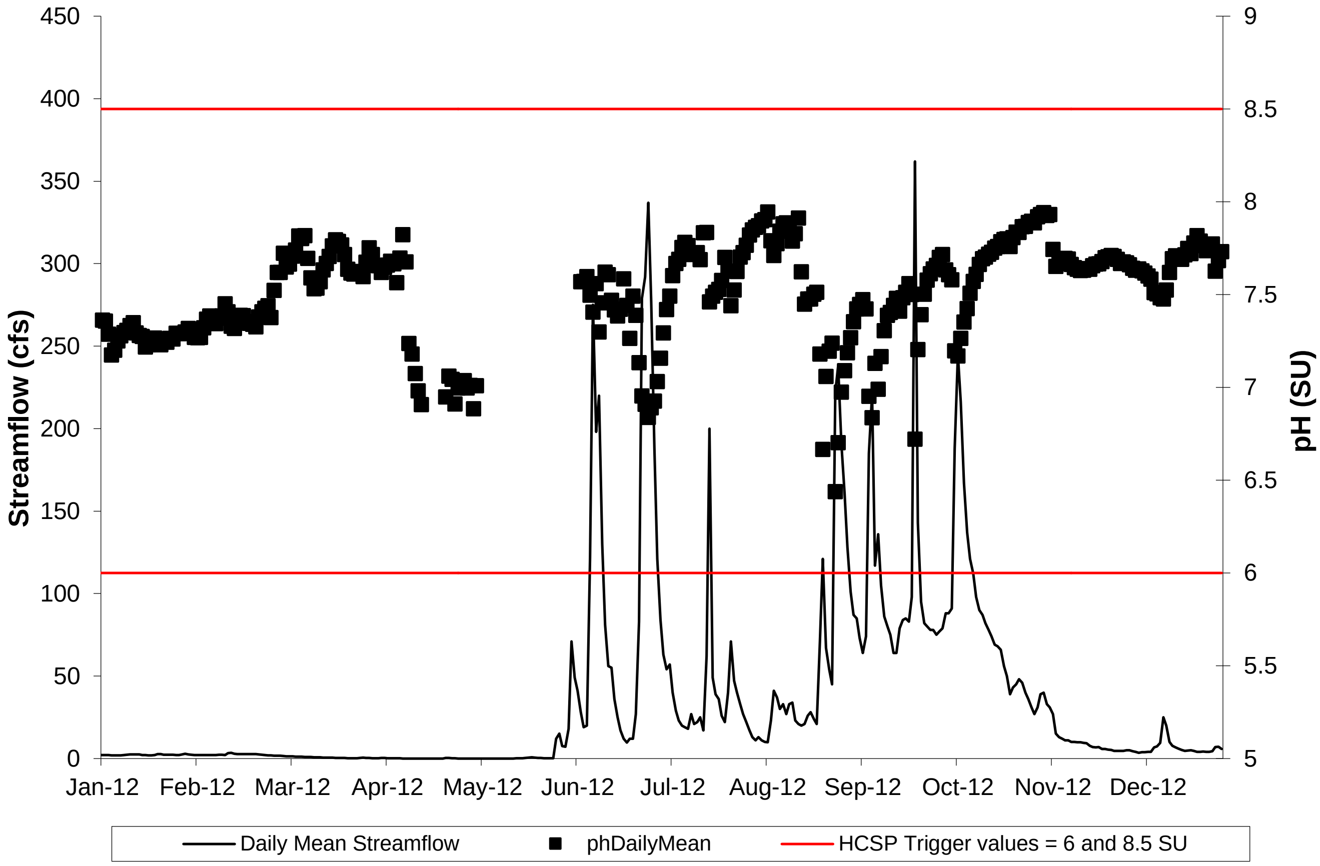

Figure 21. Relationship Between Daily Mean pH (Obtained From the Continuous Recorder at HCSW-1) and Daily Mean Streamflow for 2012. Minimum pH Detection Limit = 1 SU.

Dissolved Oxygen

Dissolved oxygen (DO) concentrations were above the trigger level and Class III Standard of 5.0 mg/l (indicating desirable conditions) during all sampling events in 2012 at HCSW-1 (Figure 22). However, levels of DO were below 5.0 mg/l at HCSW-2 for half of the year (March, May, and October-December); this station is just downstream of the Horse Creek Prairie, a blackwater swamp that typically has low DO concentrations. The Brushy Creek location had DO values below the trigger value during four of the five sampling events that occurred (July-October), which may have also contributed to the low DO concentrations at HCSW-2. DO was below the trigger value at HCSW-3 from June through September 2012 and at HCSW-4 during July and September, corresponding to times of high temperatures and relatively normal streamflow that followed periods of low-flow and stagnant conditions. DO concentrations at HCSW-1, HCSW-2, HCSW-3, and HCSW-4 obtained during biological sampling events and from the continuous recorder at HCSW-1 were fairly consistent with those found during the monthly water quality sampling (Figures 22 and 25).

The DO levels at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources (Figures 23 and 24) and exhibited no monotonic trend between 2003 and 2012 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15). Levels of dissolved oxygen were significantly different among stations in 2003 - 2012 (ANOVA, $F = 131.09$, $p < 0.001$, Table 16), with HCSW-2 significantly lower than other stations followed by HCSW-3 (Duncan's multiple range test, $p < 0.05$). Dissolved oxygen was negatively correlated with streamflow, rainfall, and NPDES discharge at both HCSW-1 and HCSW-4 (Spearman's rank correlations $-0.41 > r > -0.71$, $p < 0.05$, Table 17). During the wet season, higher temperatures in the stream drive down the oxygen saturation, and the decomposition of woody debris washed into the stream during high rains can increase oxygen demand.

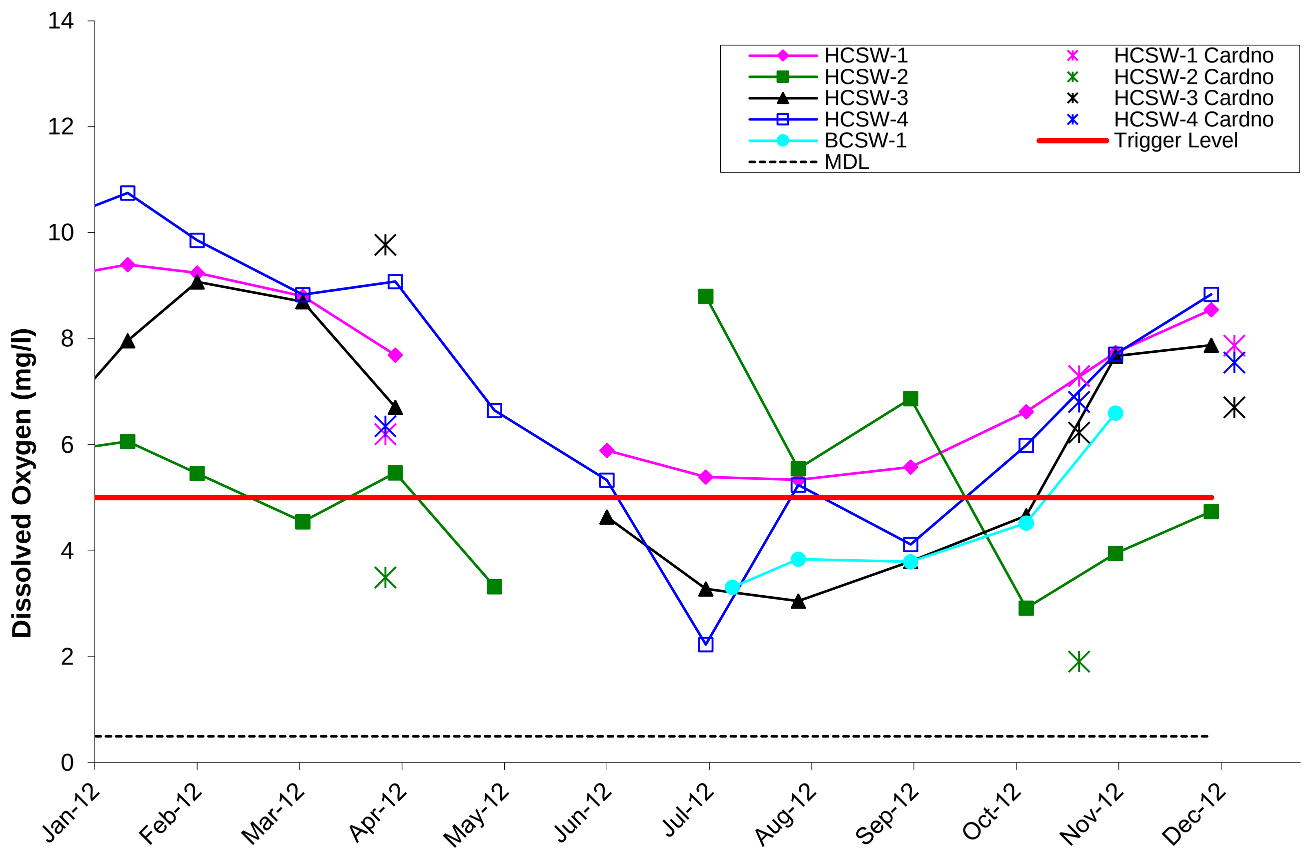

Figure 22. Dissolved Oxygen Levels Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events in 2012.

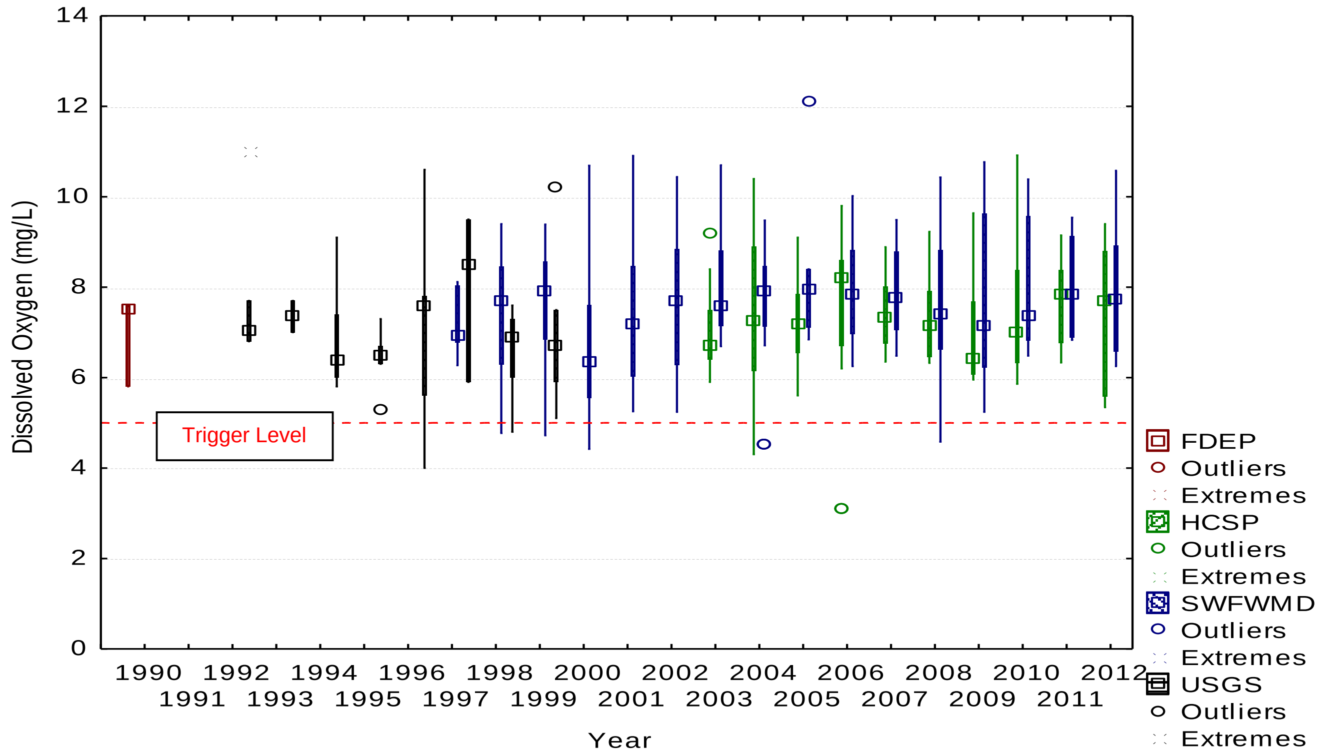

Figure 23. HCSW-1 Values of Dissolved Oxygen Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

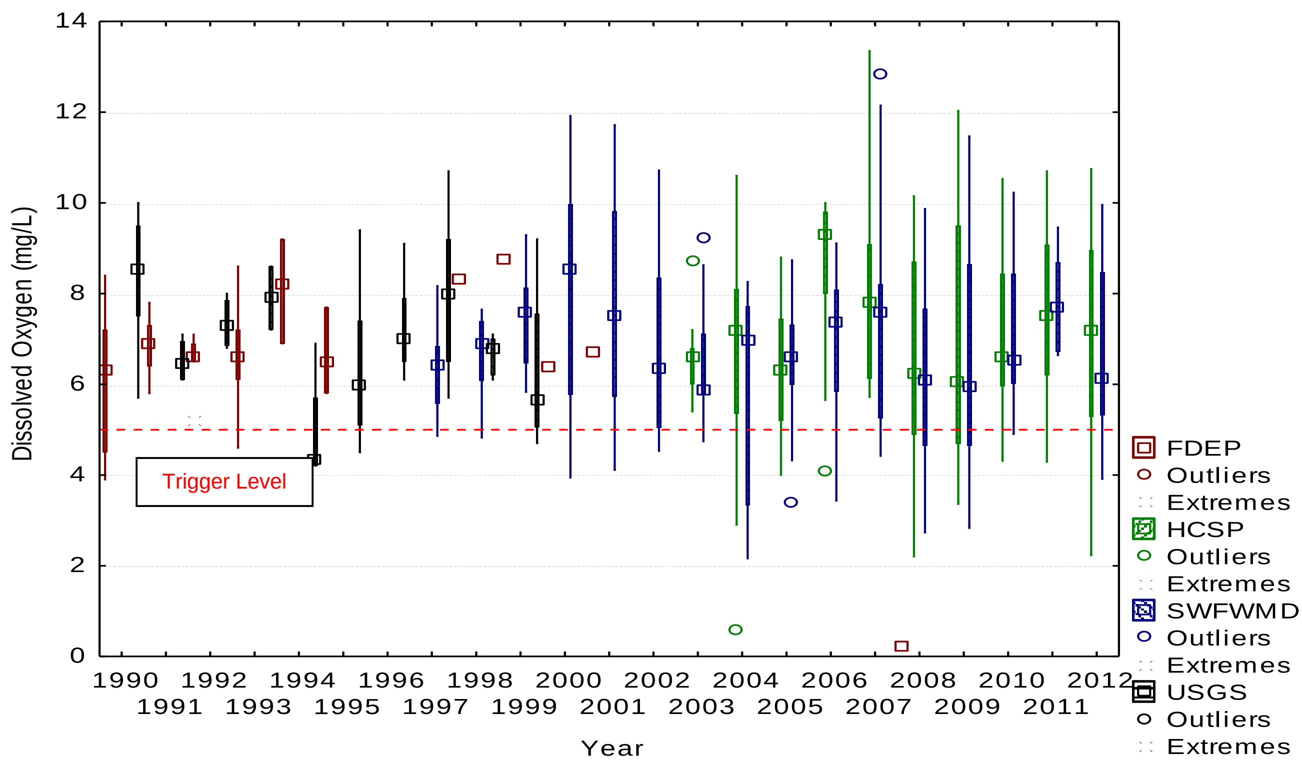

Figure 24. HCSW-4 Values of Dissolved Oxygen Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

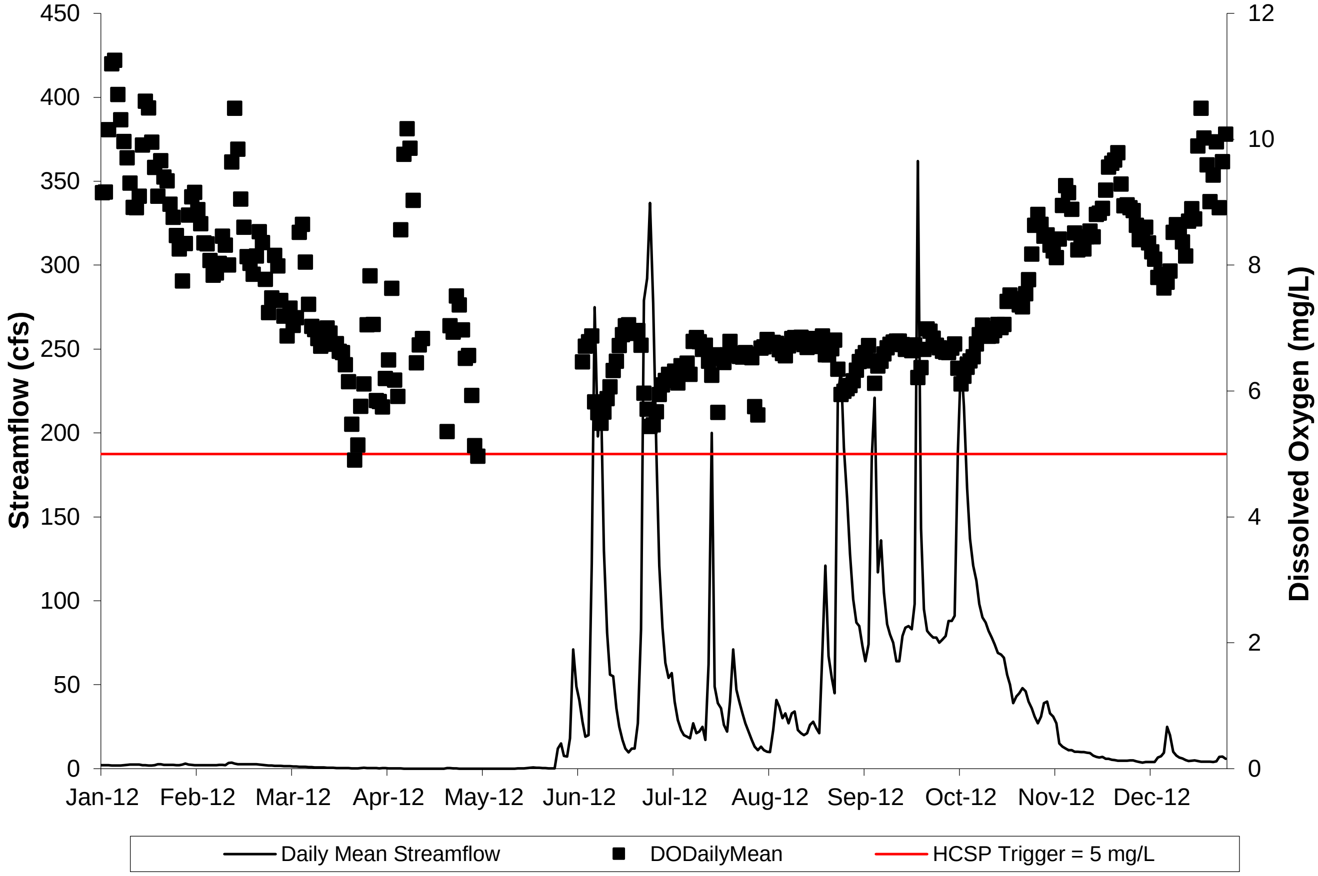

Figure 25. **Relationship Between Daily Mean DO (Obtained From the Continuous Recorder at HCSW-1) and Daily Mean Streamflow for 2012. Minimum DO Detection Limit = 0.50 mg/L.**

Turbidity

Turbidity levels obtained during biological sampling events at HCSW-1 were similar to those found during monthly water quality sampling events (Figure 26). Turbidity measured with the continuous recorder from mid-August through December 2012 was similar to the monthly measurements at HCSW-1 (Figure 29). The higher turbidity measurements from January through May 2012 were likely caused by the low water levels and the proximity of the probes to the creek bottom. Turbidity levels at all stations in 2012 were below the trigger level and Class III Surface Water Quality Standard of 29 nephelometric turbidity units (NTUs).

The turbidity levels at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15, Figures 27 and 28). Turbidity levels as measured monthly were not significantly different among stations in 2003 - 2012 (ANOVA, $F = 2.02$, $p = 0.11$, Table 16). Turbidity was positively correlated with streamflow, rainfall, and NPDES discharge at HCSW-1 and HCSW-4 (Spearman's rank correlations $0.31 < r < 0.60$, Table 17). Turbidity measurements at Brushy Creek were similar to the Horse Creek stations.

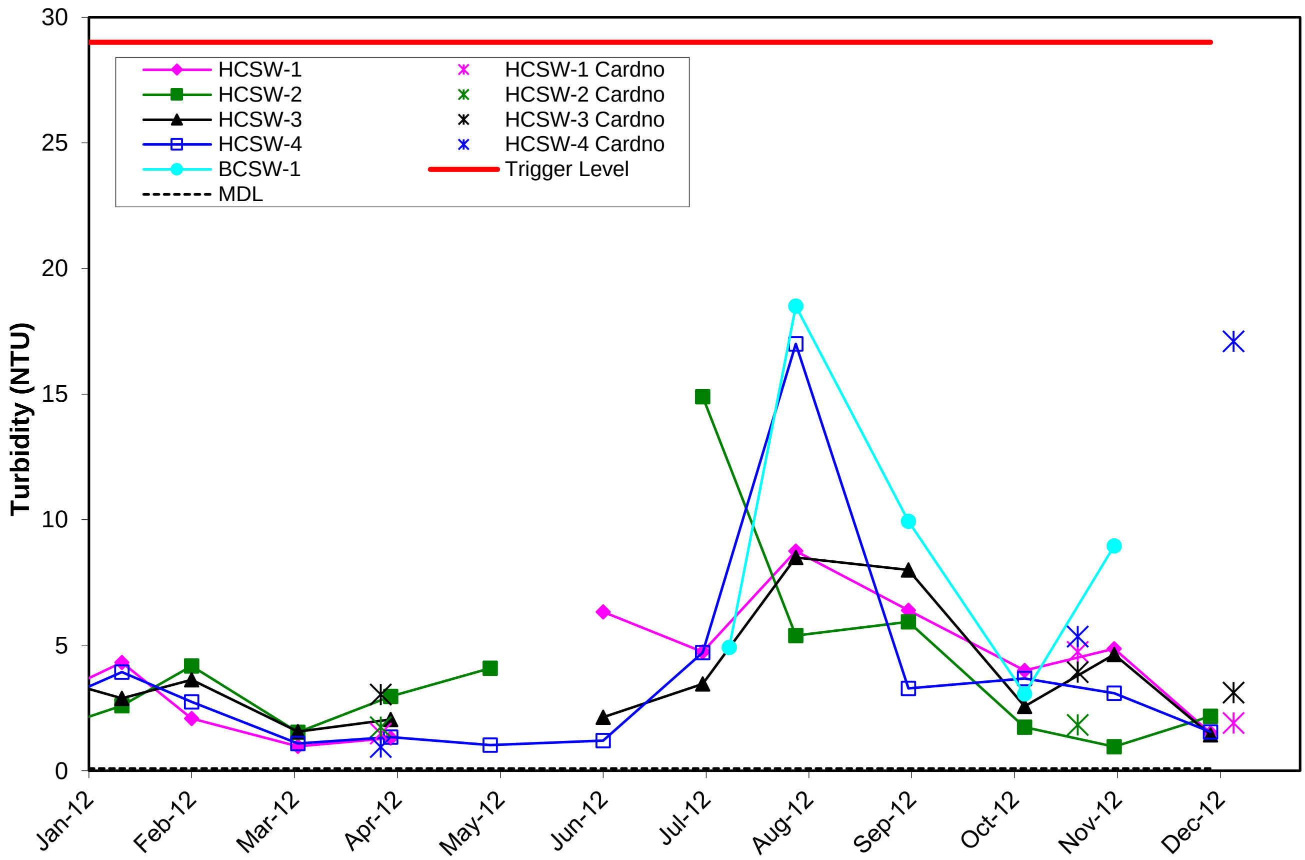

Figure 26. Turbidity Levels Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events in 2012.

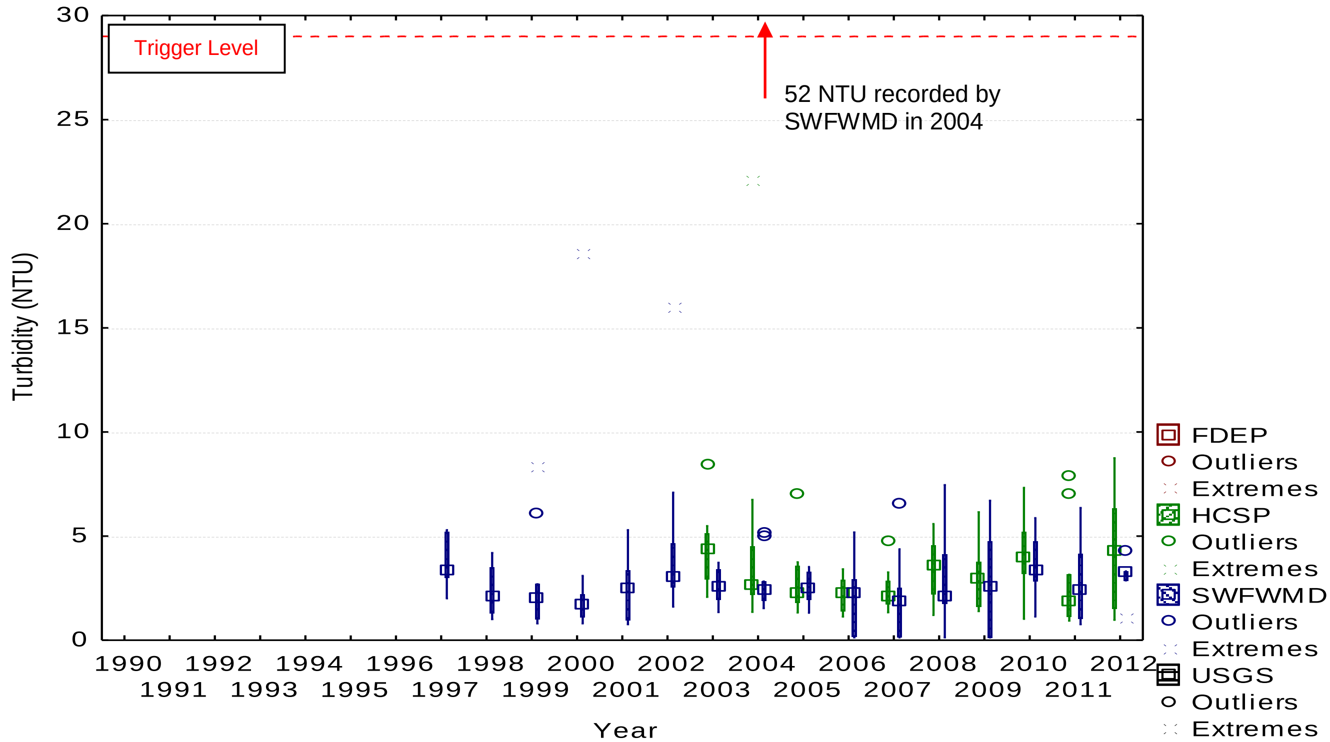

Figure 27. **HCSW-1 Values of Turbidity Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

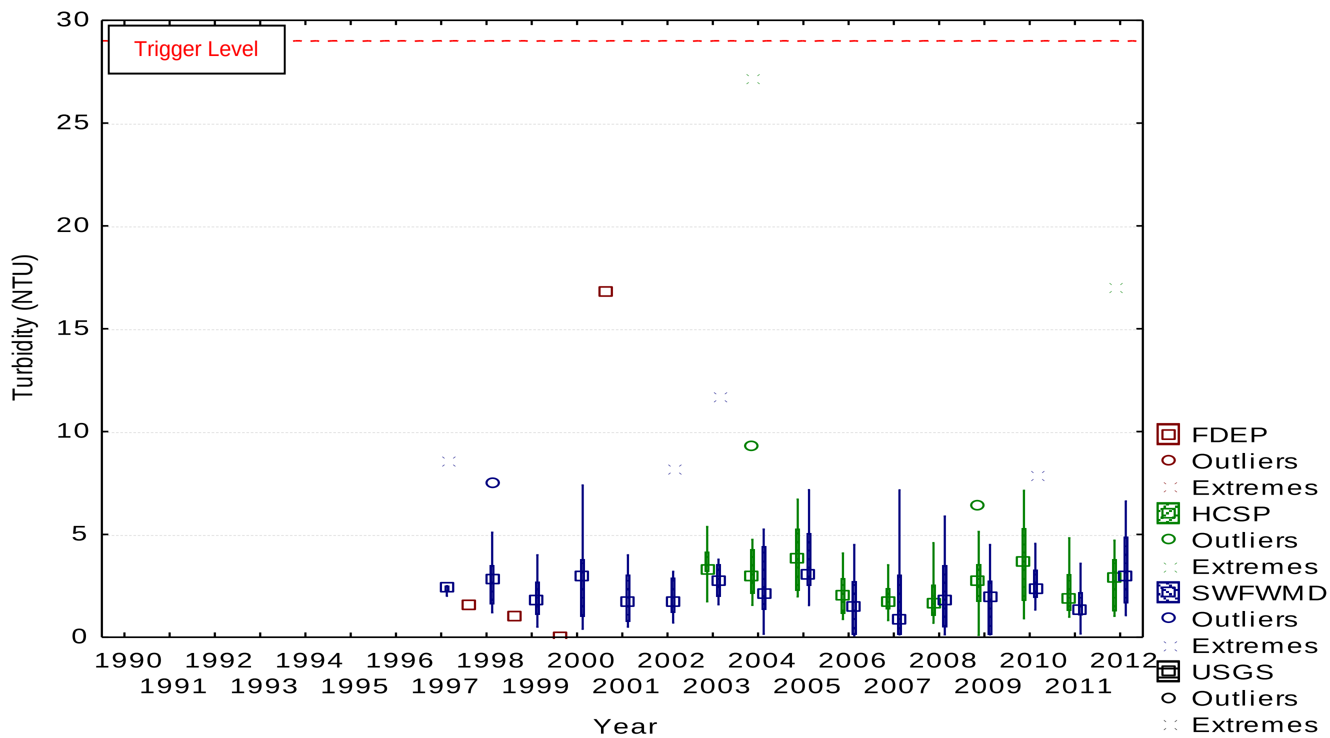

Figure 28. **HCSW-4 Values of Turbidity Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

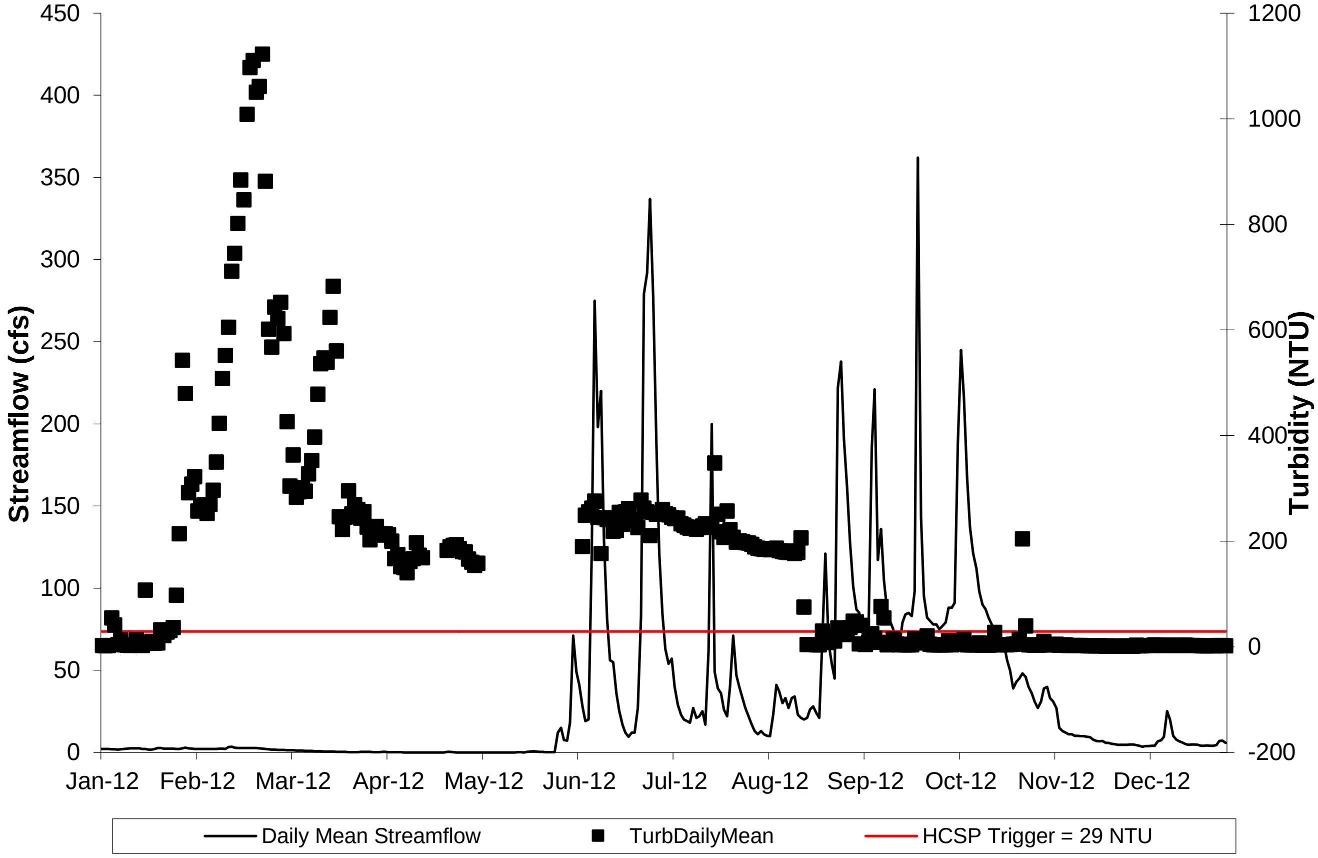

Figure 29. Relationship Between Daily Mean Turbidity (Obtained From the Continuous Recorder at HCSW-1) and Daily Mean Streamflow for 2012. Minimum Detection Limit = 0.1 NTU.

Color

All color values in 2012 were above the trigger level of 25 Platinum-Cobalt units (PCU) (indicating desirable conditions) during all events at all stations (Figure 30). The color levels at HCSW-1 measured by the HCSP are consistent with other water quality data sources and exhibited a slight increasing monotonic trend from 2003-2012 (Seasonal Kendall Tau with LOWESS, Tau = 0.32, p = 0.03, Sen slope = 5.25 PCU per year, Table 15, Figure 31). HCSW-4 also exhibited an increasing monotonic trend over the 2003-2012 time period (Seasonal Kendall Tau with LOWESS, Tau = 0.50, p = 0.001, Sen slope = 10.6 PCU per year, Table 15, Figure 32). The trigger level for color in the HCSP is expressed as a minimum, so the observed upward trends at HCSW-1 and HCSW-4 are not of concern (Appendix I) as it relates to a defined trigger level; over time, the program will continue to monitor this trend.

Color levels were significantly different among stations in 2003 - 2012 (ANOVA, F = 6.58, p < 0.001, Table 16), with HCSW-2 having higher color than other stations (Duncan's multiple range test, p < 0.05). HCSW-2 receives input from Horse Creek Prairie which contributes higher color levels to this station. Brushy Creek generally has higher color than the Horse Creek stations and also flows into Horse Creek above HCSW-2. Color was positively correlated with streamflow, rainfall, and NPDES discharge at HCSW-1 and HCSW-4 (Spearman's rank correlations 0.29 < r < 0.71, Table 17).

The similar pattern among the stations, with higher color in the wet, summer months, and lower levels in the dry, winter months, suggest that color is affected by the differential inputs of surface water and groundwater seepage. During the wet season when surface flows from wetland areas are highest, the transport of tannins to Horse Creek adds more color to the water (Reid and Wood 1976). This was very evident during and after Hurricane Charley in 2004. As the dry season begins, groundwater seepage provides a proportionally higher contribution of clearer water to Horse Creek, thereby decreasing the color of the water. It is likely that agricultural irrigation return flows also have some impact on color in the stream by introducing clearer water during the drier parts of the year or during dry years like 2006 and 2007. This agricultural influence is also noted below with respect to several other parameters.

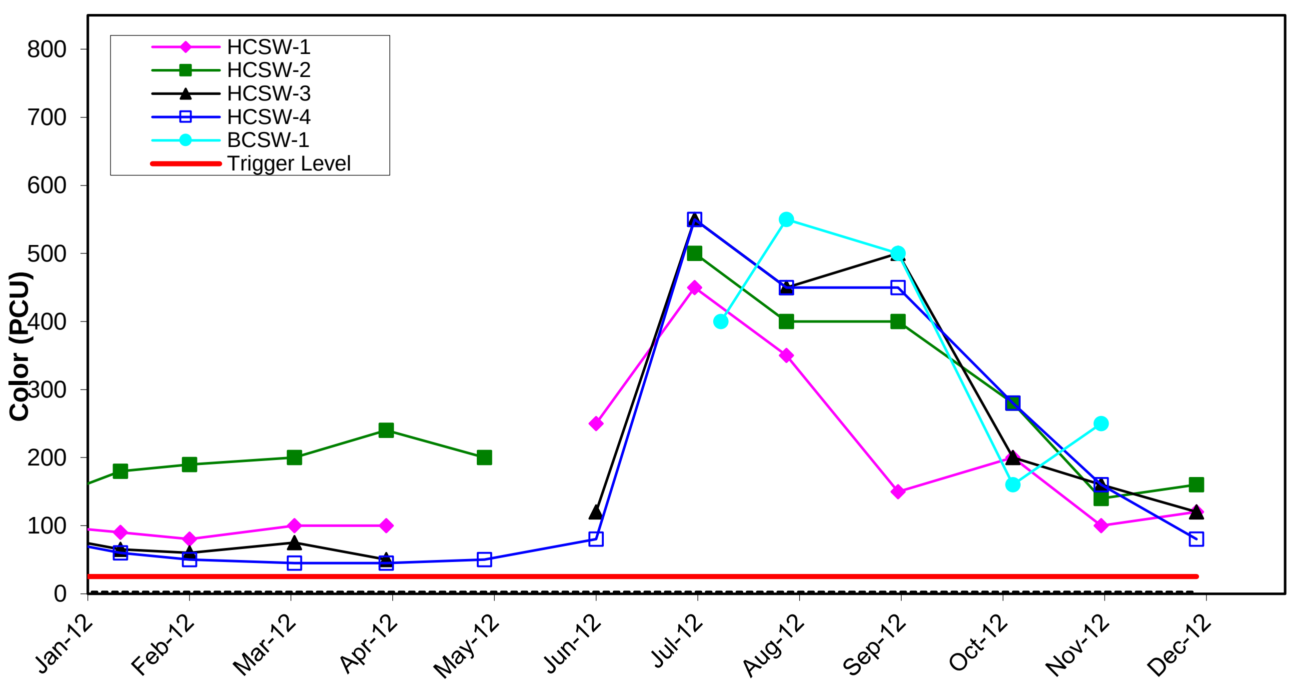

Figure 30. Color Levels Obtained During Monthly HCSP Water Quality Sampling in 2012.

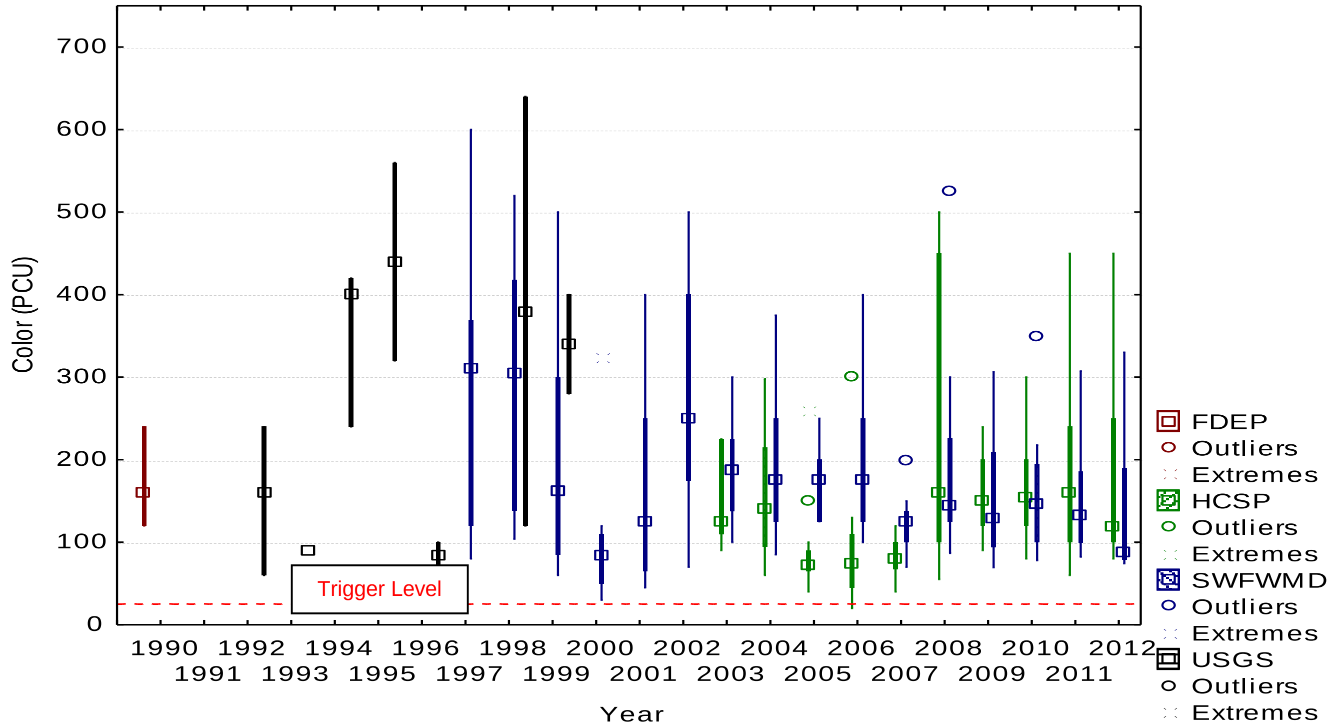

Figure 31. **HCSW-1 Values of Color Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

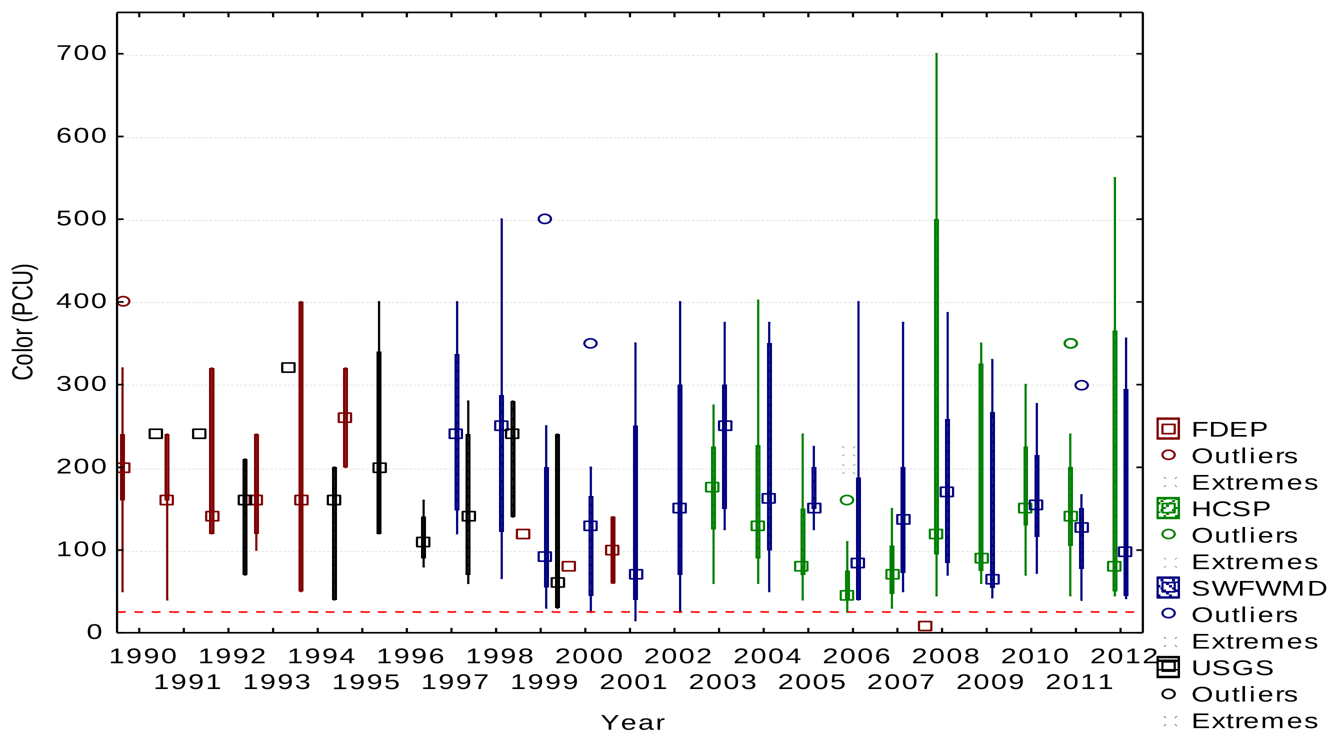

Figure 32. **HCSW-4 Values of Color Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

5.2.3 Nutrients

Total Nitrogen

Total nitrogen[13] concentrations were between 0.6 and 2.8 mg/l during all sampling events at all stations in 2012 (Figure 33). During 2012, total nitrogen was consistently below the trigger value of 3.0 mg/l. The major component of total nitrogen in nearly all samples was organic nitrogen. The total nitrogen concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Figures 34 and 35, Table 15). Total nitrogen concentrations were significantly different among stations for 2003 – 2012 (ANOVA, $F = 8.68$, $p < 0.001$, Table 16), with lower concentrations at HCSW-1 than other stations (Duncan's multiple range test, $p < 0.05$). Total nitrogen was positively correlated with streamflow (r_s= 0.48), rainfall (r_s= 0.51), and NPDES discharge (r_s= 0.33) at HCSW-1, while it was only positively correlated with streamflow (r_s= 0.37) at HCSW-4 (Spearman's rank correlations, $p < 0.05$, Table 17). Total nitrogen concentrations at Brushy Creek were slightly higher than concentrations at the Horse Creek stations.

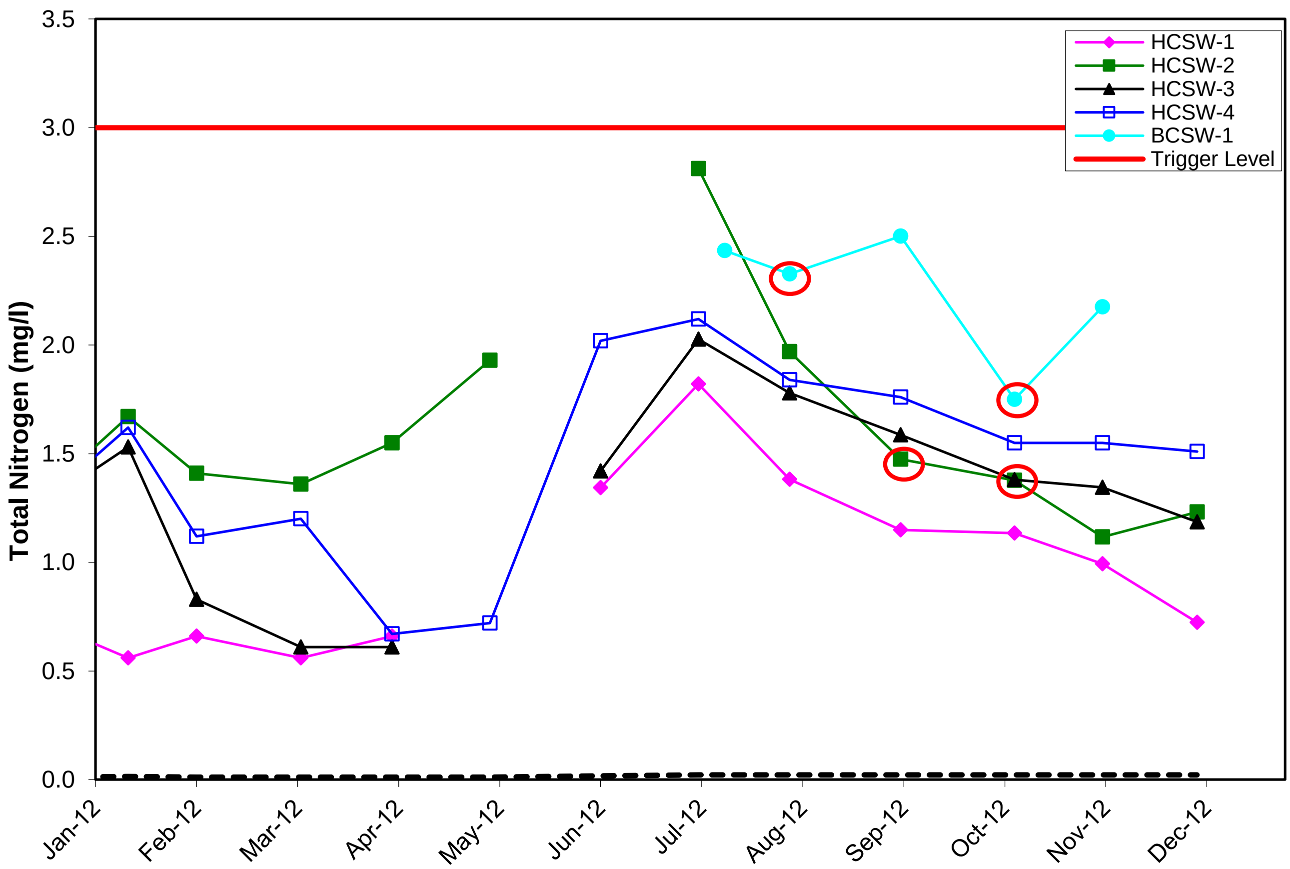

Figure 33. Total Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012. (Data from samples where Nitrate-Nitrite Nitrogen was undetected are circled in red.)

[13] Total nitrogen is calculated as the arithmetic sum of TKN and nitrate+nitrite. As requested by the PRMRWSA, if either TKN or nitrate+nitrite is undetected, the MDL of the undetected constituent will be used as part of the total nitrogen calculation. Note that this use of MDL for undetected constituents is inconsistent with typical laboratory and DEP SOPs and may result in artificially high estimates of total nitrogen.

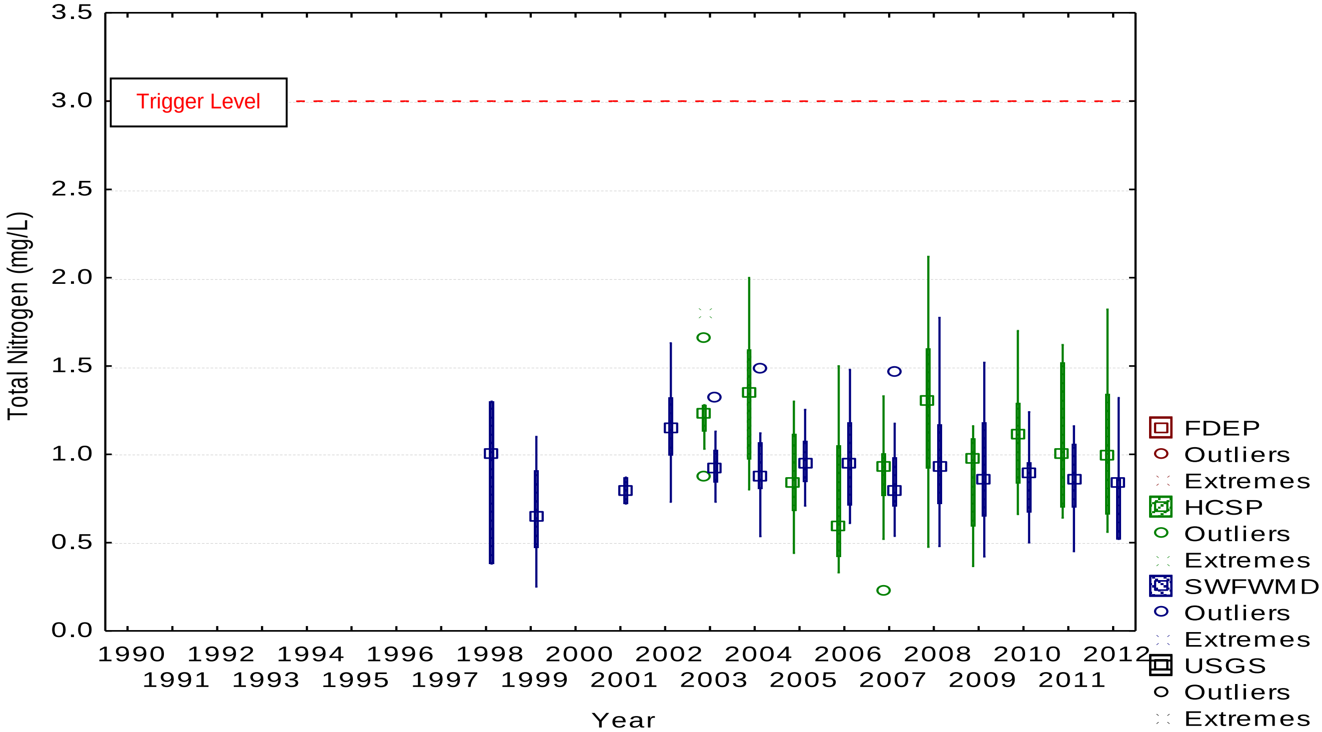

Figure 34. HCSW-1 Values of Total Nitrogen Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

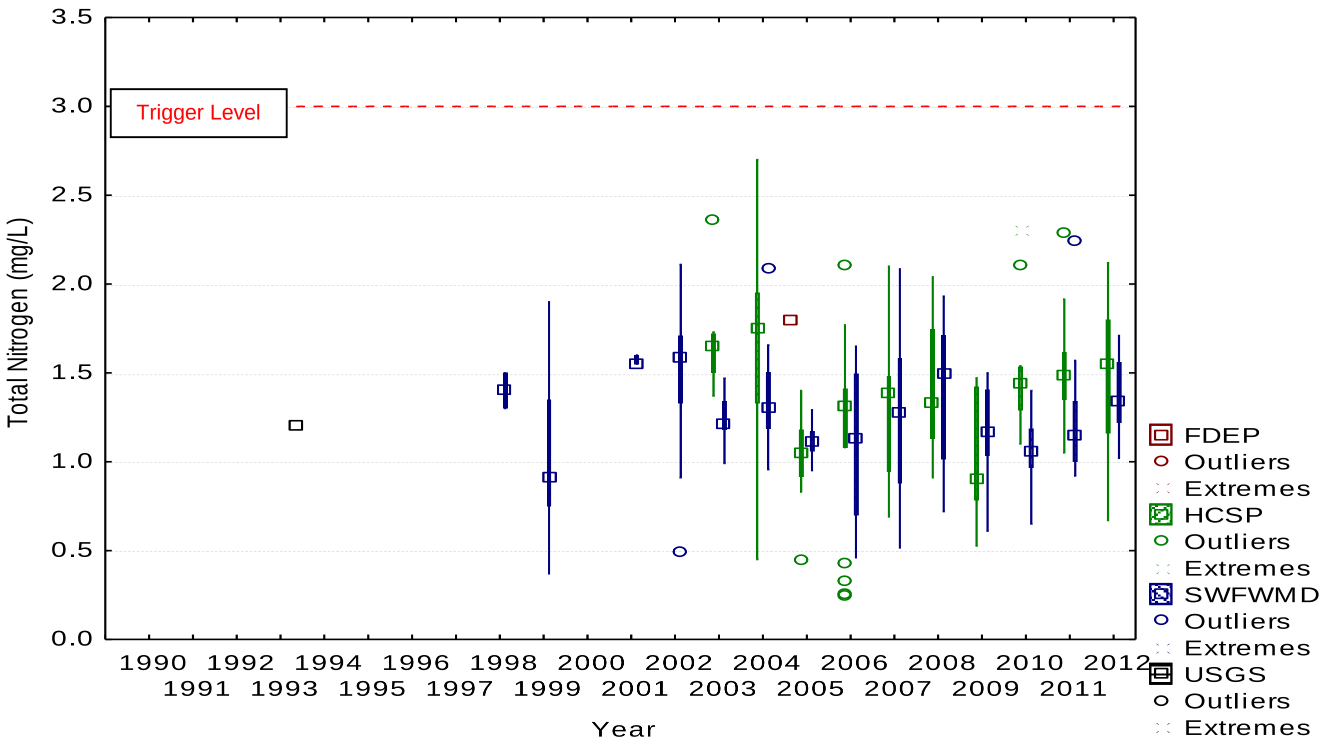

Figure 35. HCSW-4 Values of Total Nitrogen Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

Total Kjeldahl Nitrogen

Total Kjeldahl nitrogen (TKN) comprised the majority of total nitrogen in most samples (Figure 36, compare with Figure 33). The HCSP does not have an independent trigger value for TKN. The total Kjeldahl nitrogen concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Figures 37 and 38, Table 15). Concentrations of TKN were significantly different among stations (ANOVA, $F = 16.94$, $p < 0.001$, Table 16), with HCSW-2 having a higher concentration than the other three stations (Duncan's multiple range test, $p < 0.05$). Brushy Creek, which contributes to HCSW-2, has higher TKN concentrations than the Horse Creek stations. Total Kjeldahl nitrogen was positively correlated with streamflow, rainfall, and NPDES discharge at HCSW-1 and HCSW-4 (Spearman's rank correlations $0.36 < r < 0.59$, Table 17).

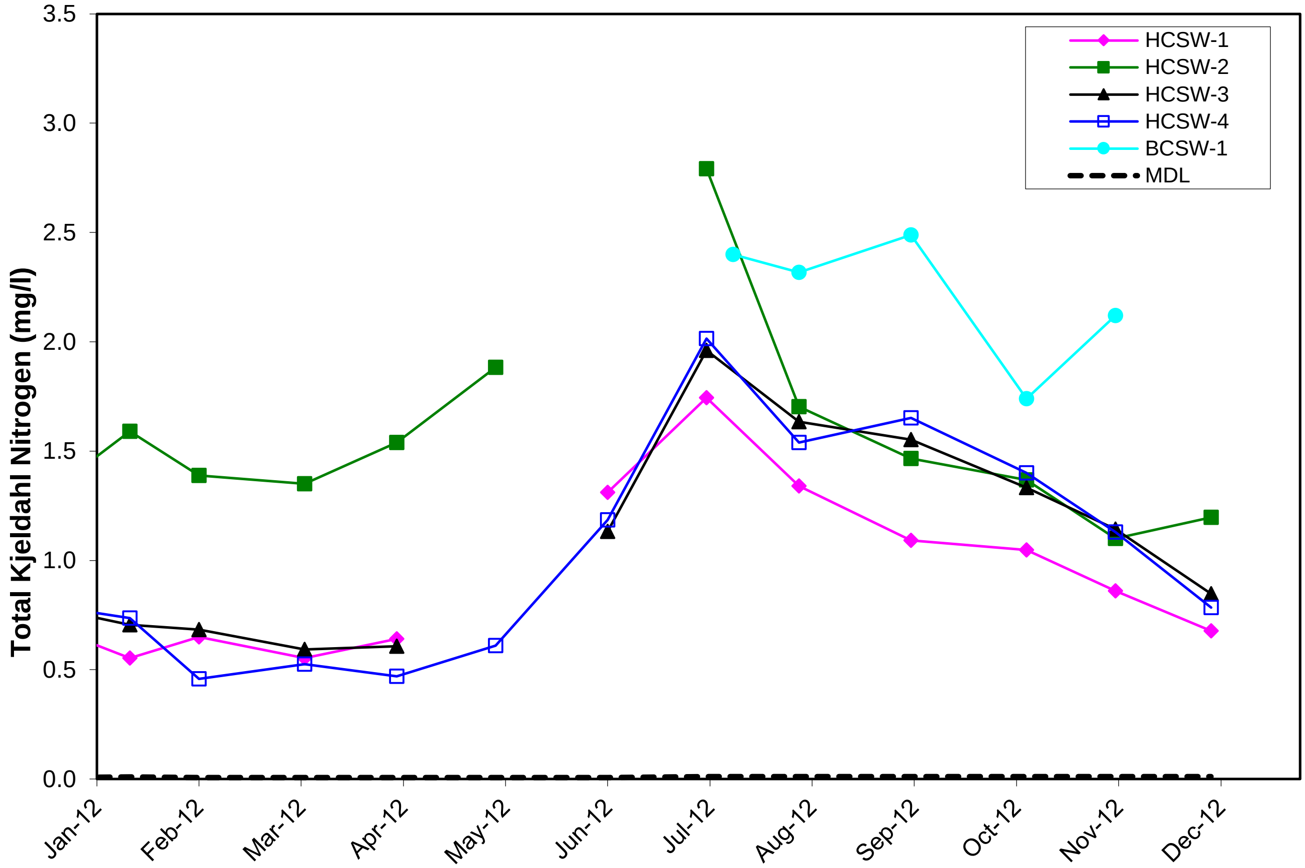

Figure 36. Total Kjeldahl Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

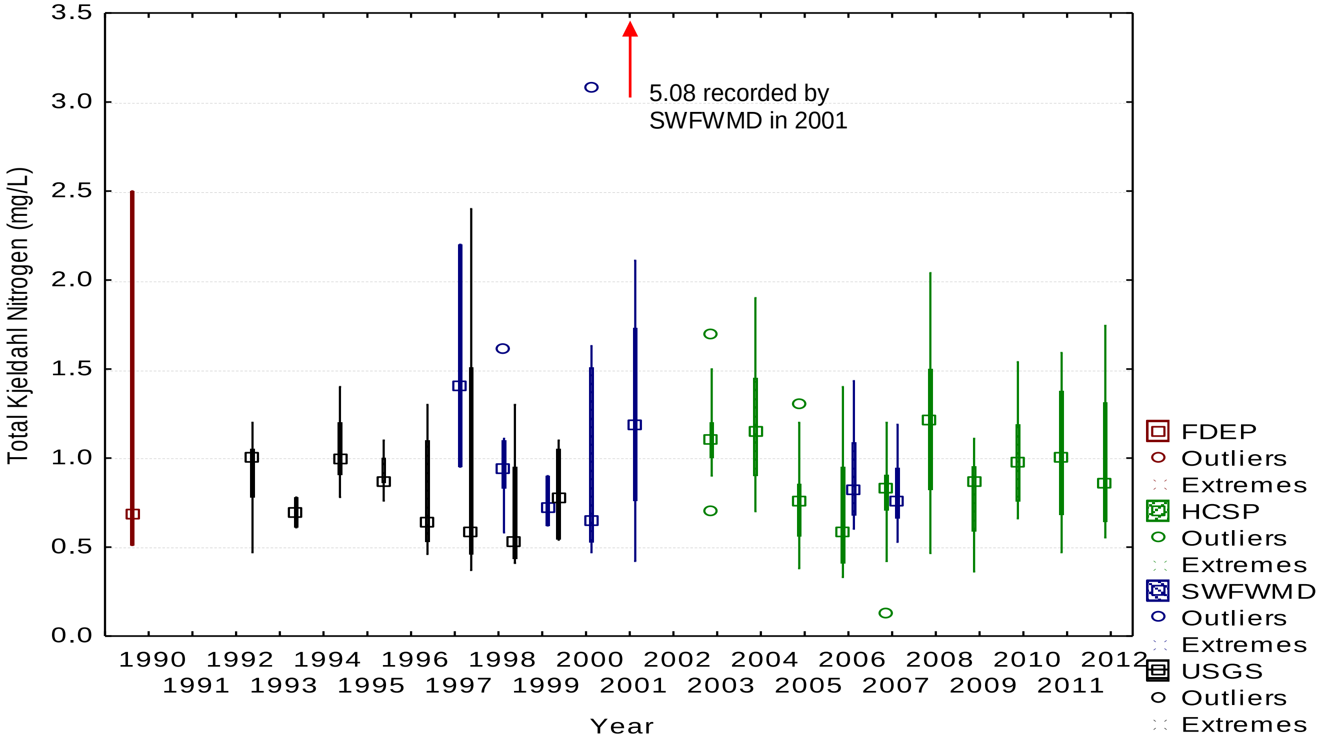

Figure 37. **HCSW-1 Values of Total Kjeldahl Nitrogen Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

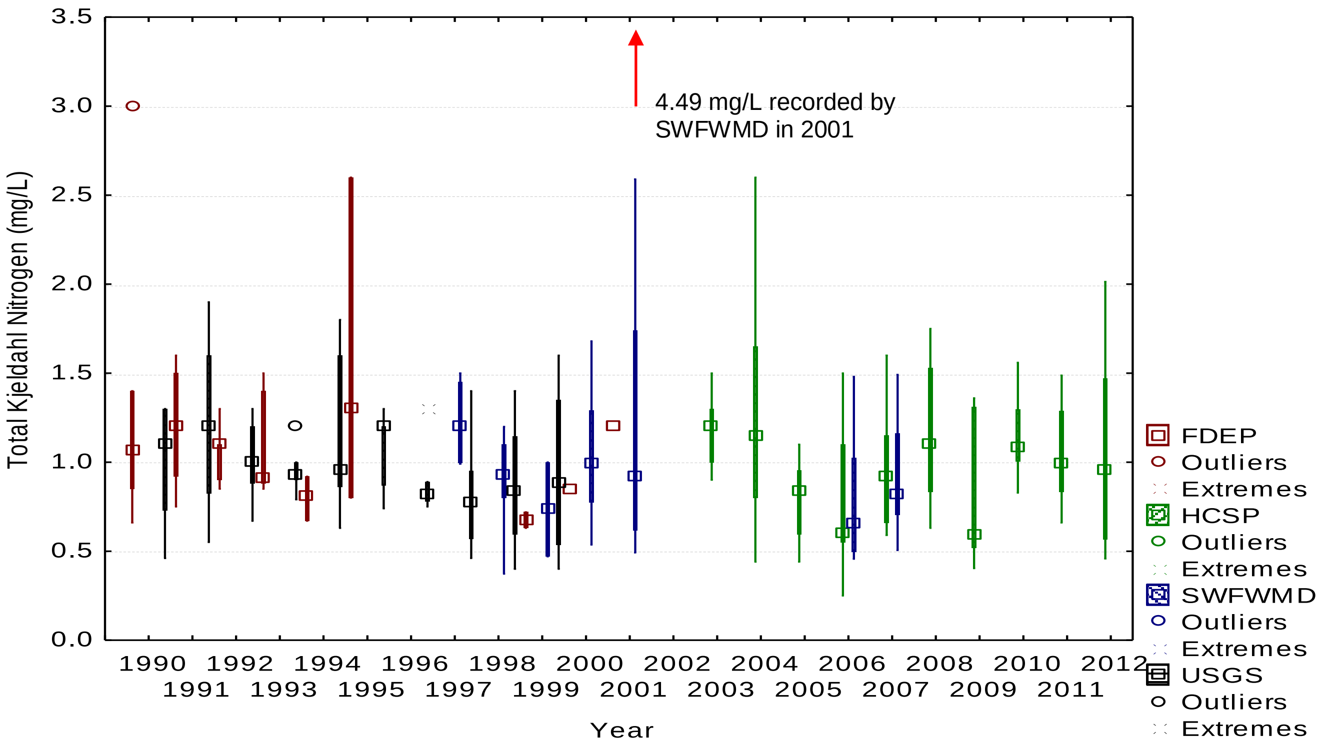

Figure 38. **HCSW-4 Values of Total Kjeldahl Nitrogen Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Nitrate-Nitrite Nitrogen

In general, nitrate+nitrite concentrations are greater at the downstream Horse Creek stations possibly because of agriculture (Figure 39). Nitrate-nitrite concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources (Figures 40 and 41), but could not be analyzed for monotonic trends or correlations with water quantity because of changes in MDL's over the course of the HCSP (Appendix C). Based on an alternate trend analysis performed on annual medians of data collected by SWFWMD from 2003-2012, there are no monotonic trends in nitrate-nitrite for HCSW-1 or HCSW-4 (Annual Kendall Tau with LOWESS, $p > 0.05$, Table 15).

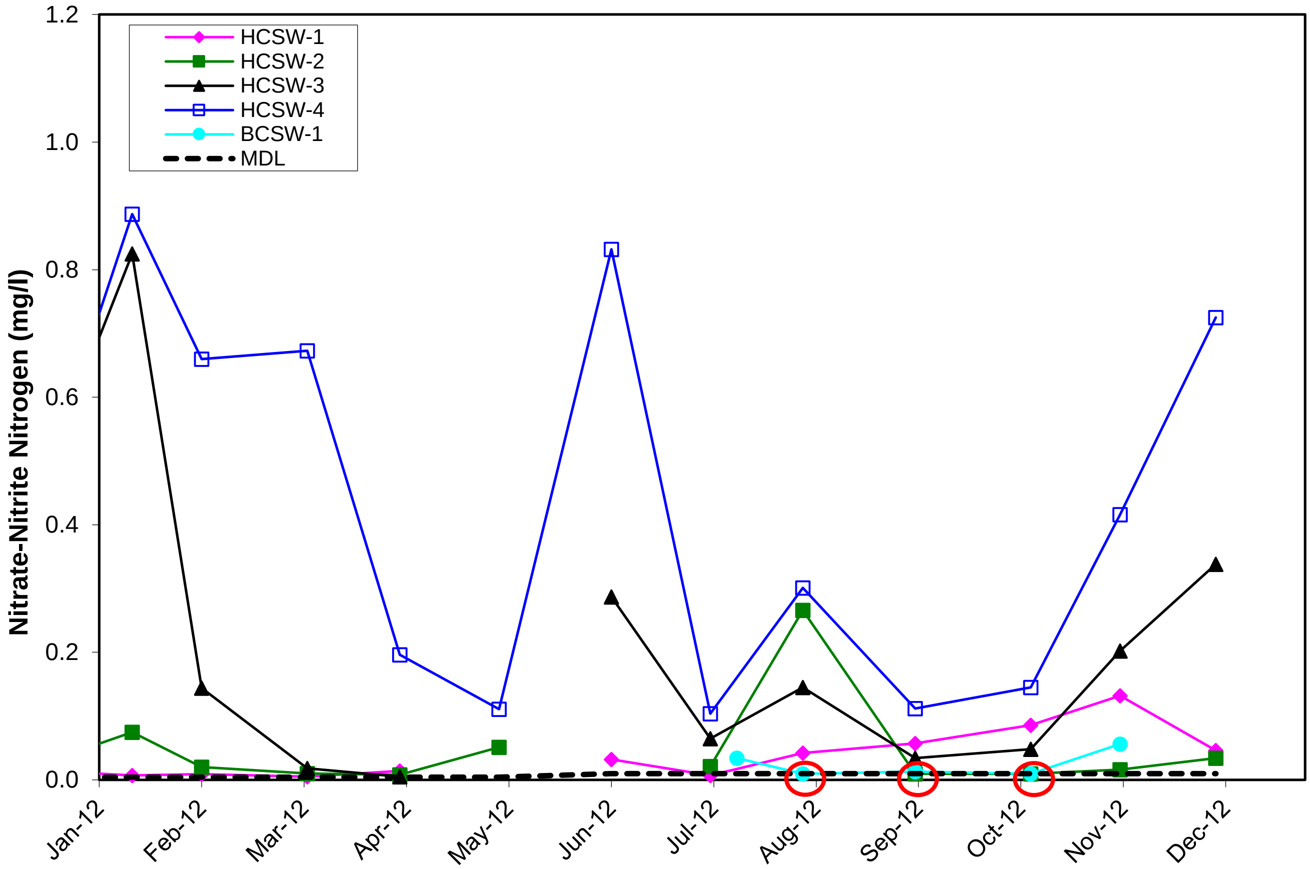

Figure 39. Nitrate-Nitrite Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012. (Data from samples where Nitrate-Nitrite Nitrogen was undetected are circled in red.)

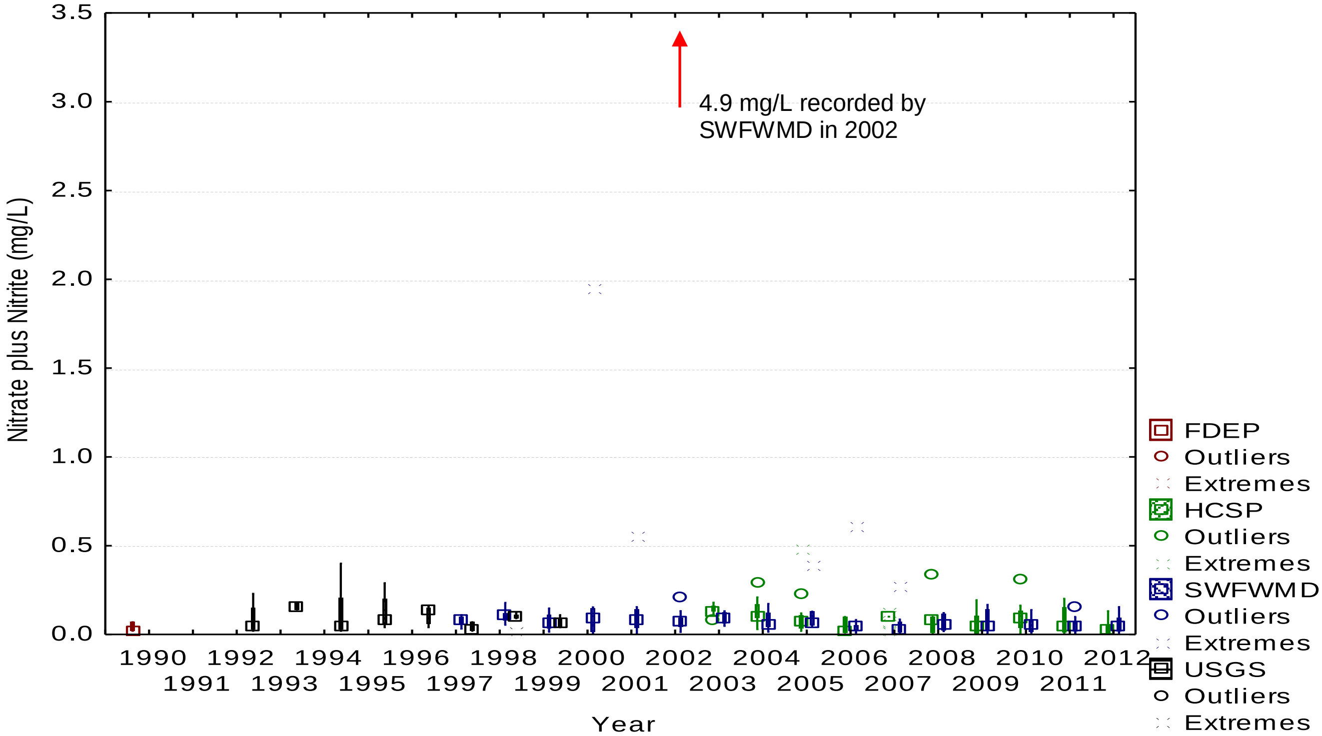

Figure 40. HCSW-1 Values of Nitrate plus Nitrite Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

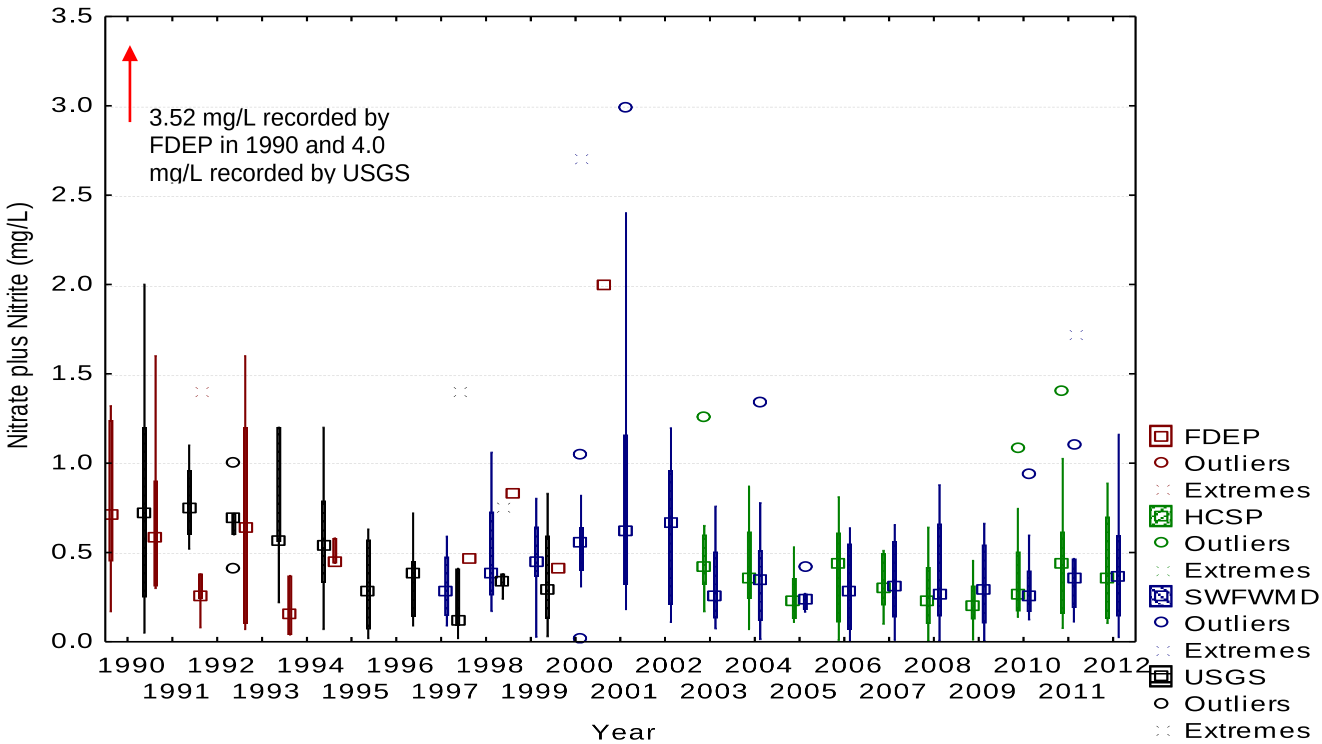

Figure 41. HCSW-4 Values of Nitrate plus Nitrite Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

Total Ammonia Nitrogen

Total ammonia nitrogen levels were within a similar range during all sampling events at all stations, with no stations exceeding the trigger level during 2012 (Figure 42). The ammonia concentrations at HCSW-1 and HCSW-4 measured by the HCSP are at levels within the normal range for the last decade of data (Figures 43 and 44). They could not be analyzed for monotonic trends using seasonal data or correlations with water quantity because of changes in MDL's over the course of the HCSP. Based on an alternative trend analysis performed on annual medians from data collected by SWFWMD since 2003, there are no monotonic trends in total ammonia nitrogen for HCSW-4 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15). However, at HCSW-1 there is a slightly decreasing monotonic trend for total ammonia nitrogen from 2003-2012 (Annual Kendall Tau with LOWESS, Tau = -0.56, $p = 0.05$, Sen slope estimator = -0.0003 mg/L per year, Figures 43 and 44). Because the direction of this potential trend is opposite that of the HCSP trigger level, it is not of concern (Appendix I).

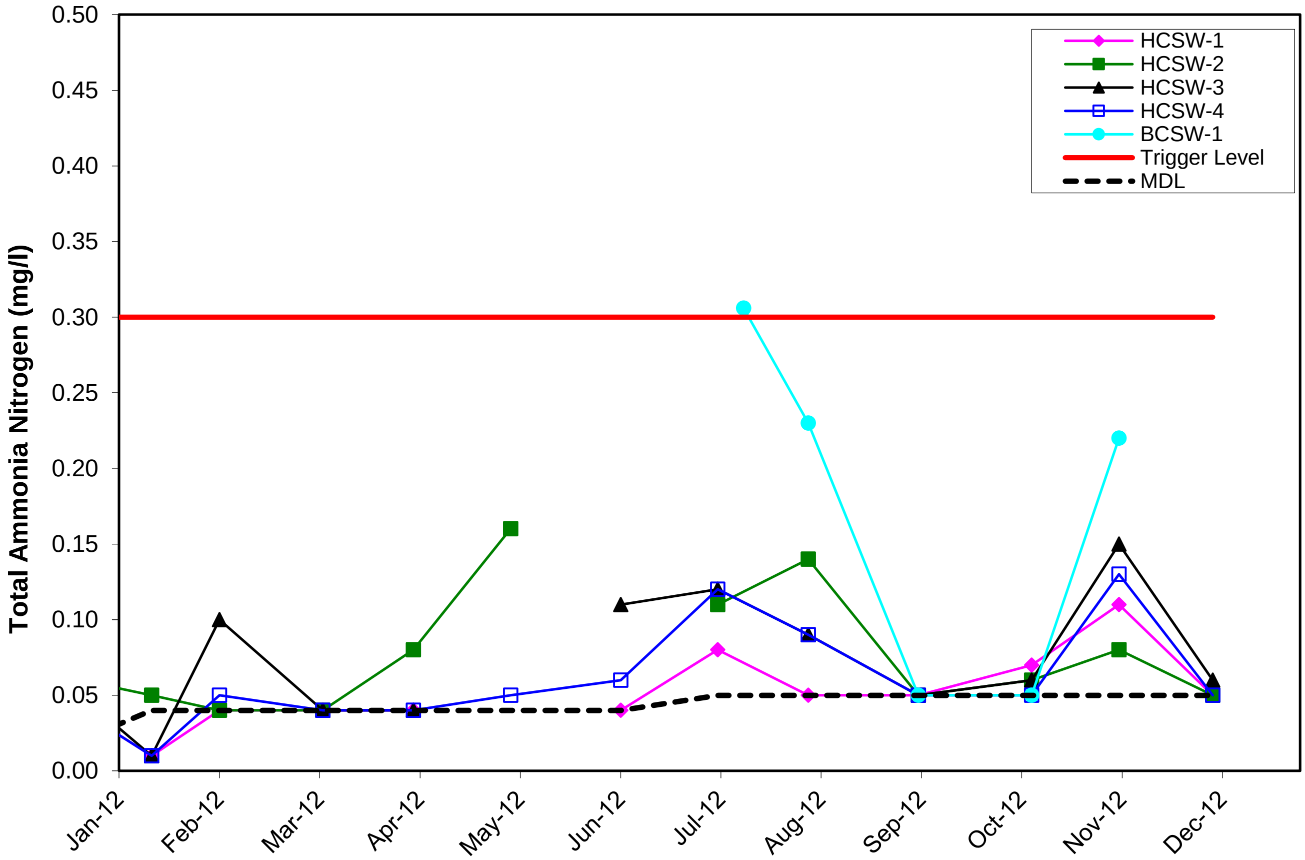

Figure 42. Total Ammonia Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

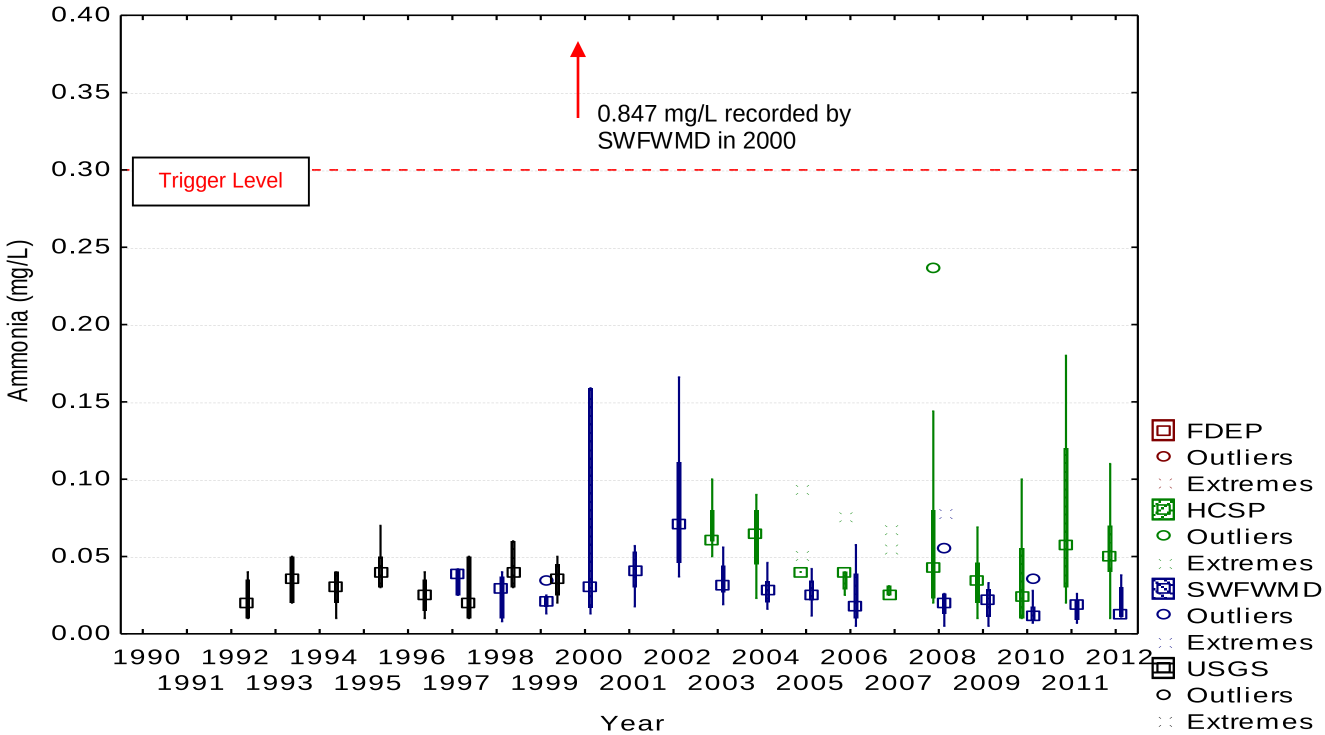

Figure 43. **HCSW-1 Values of Ammonia Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

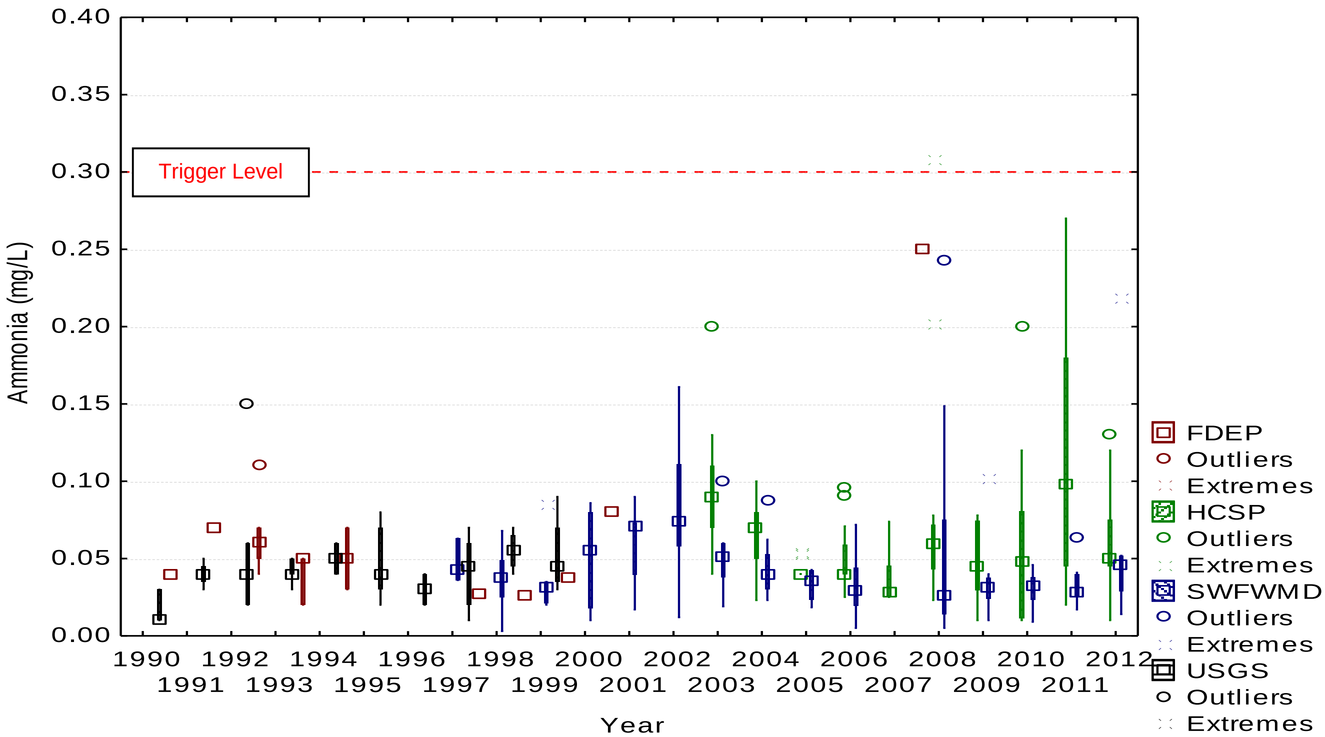

Figure 44. **HCSW-4 Values of Ammonia Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Orthophosphate

Levels of orthophosphate were well below the trigger level of 2.5 mg/l in 2012 (Figure 45). The orthophosphate concentrations at HCSW-1 measured by the HCSP are consistent with other water quality data sources and exhibited a slightly increasing monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, Tau=0.36, p=0.01, Sen slope estimator=0.02 mg/L per year). The previously noted increasing monotonic trend at HCSW-4 (2010 Annual Report, Robbins et al. 2014) was no longer significant when data from 2012 was included in the analysis (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15). The impact assessment in Appendix I shows that the apparent trend in orthophosphate from 2003-2012 is caused by a data bias, and that extending the period of record eliminates this trend. The evidence of a data bias and the very small magnitude of the trend slope indicate that orthophosphate is not a concern at this time, although the program will continue to monitor potential trends. In addition, the higher slope observed in the 2003-2010 trend analysis (0.27 mg/L/yr) was reduced to 0.02 mg/L/yr when adding on data from both 2011 and 2012 to the seasonal analysis.

Orthophosphate concentrations were significantly different among stations (ANOVA, $F = 23.7$, $p < 0.001$, Table 16), with concentrations at HCSW-2 lowest (Duncan's multiple range test, $p < 0.05$). Orthophosphate was only negatively correlated with streamflow ($r_s = -0.21$) at HCSW-1 (Figure 46), while it was not correlated with streamflow, rainfall, or NPDES discharge at HCSW-4 (Spearman's rank correlation, $p > 0.05$, Figure 47, Table 17). Orthophosphate concentrations at Brushy Creek were lower than all stations in Horse Creek during all sampling events with the exception of November 2012 (Figure 45).

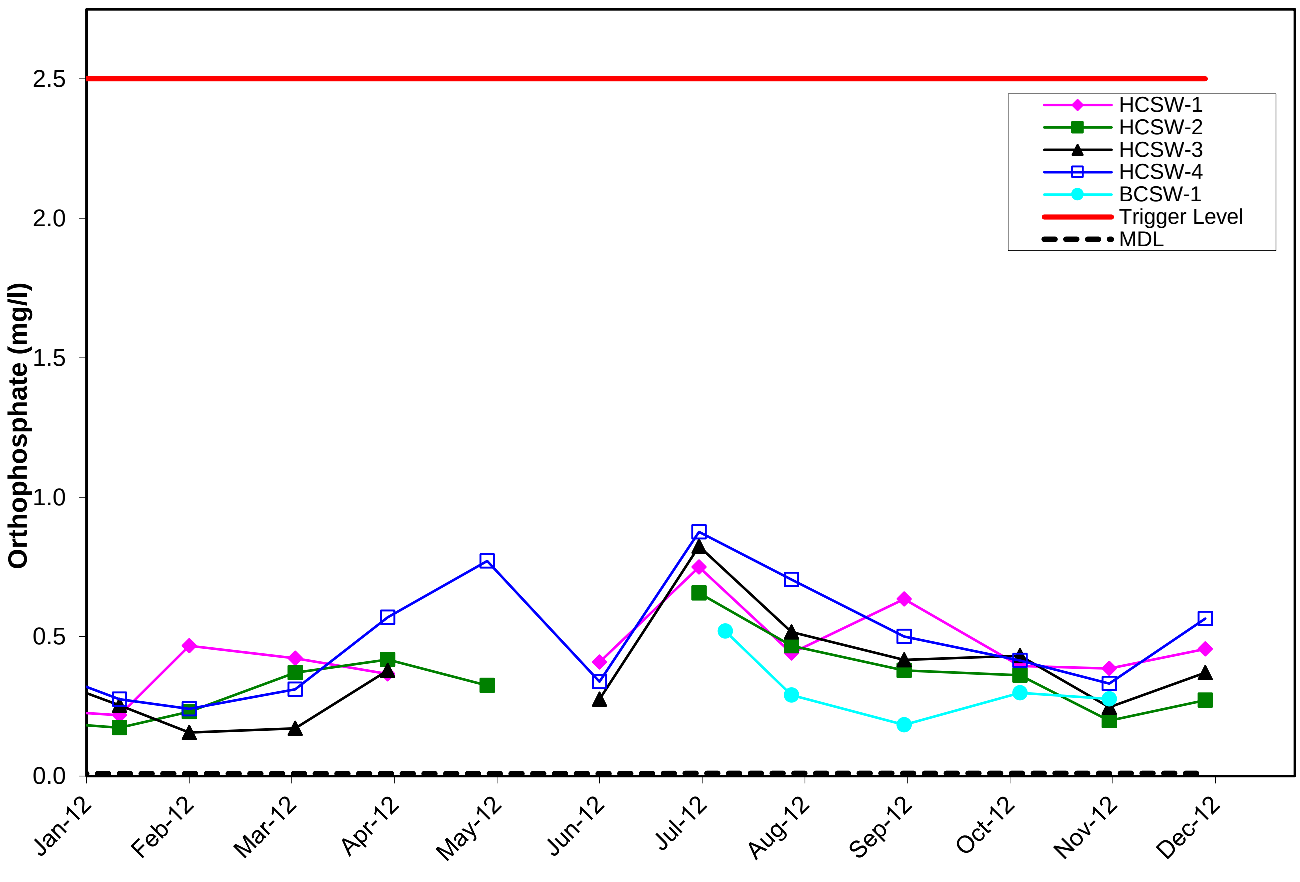

Figure 45. Orthophosphate Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

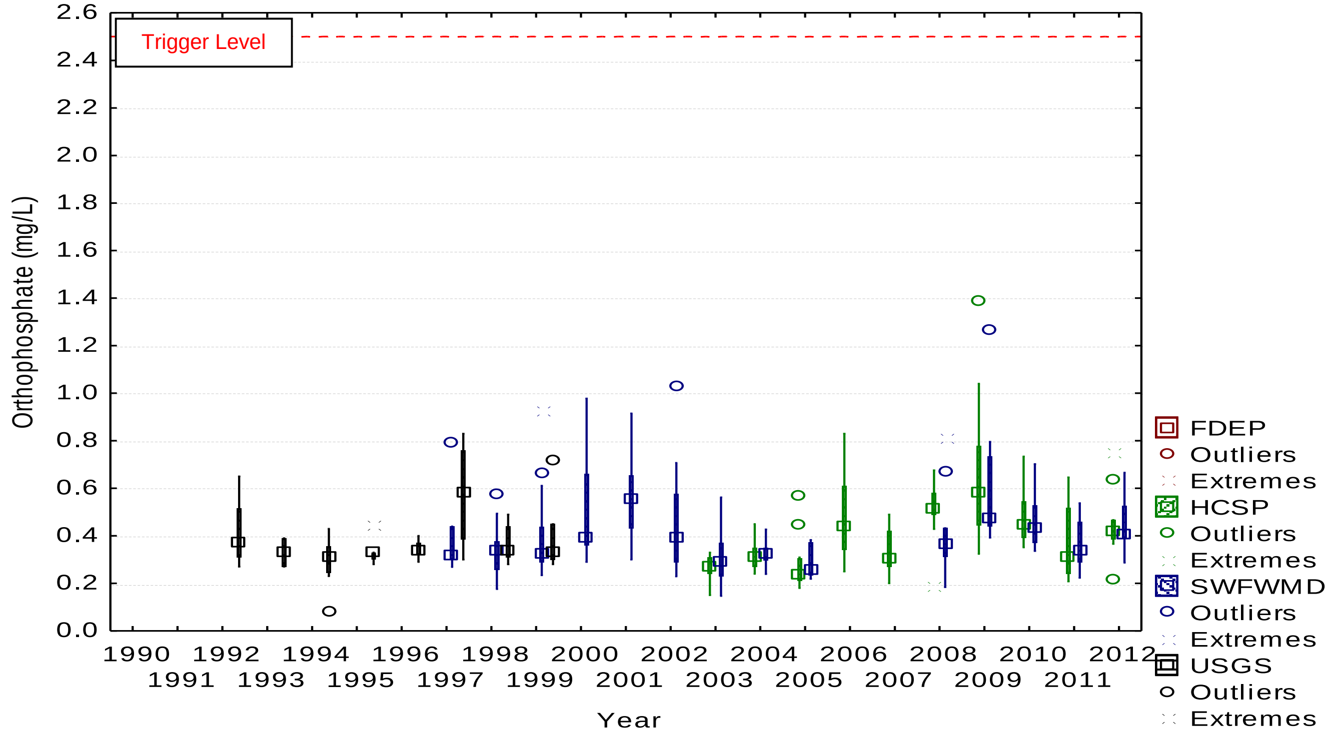

Figure 46. **HCSW-1 Values of Orthophosphate Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

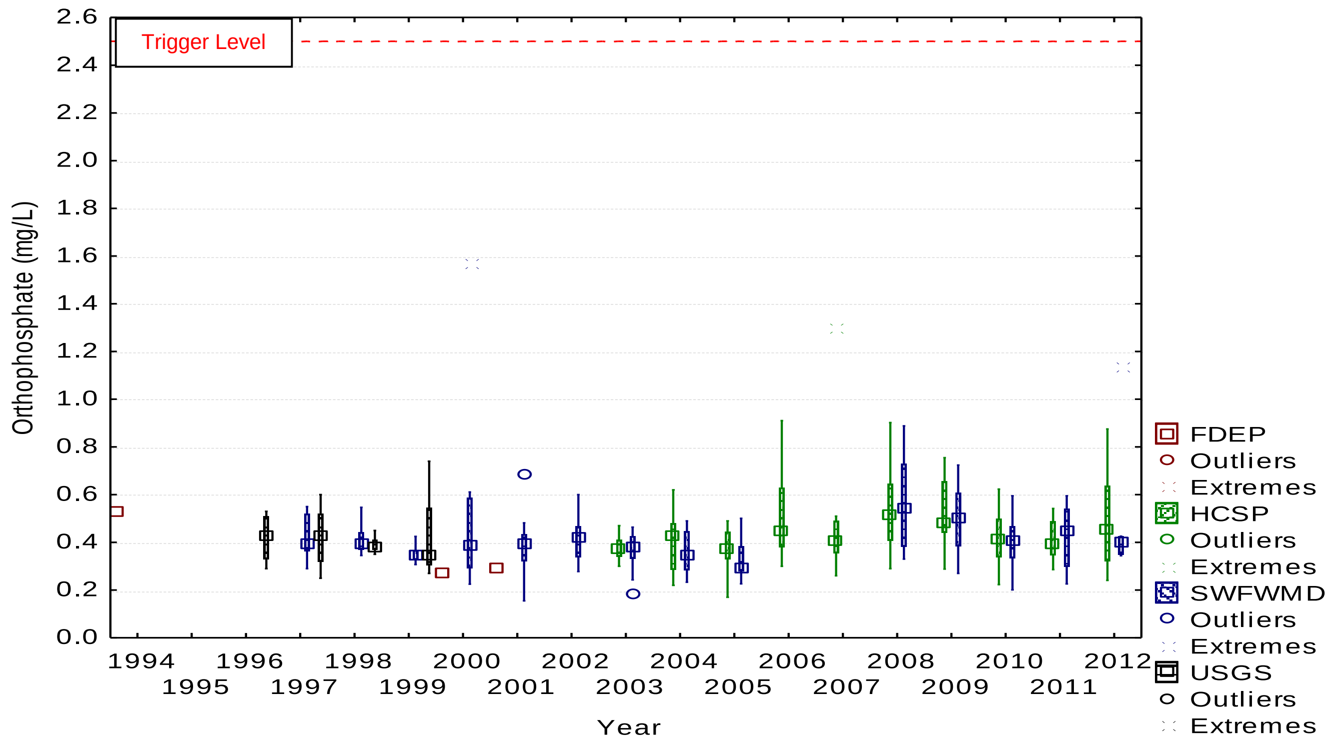

Figure 47. **HCSW-4 Values of Orthophosphate Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Chlorophyll-a

Chlorophyll *a* values were well below the trigger level of 15 mg/m^3 during all sampling events at three stations in 2012, but HCSW-2 exceeded the trigger value for chlorophyll *a* during February, April, May, and December of 2012 (Figure 48). The chlorophyll *a* concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau, $p > 0.05$, Figures 49 and 50, Table 15). Chlorophyll *a* concentrations were significantly different between stations (ANOVA, $F = 28.55$, $p < 0.001$, Table 16), with HCSW-2 significantly higher than other stations (Duncan's multiple range test, $p < 0.05$). Chlorophyll *a* was not correlated with streamflow, rainfall, and NPDES discharge at HCSW-1 or HCSW-4 (Spearman's rank correlation, $p > 0.05$, Table 17). Chlorophyll concentrations at Brushy Creek were higher than concentrations at Horse Creek stations.

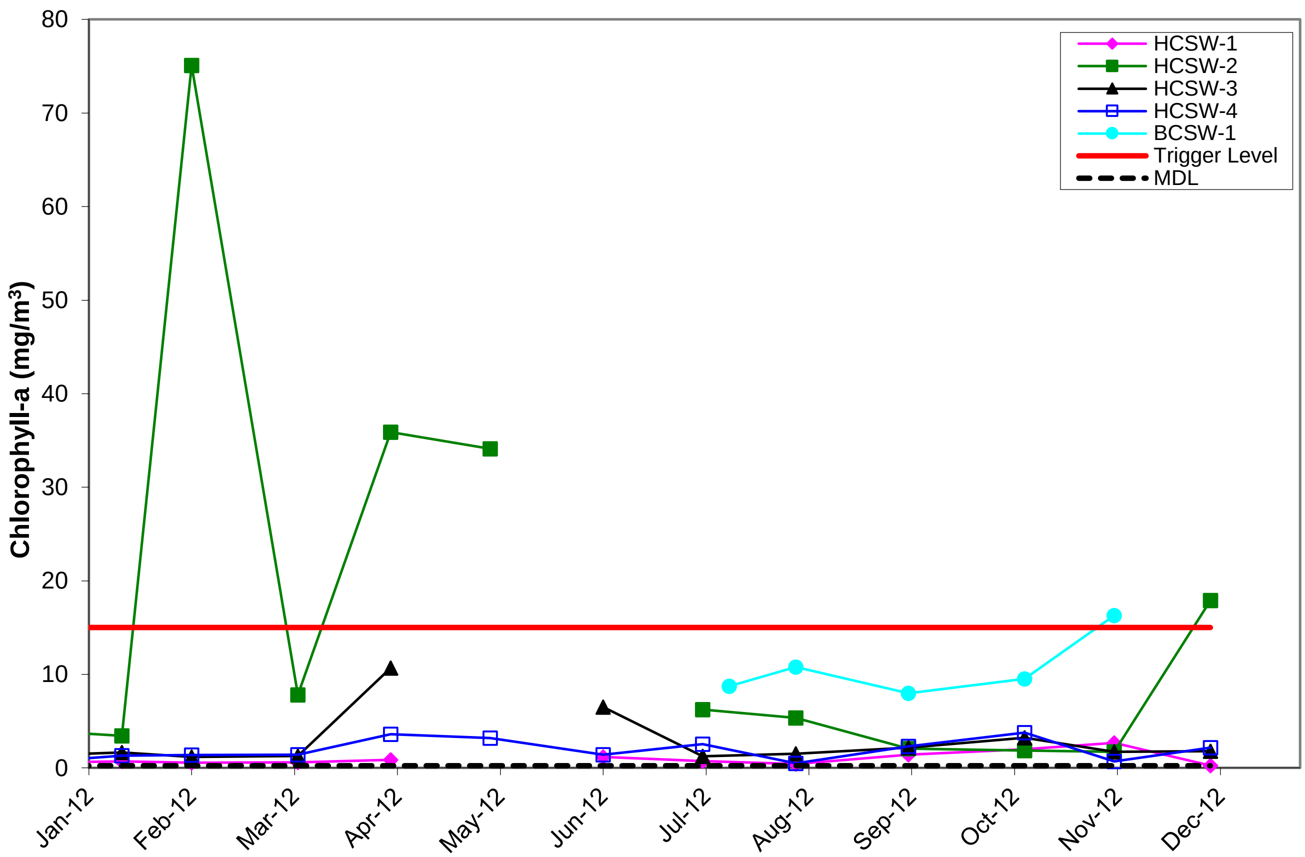

Figure 48. Chlorophyll-a Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

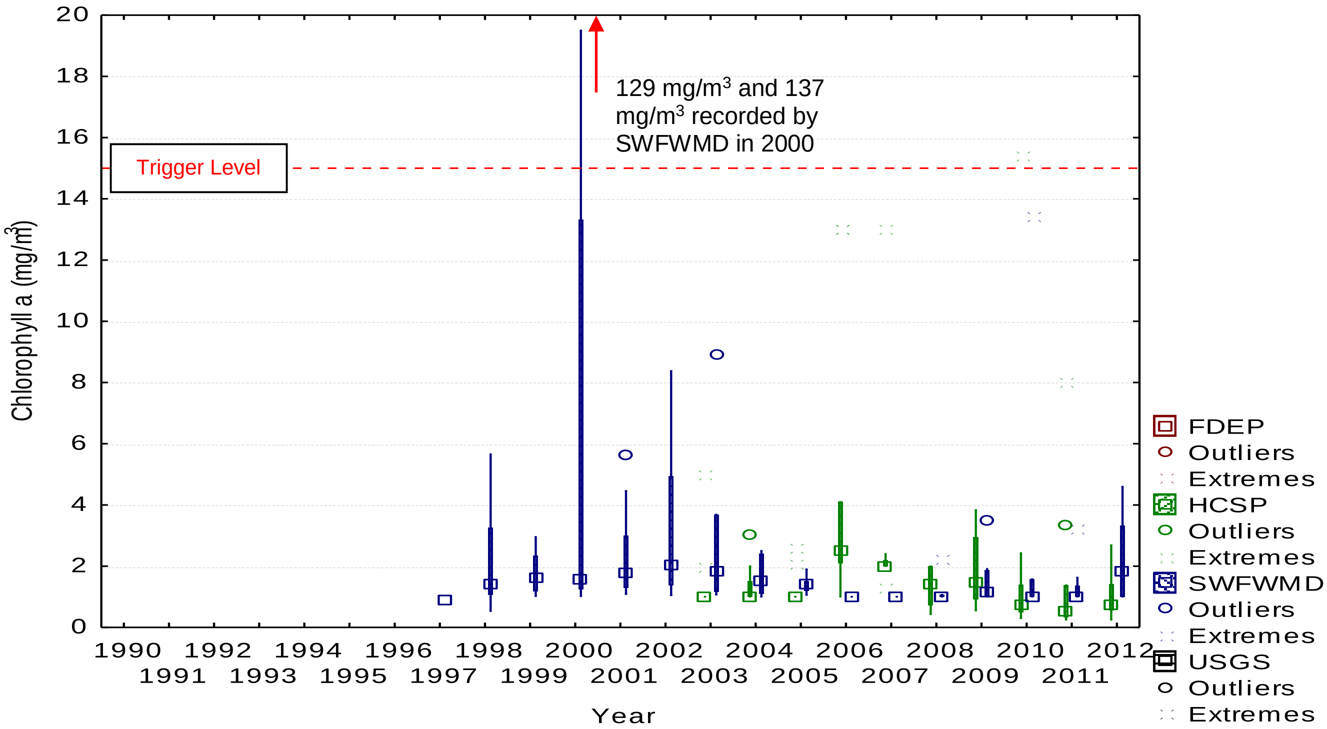

Figure 49. **HCSW-1 Values of Chlorophyll a Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

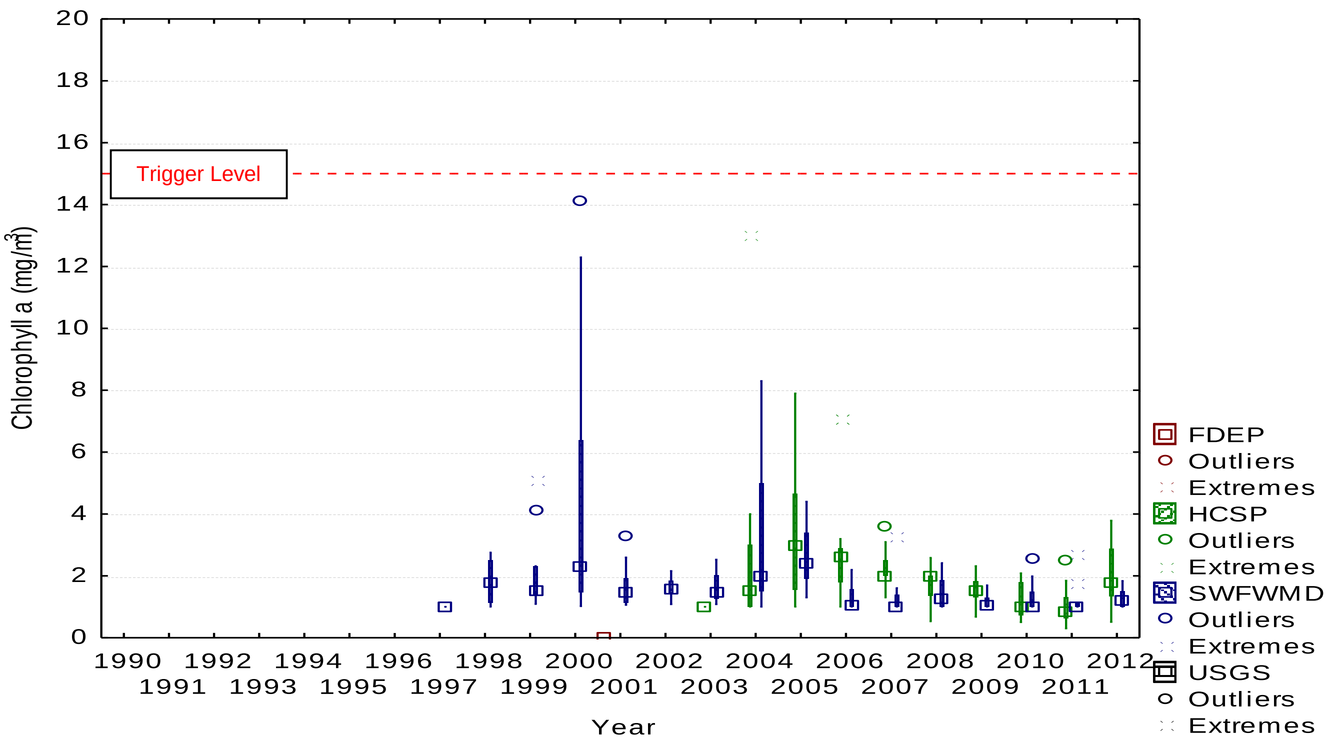

Figure 50. **HCSW-4 Values of Chlorophyll a Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

5.2.4 Dissolved Minerals, Mining Reagents and Radionuclides

Specific Conductivity

During all sampling events and stations, specific conductivity levels were well below the trigger level of $\leq$1275 µmhos/cm^2 with the exception of HCSW-4 in June 2012 (Figure 51). Levels of specific conductivity determined during each biological sampling event were consistent with those obtained during monthly water quality sampling events (Figure 51). Mean daily specific conductivity values obtained from the recorder at HCSW-1 were within the range obtained during the monthly water quality sampling events (Figure 54). The specific conductivity at HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15, Figure 53), but there was an increasing monotonic trend at HCSW-1 (Seasonal Kendall Tau with LOWESS, Tau = 0.51, $p = 0.004$, Sen slope = 10.6 µmhos/cm per year, Figure 51). This potential trend is discussed in the impact analysis in Appendix I. Changes in mining practices over time, in addition to regional climatic or other landuse effects, have contributed to a step-change in specific conductivity and other dissolved ions since 2006 at HCSW-1. However, concentrations from 2008 to 2012 have remained consistent, with the trend slope actually decreasing slightly since the analysis completed for the 2003-2010 dataset, giving no evidence of further increases over time. The biological data from HCSW-1 do not indicate any significant effects of the step-change in conductivity, but the program will continue to monitor this closely.

Specific conductivity was significantly different among stations over the 2003-2012 time period (ANOVA, $F = 37.69$, $p < 0.001$, Table 16), with the lowest overall readings at HCSW-2 followed by HCSW-1 (Duncan's multiple range test, $p < 0.05$). Specific conductivity was negatively correlated with streamflow, rainfall, and NPDES discharge at HCSW-4 (Spearman's rank correlations $-0.34 > r > -0.79$, $p < 0.05$, Table 17), but positively correlated with NPDES discharge ($r_s = 0.20$) and negatively correlated with rainfall ($r_s = -0.23$) at HCSW-1. Concentrations at Brushy Creek were lower than Horse Creek stations throughout 2012.

Higher conductivity at downstream stations over the course of the HCSP was probably the cumulative result of contributions of groundwater that either seeped into Horse Creek directly or ran off of agricultural lands as a result of irrigation water pumped from the aquifer. This pattern has been present for many years and is more apparent in the review of the long-term data in a separate report (Durbin and Raymond 2006). It is possible that some of the conductivity differential may simply be the result of changes in geology of the watershed from high elevations in the upper part of the basin to low elevations in the lower part of the basin near the Peace River. Groundwater, which generally contains more concentrated dissolved ions than surface water, is closer to the surface in the lower Horse Creek Basin, making seepage into the stream more likely. A review of land use types in the basin also shows more land under agriculture use in the lower basin than the upper basin suggesting a higher potential for higher ion levels in the lower basin due to agriculture irrigation runoff. In recent years, changes in mining practices have raised the conductivity at the HCSW-1 station; although the new levels appear to be stable and not biologically harmful (see Appendix I).

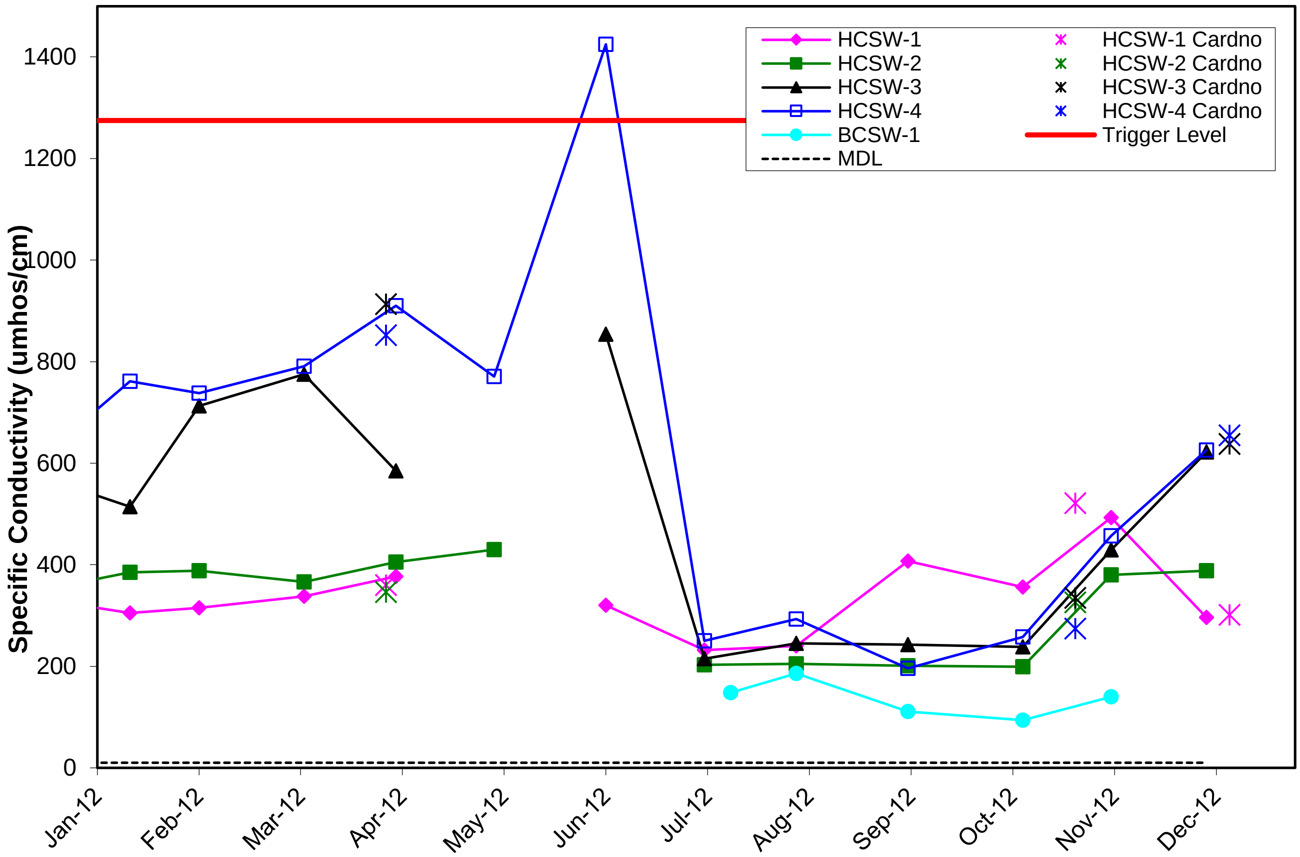

Figure 51. **Levels of Specific Conductivity Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events in 2012.**

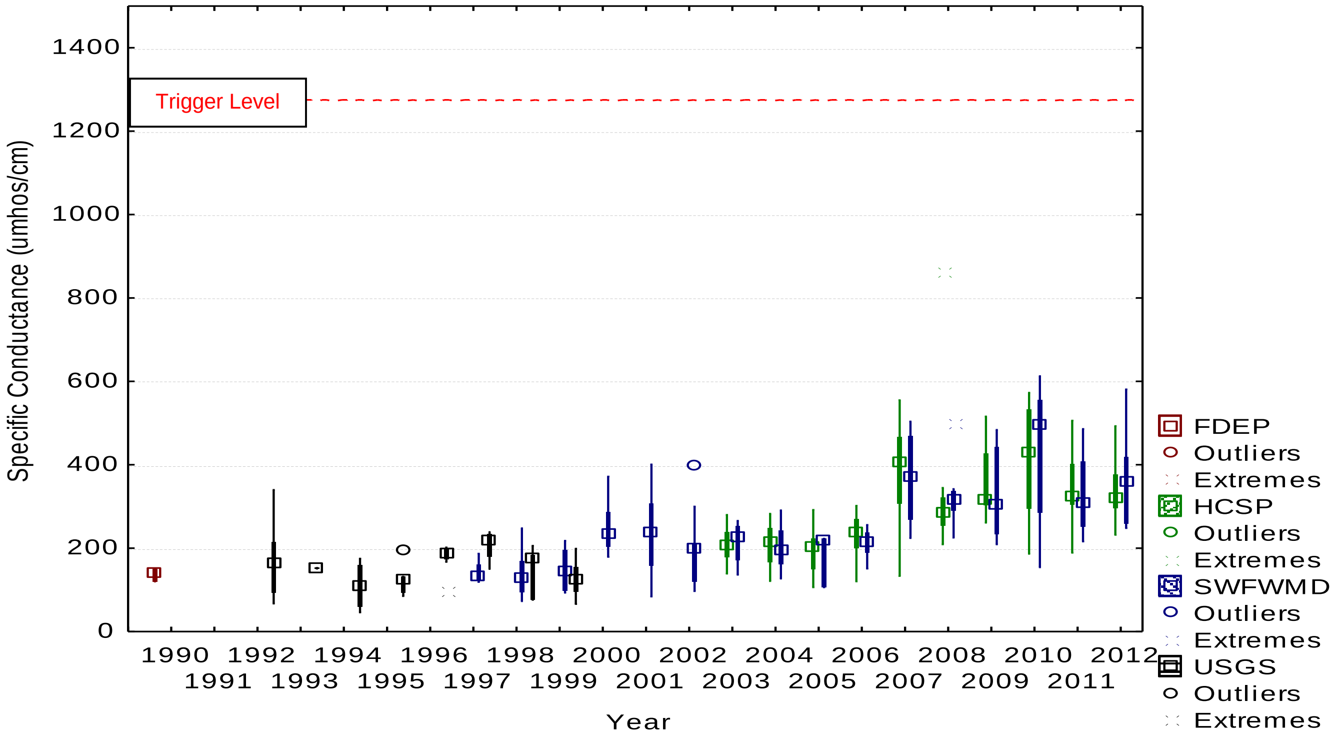

Figure 52. **HCSW-1 Values of Specific Conductance Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

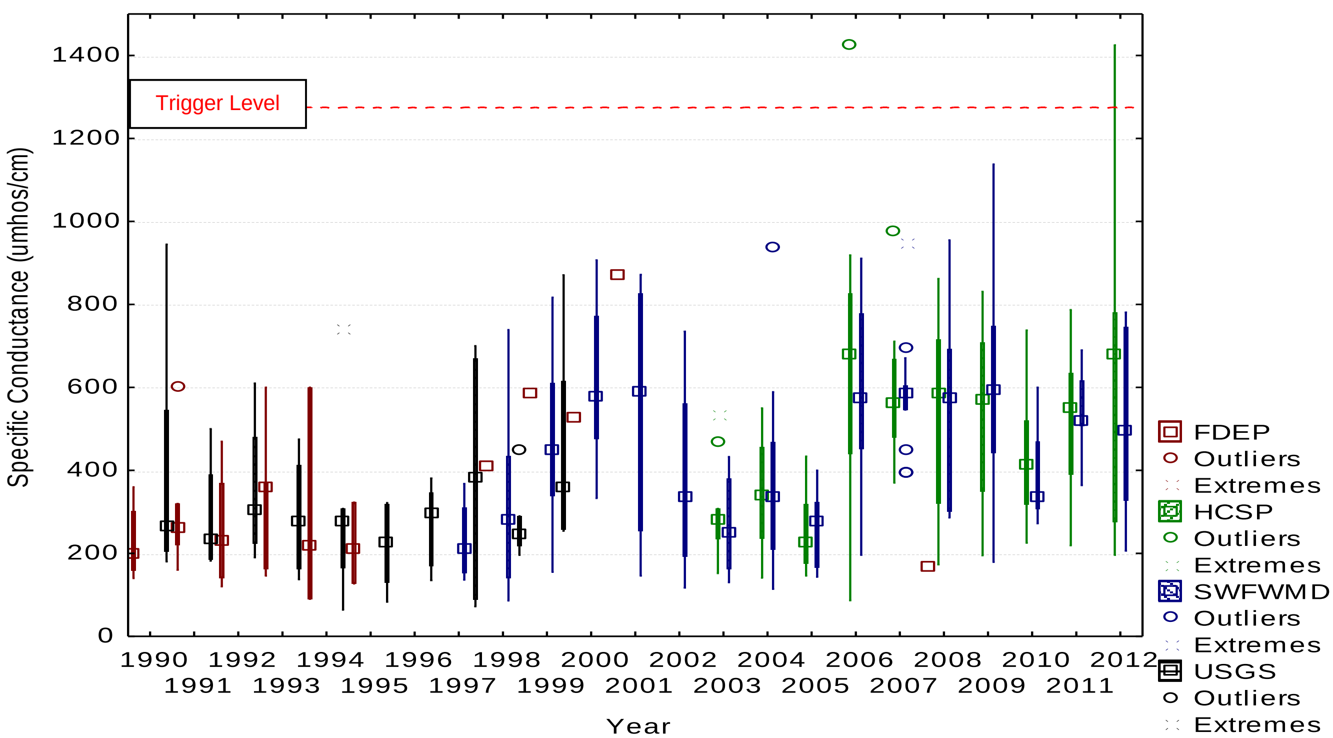

Figure 53. **HCSW-4 Values of Specific Conductance Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

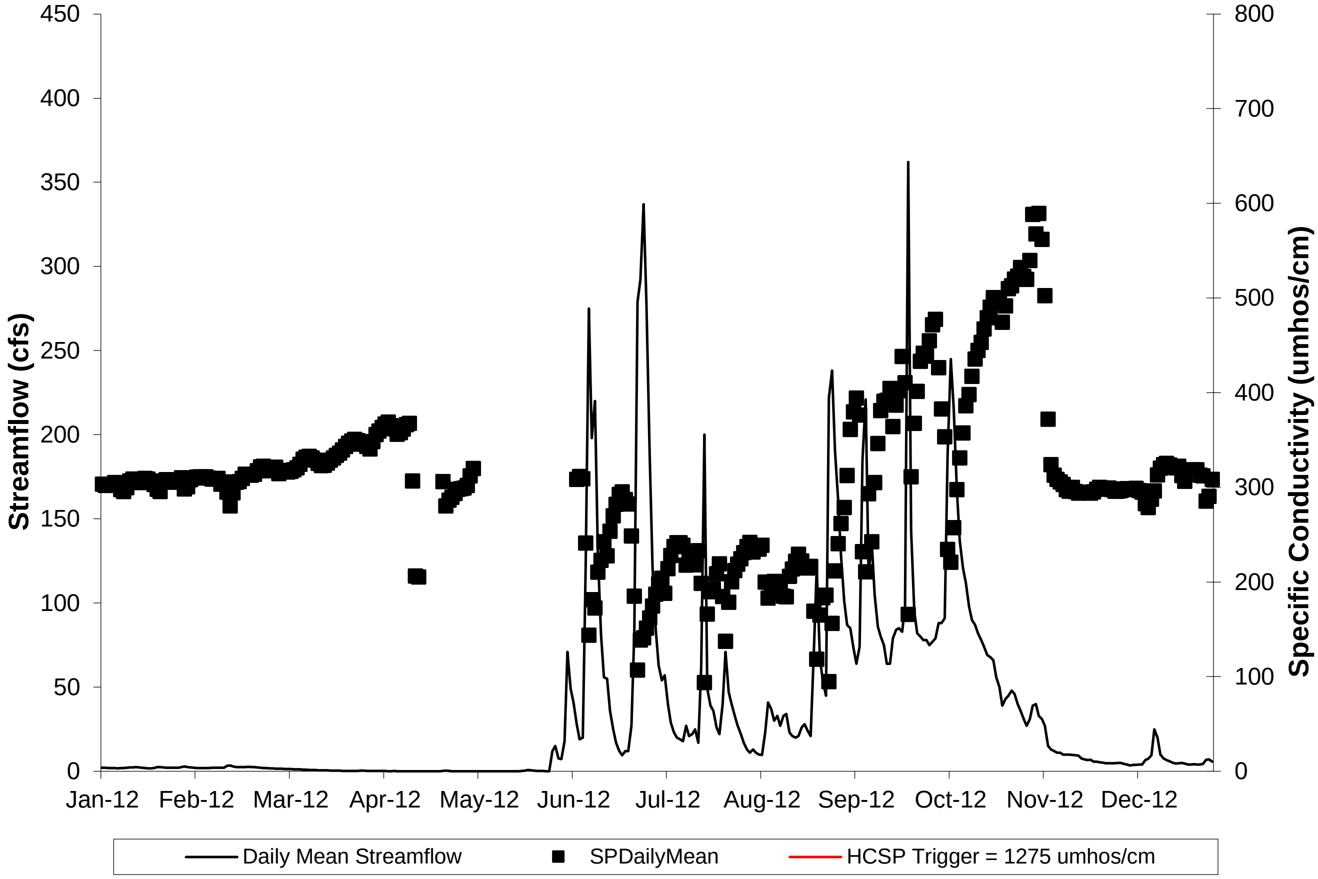

Figure 54. Relationship Between Daily Mean Specific Conductivity (Obtained From the Continuous Recorder at HCSW-1) and Daily Mean Streamflow for 2012. Min. Detection Limit = 10 µmhos/cm.

Dissolved Calcium

Calcium levels were lower than the trigger value of 100 mg/l at HCSW-1 and HCSW-2 during all events in 2012; HCSW-3 exceeded the trigger value in April and HCSW-4 exceeded the trigger value from April to June 2012 (Figure 55). The calcium concentrations at HCSW-4 measured by the HCSP show no monotonic trend from 2003 to 2012 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Figure 57, Table 15); however, HCSW-1 exhibited an increasing trend since 2003 (Seasonal Kendall Tau with LOWESS, Tau = 0.51, $p = 0.004$, Sen slope = 1.05 mg/L per year, Figure 56, Table 15). The trend for HCSW-1 is small compared to historic HCSP differences between primary and field duplicate samples (≤ 8.0 mg/L). The trend for calcium and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful (Appendix I).

Concentrations of calcium were significantly different between stations (ANOVA, $F = 61.93$, $p < 0.001$), with significantly higher levels at HCSW-4 and significantly lower levels at HCSW-2 (Duncan's post hoc test, $p < 0.05$, Figure 55, Table 16). As with specific conductivity, calcium concentrations may be higher downstream because of increased groundwater seepage or irrigation runoff, especially during the dry years of 2006 to 2007. Calcium was negatively correlated with streamflow, rainfall, and NPDES discharge at HCSW-4 (Spearman's rank correlations $-0.32 > r > -0.81$, $p < 0.05$, Table 17), but positively correlated with NPDES discharge ($r_s = 0.21$) and negatively correlated with rainfall ($r_s = -0.27$) at HCSW-1. Brushy Creek had lower calcium concentrations than the Horse Creek stations.

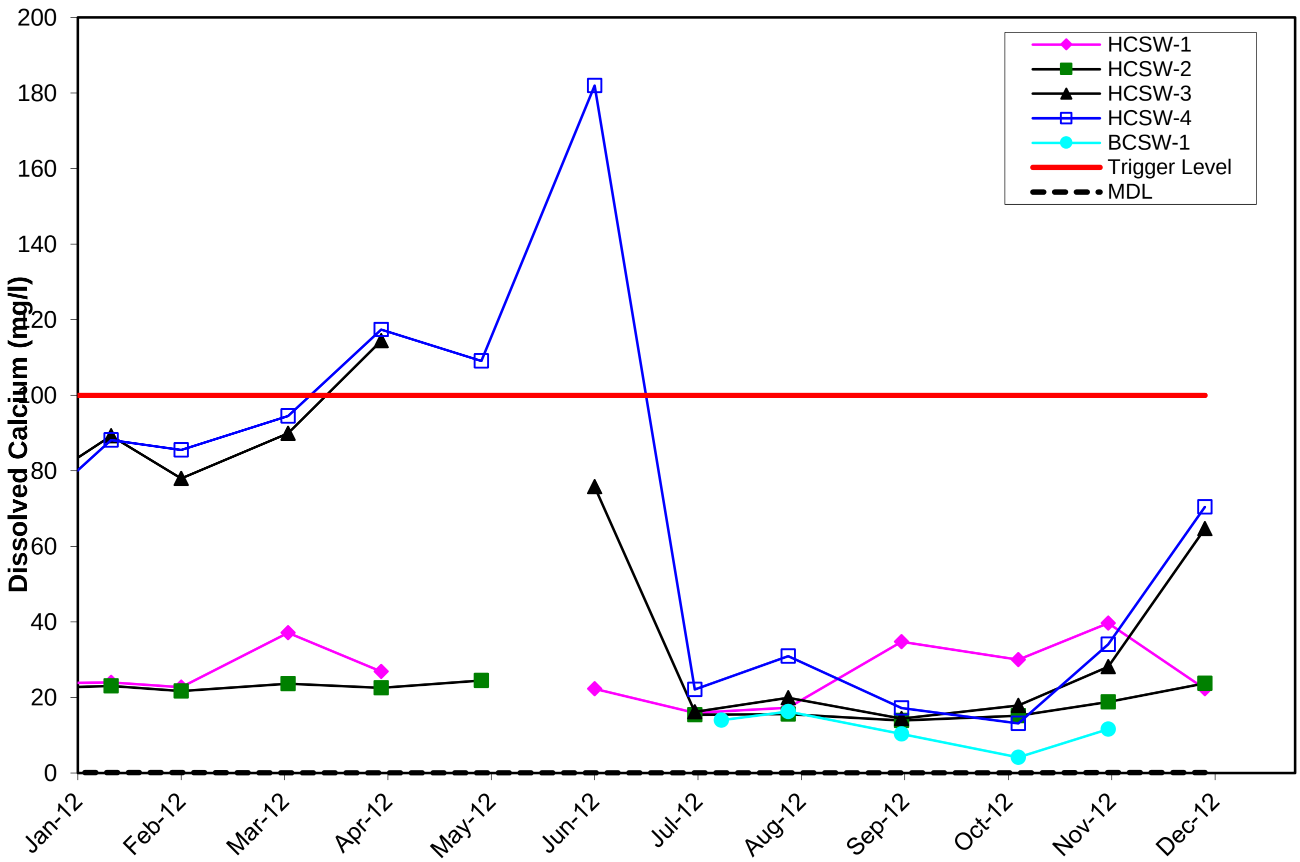

Figure 55. **Dissolved Calcium Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012. Minimum Detection Limit = 0.1 mg/l.**

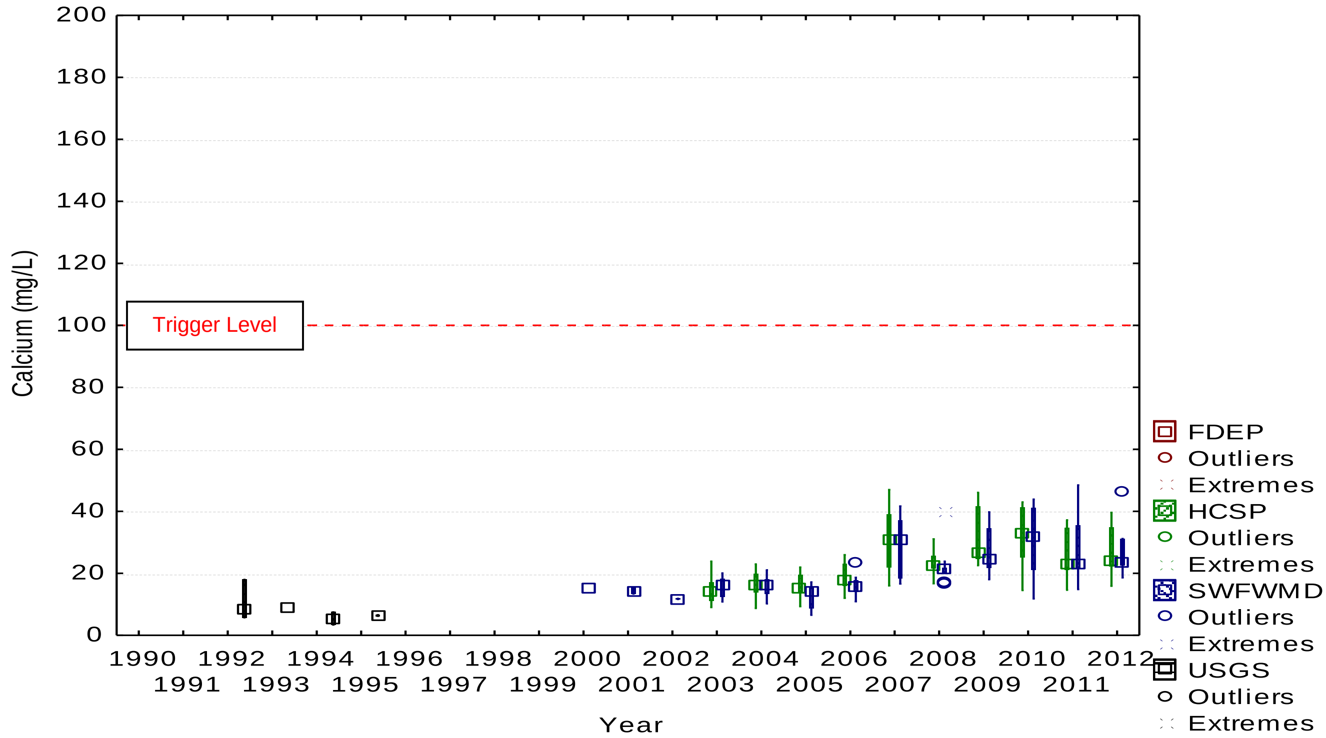

Figure 56. HCSW-1 Values of Calcium Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

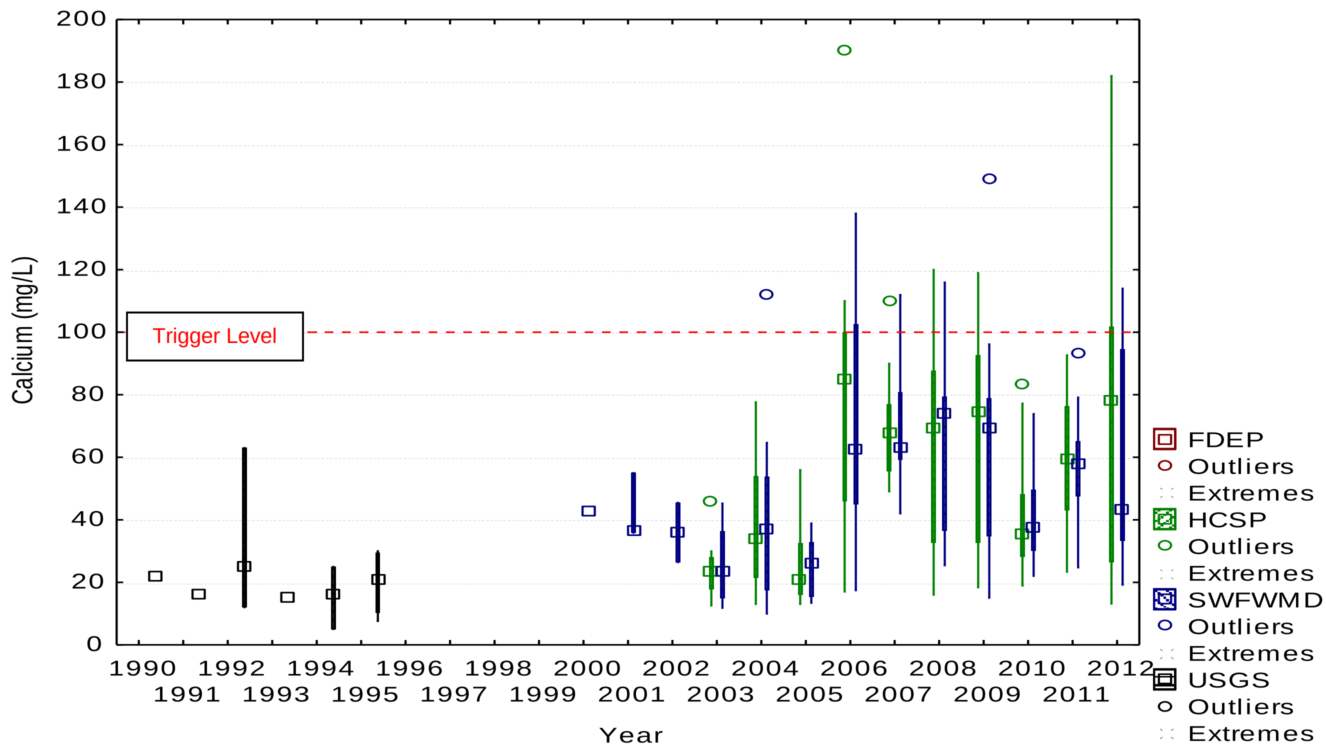

Figure 57. HCSW-4 Values of Calcium Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

Dissolved Iron

Levels of dissolved iron at all stations were below the trigger level of 1 mg/l during all sampling events in 2012 (Figure 58). Dissolved iron concentrations at HCSW-4 exceeded the trigger value of 0.3 mg/l established for that sampling station from July through October. HCSW-4 has a different trigger level for iron because of its location upstream of a segment of the Peace River that is designated as Class I waters, which carries a lower standard value for iron (0.3 mg/l) than Class III waters (1.0 mg/l). The iron concentrations at HCSW-1 and HCSW-4 measured by the HCSP were not compared to data collected by other sources because the historical data is limited for this water quality parameter.

There were decreasing monotonic trends for dissolved iron since 2003 at both HCSW-1 (Seasonal Kendall Tau with LOWESS, Tau = -0.35, p = 0.02, Sen Slope = -0.02 mg/L per year) and HCSW-4 (Seasonal Kendall Tau with LOWESS, Tau = -0.35, p = 0.02, Sen Slope = -0.01 mg/L per year, Table 15). Because the direction of this potential trend is opposite that of the HCSP trigger level, it is not of concern (Appendix I). The program will continue to monitor this condition over time.

Dissolved iron concentrations were not significantly different among stations (ANOVA, F = 0.54, p = 0.66, Table 16). Iron was positively correlated with streamflow, rainfall, and NPDES discharge at HCSW-1 and HCSW-4 (Spearman's rank correlations 0.35 < r < 0.79, p < 0.05, Table 17). Brushy Creek had slightly higher iron concentrations than the Horse Creek stations.

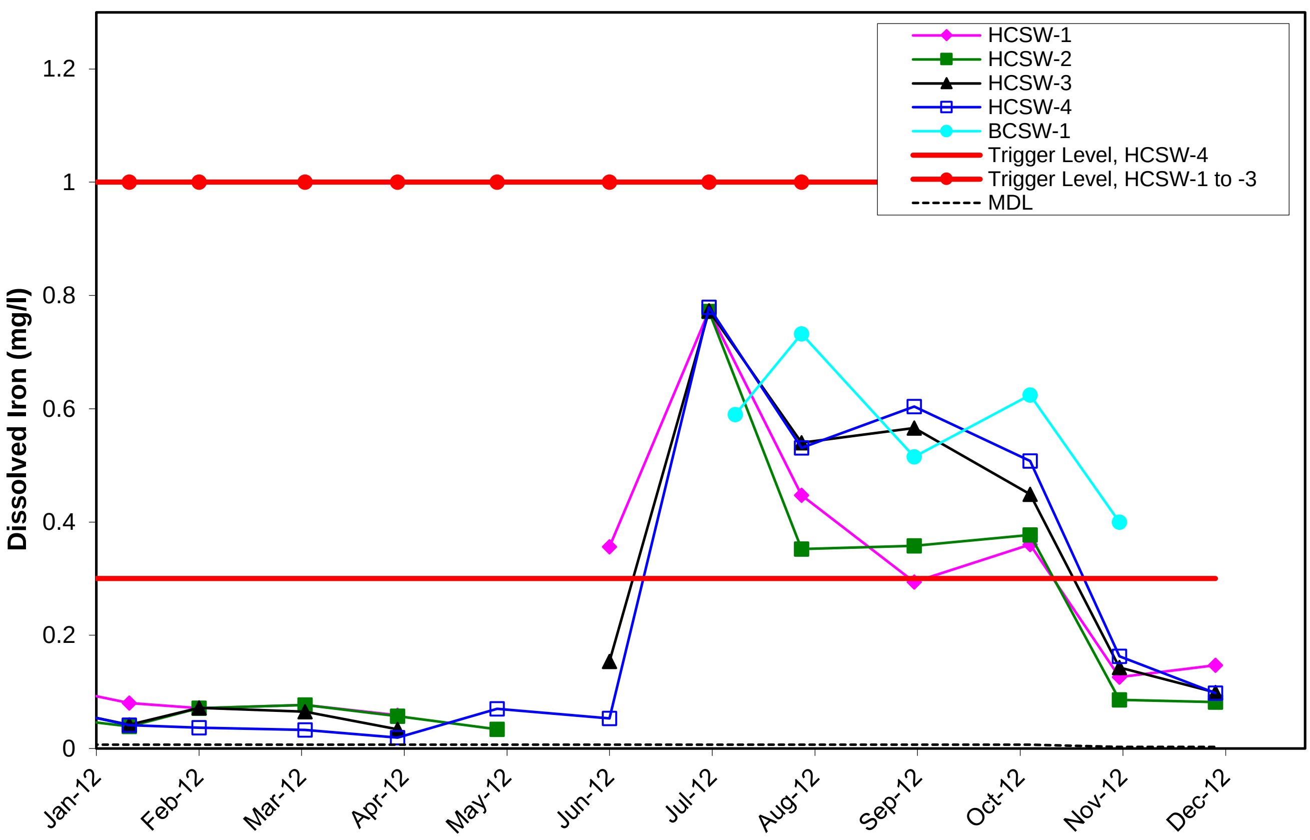

Figure 58. Dissolved Iron Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

Total Alkalinity

Levels of total alkalinity were well below the trigger value of 100 mg/l during 2012 except for the May sample at HCSW-4 and November sample at HCSW-1 (Figure 59). The alkalinity levels at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources, with slightly elevated alkalinity measurements at HCSW-4 than in the past (Figures 60 and 61). There was an increasing monotonic trend present from 2003-2012 at both HCSW-1 (Seasonal Kendall Tau with LOWESS, Tau = 0.42, $p = 0.004$, Sen slope = 2.96 mg/L per year) and HCSW-4 (Seasonal Kendall Tau with LOWESS, Tau = 0.53, $p = 0.0003$, Sen slope = 1.66 mg/L per year, Table 15). The estimated slope for HCSW-1 and HCSW-4 is small compared to the differences between primary and field duplicate samples (≤17 mg/L). The trend for alkalinity and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful (Appendix I).

Total alkalinity was significantly different among stations (ANOVA, $F = 35.86$, $p < 0.001$, Table 16), with highest levels at HCSW-1 followed by HCSW-4 (Duncan's multiple range test, Figure 59). Alkalinity was negatively correlated with streamflow, rainfall, and NPDES discharge at HCSW-4 (Spearman's rank correlation $-0.40 > r > -0.76$, $p < 0.05$, Table 17), which is consistent with the concept that higher flows from rainfall would reflect the lower alkalinity of rainwater, compared with dry season inputs of groundwater. This condition suggest that groundwater seepage and agriculture irrigation runoff may also contribute to higher levels of alkalinity at HCSW-4. However, NPDES discharge was positively correlated with alkalinity at HCSW-1 ($r_s = 0.23$) and rainfall was negatively correlated ($r_s = -0.20$). High levels of alkalinity at HCSW-1 may be partly attributed to the exposed rock in the stream banks that is unique to that station and recent mining changes to mining practices described in Appendix I. Brushy Creek had lower alkalinity concentrations than the Horse Creek stations.

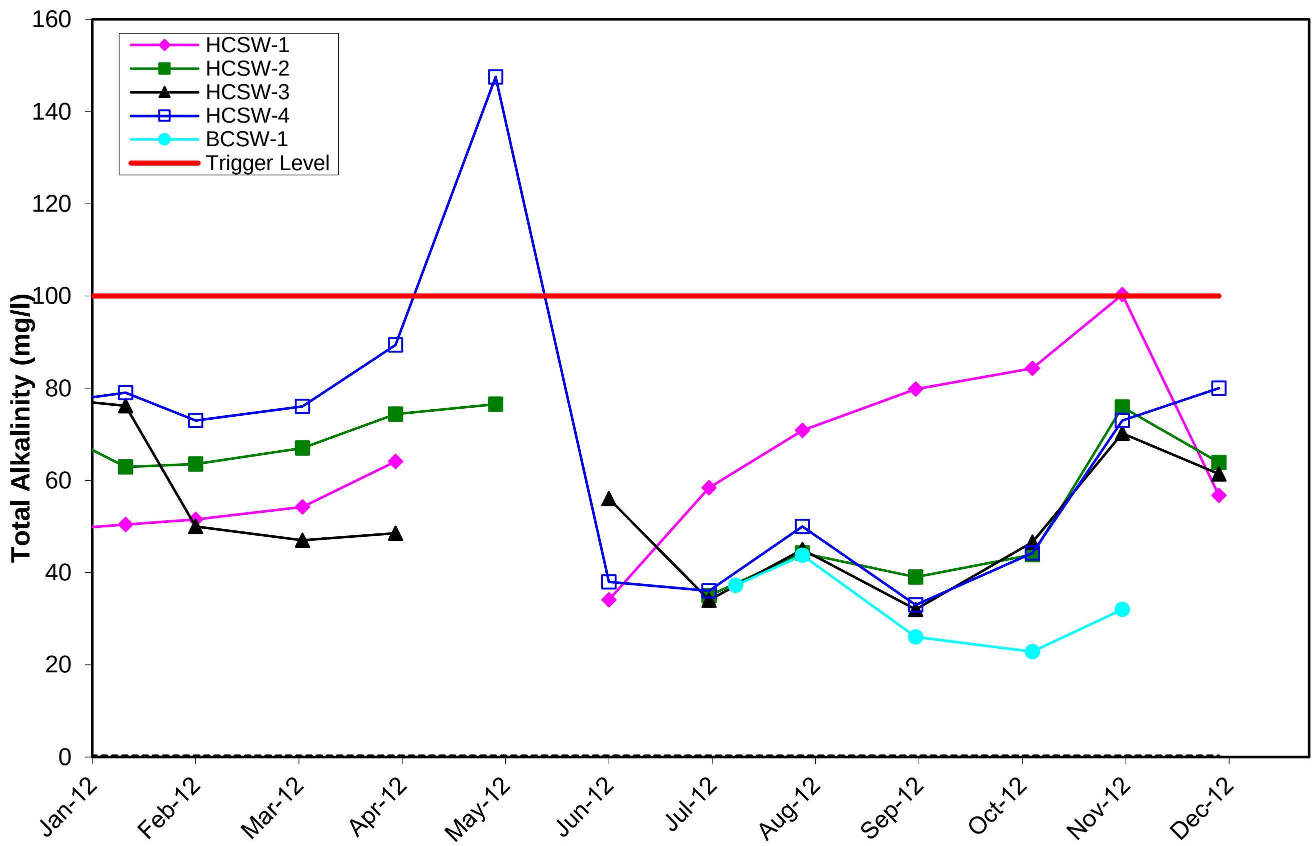

Figure 59. Levels of Total Alkalinity Obtained During Monthly HCSP Water Quality Sampling in 2012.

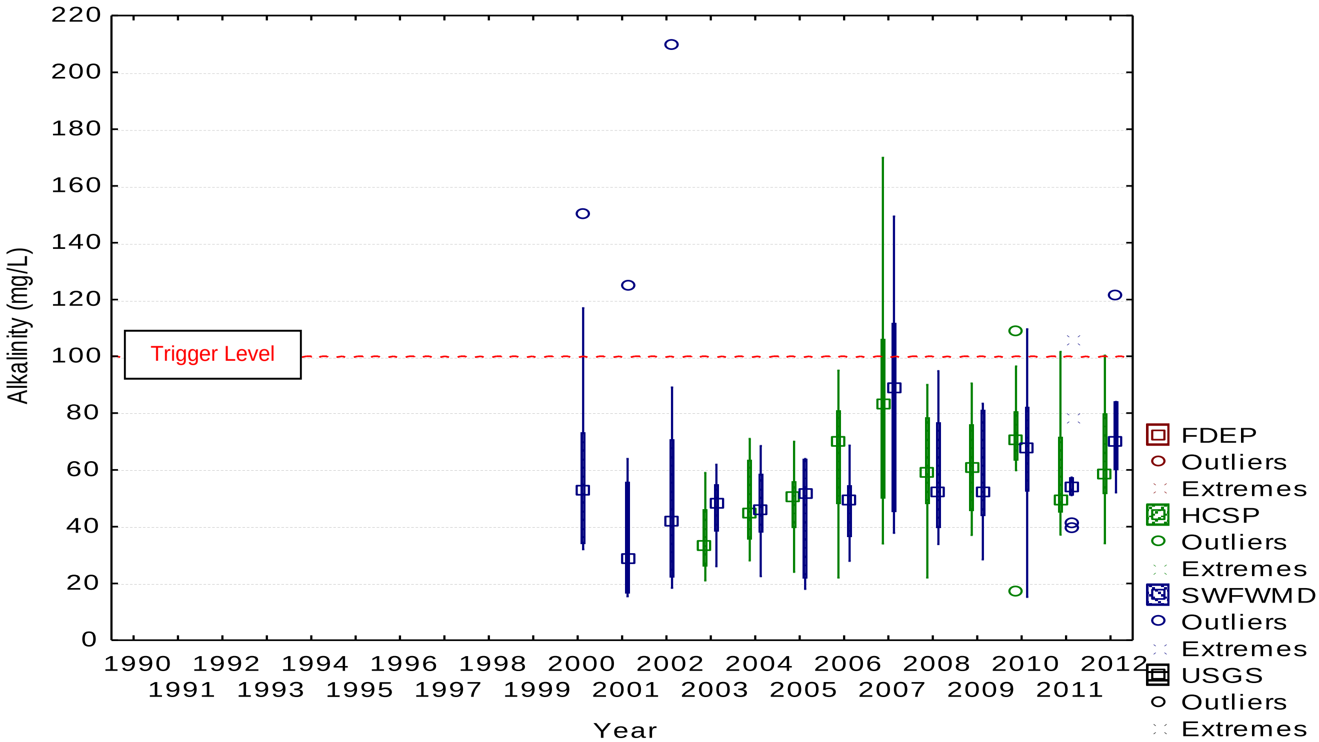

Figure 60. **HCSW-1 Values of Alkalinity Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

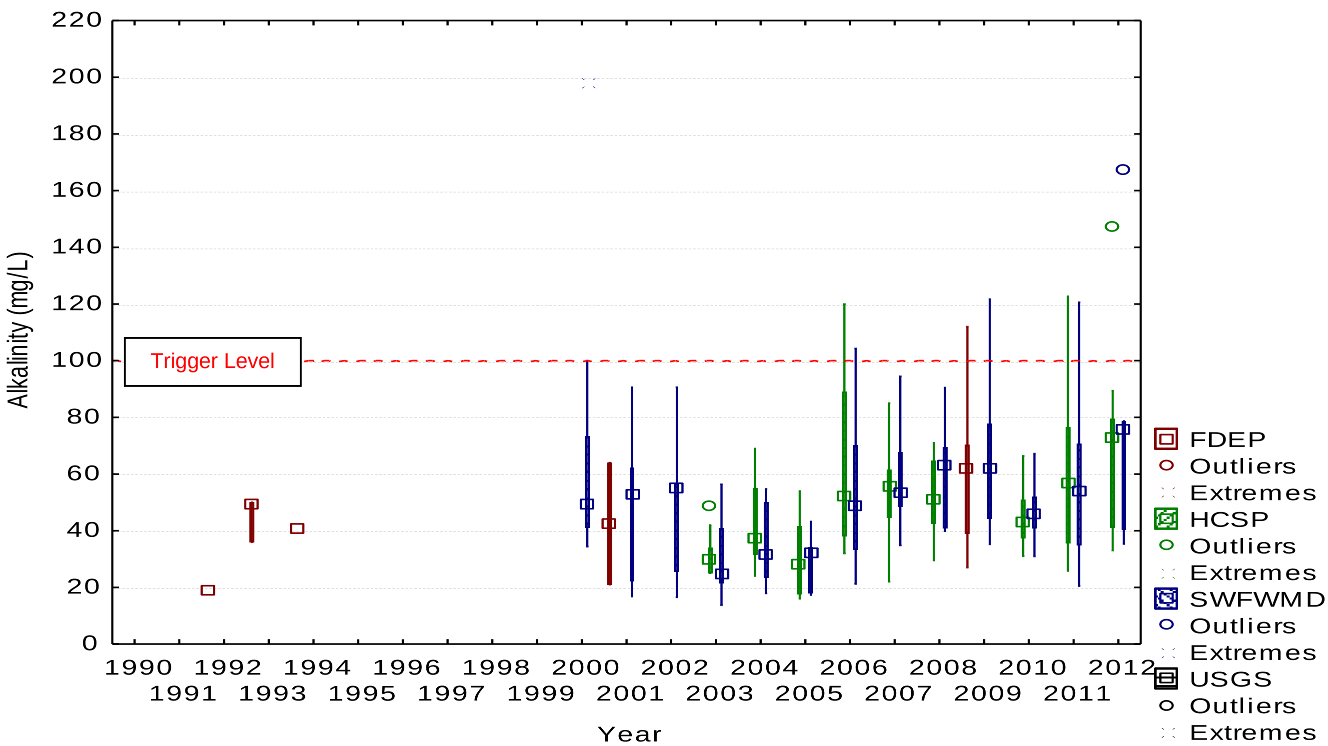

Figure 61. **HCSW-4 Values of Alkalinity Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Chloride

Levels of chloride were below 45 mg/l during 2003 - 2012, considerably lower than the trigger level of 250 mg/l (Figure 62). The chloride concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Figures 63 and 64, Table 15). Chloride concentrations were significantly different among stations during all sampling events (ANOVA, $F = 23.55$, $p < 0.001$, Table 16), with a pattern of increasing concentration downstream suggesting again the possible influence from groundwater seepage and agriculture irrigation runoff (Figure 62). Chloride was negatively correlated with streamflow, rainfall, and NPDES discharge at HCSW-1 and HCSW-4 (Spearman's rank correlations $-0.33 > r > -0.81$, $p < 0.05$, Table 17). Brushy Creek had similar concentrations to the Horse Creek stations.

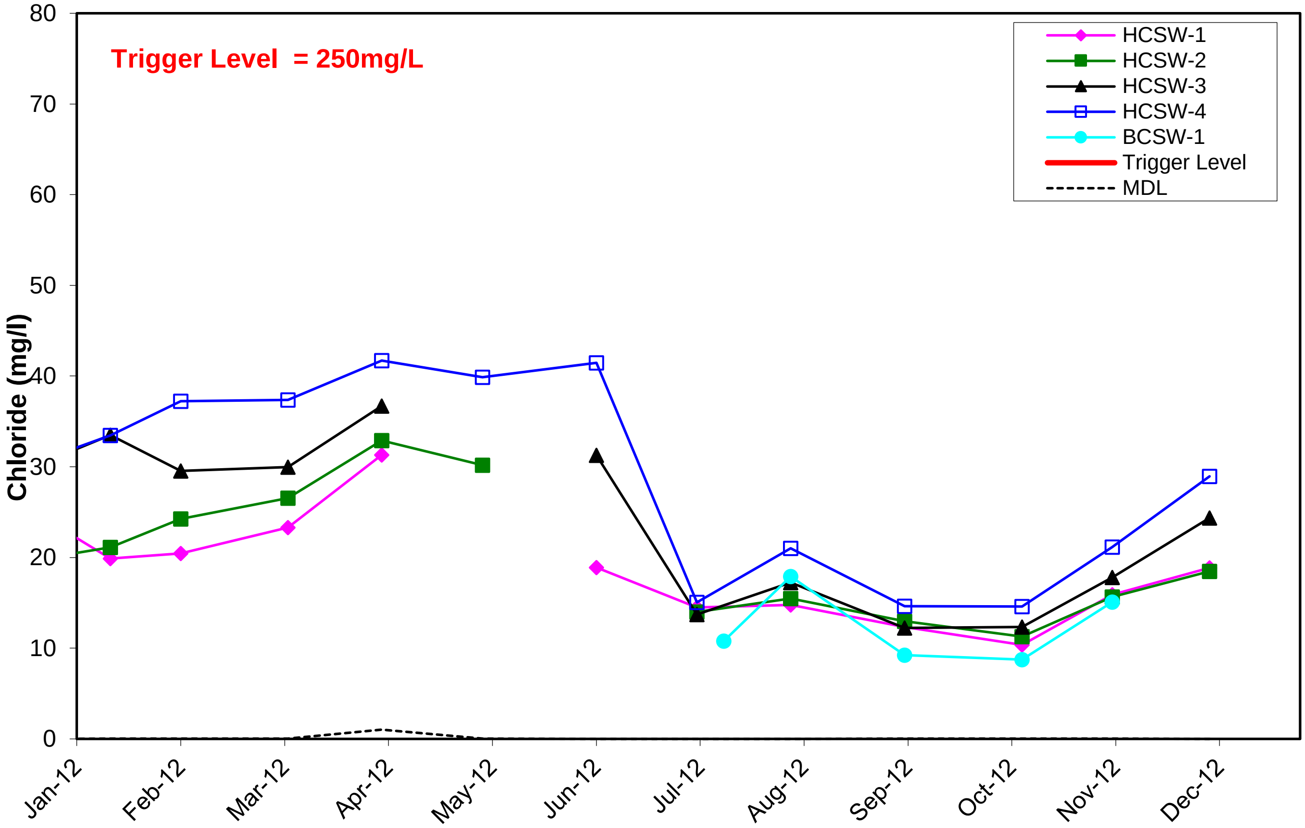

Figure 62. **Chloride Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012. (HCSP trigger value for Chloride is 250 mg/L.)**

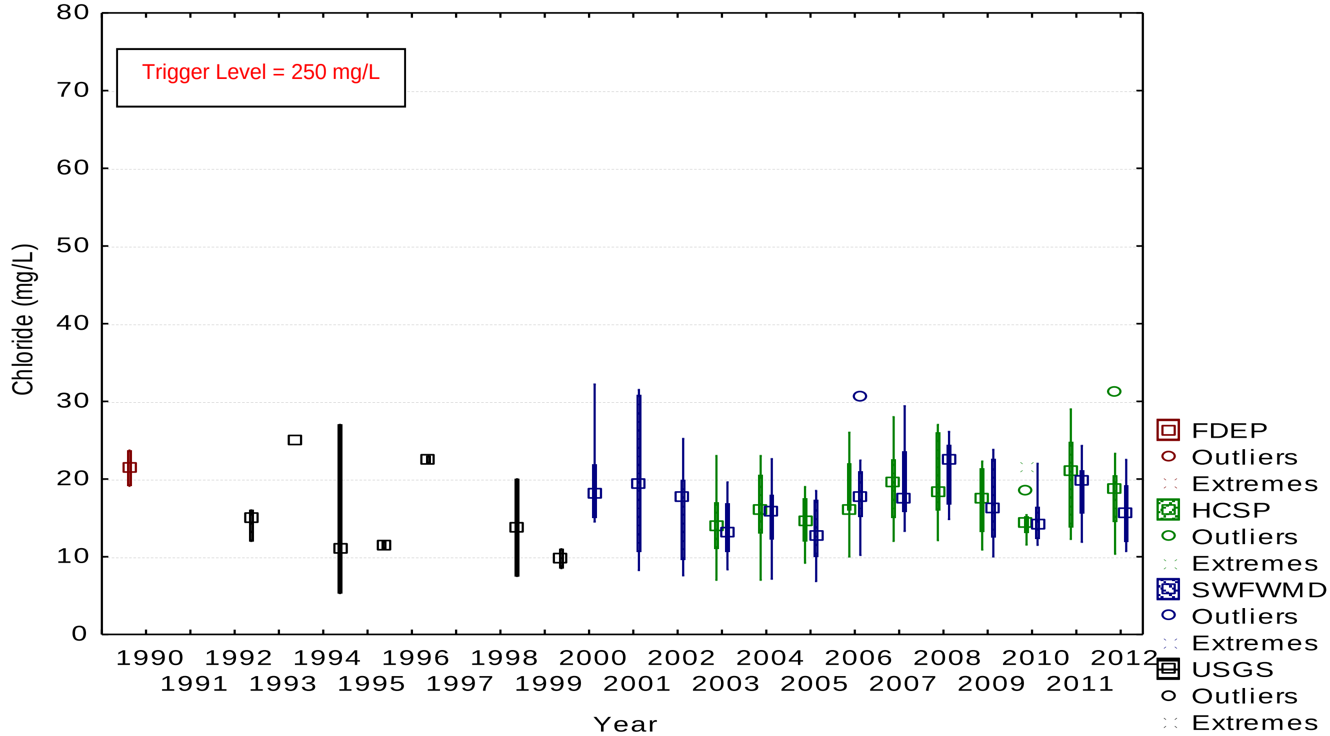

Figure 63. **HCSW-1 Values of Chloride Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

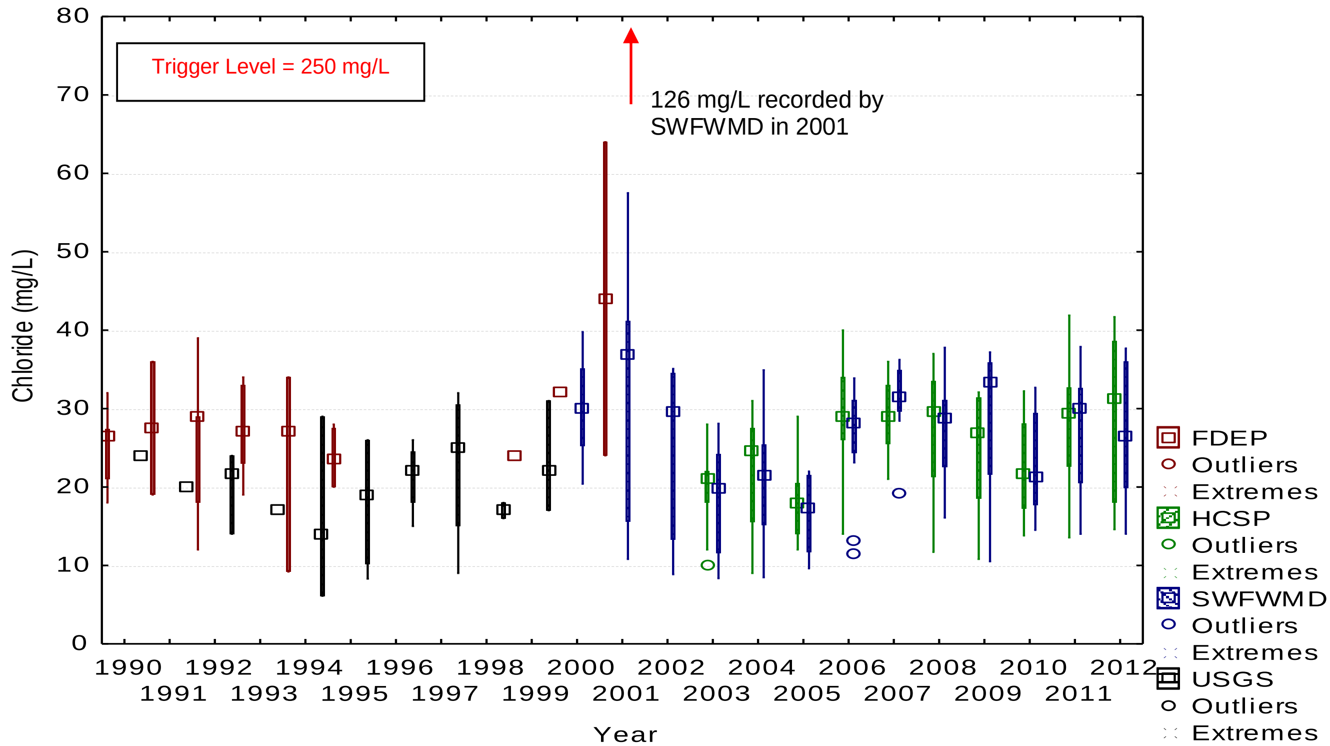

Figure 64. **HCSW-4 Values of Chloride Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Fluoride

Concentrations of fluoride were well below the trigger levels of 4.0 mg/L established for HCSW-1, HCSW-2, and HCSW-3, as well as the 1.5 mg/L trigger level for HCSW-4 (Figure 65). Brushy Creek had lower concentrations than the Horse Creek stations. After dramatic changes with the MDL for fluoride in 2007 during a drought, the MDLs have now been minimized and did not change from April 2008 through 2012. The fluoride concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources (Figures 66 and 67). The fluoride concentrations at HCSW-1 and HCSW-4 measured by the HCSP could not be analyzed for monotonic trends using seasonal data or correlations with water quantity because of changes in MDL's over the course of the HCSP. Based on an alternative trend analysis performed on annual medians from data collected by SWFWMD from 2003-2012, no increasing or decreasing trend was observed at HCSW-1 or HCSW-4 (Annual Kendall Tau with LOWESS, $p > 0.05$, Table 15). The slope of the trend previously observed for fluoride at HCSW-1 was no longer present in 2012 when running an Annual Kendall Tau analysis.

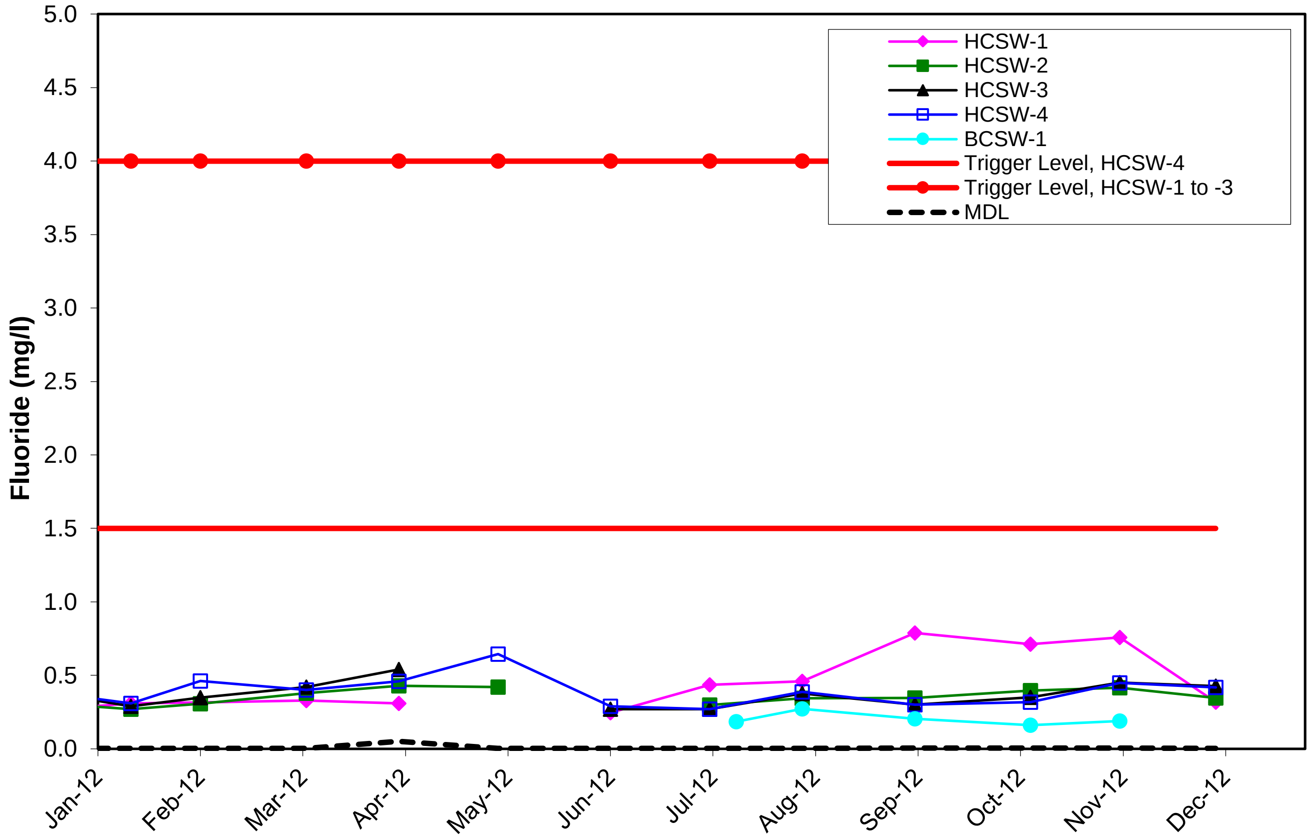

Figure 65. Fluoride Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

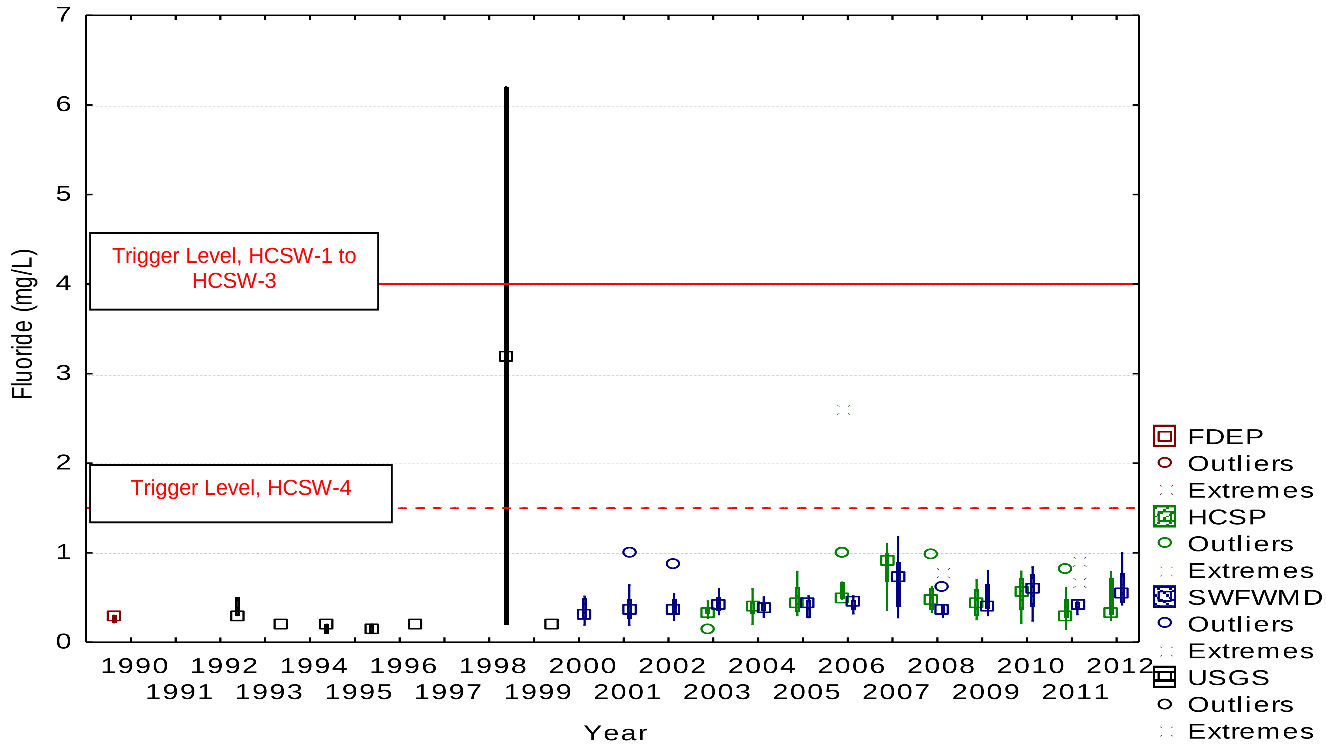

Figure 66. **HCSW-1 Values of Fluoride Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

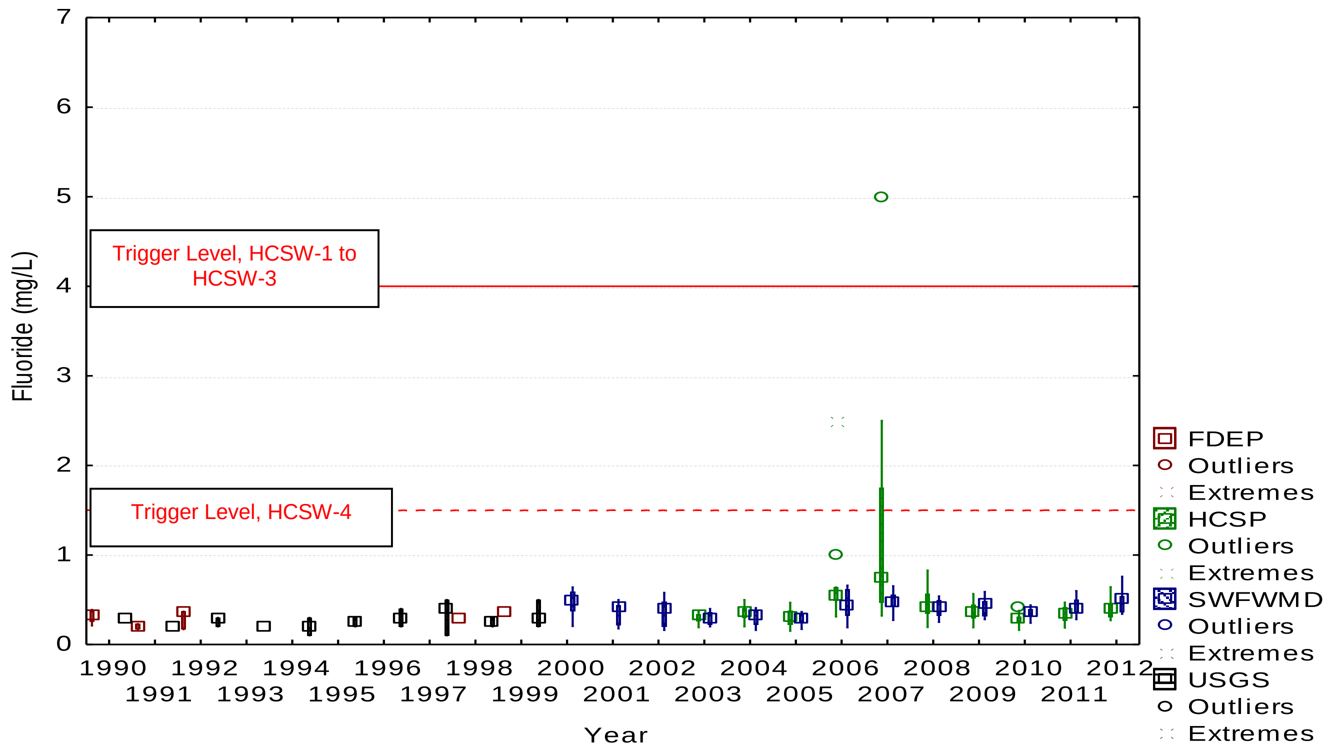

Figure 67. **HCSW-4 Values of Fluoride Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Sulfate

Sulfate concentrations were below the trigger level of 250 mg/l at HCSW-1 and HCSW-2 during all sampling events in 2012; the trigger level was exceeded at HCSW-3 during February-April and June and at HCSW-4 during March, April and June (Figure 68). The sulfate concentrations at HCSW-1 and HCSW-4 measured by the HCSP are consistent with other water quality data sources, and HCSW-4 exhibited no monotonic trends since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$, Table 15, Figures 69 and 70). There was a slightly increasing trend observed at HCSW-1 since 2003 (Seasonal Kendall Tau with LOWESS, Tau = 0.32, $p = 0.03$, Sen slope = 2.27 mg/L per year, Table 15). The trend for sulfate and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful (Appendix I).

In 2003 - 2012, levels of sulfate were significantly different among stations (ANOVA, $F = 47.25$, $p < 0.001$, Table 16), with lowest levels at HCSW-1 and HCSW-2 and highest at HCSW-4 (Duncan's multiple range test, $p < 0.05$). As with specific conductivity and calcium, sulfate concentrations may be higher downstream because of increased groundwater seepage or irrigation runoff, especially during the dry years of 2006 to 2007. Sulfate was negatively correlated with streamflow, rainfall, and NPDES discharge at HCSW-4 (Spearman's rank correlation $-0.31 > r > -0.75$ $p < 0.05$, Table 17), but was positively correlated with NPDES discharge ($r_s = 0.29$) at HCSW-1. Brushy Creek concentrations were lower than at Horse Creek stations.

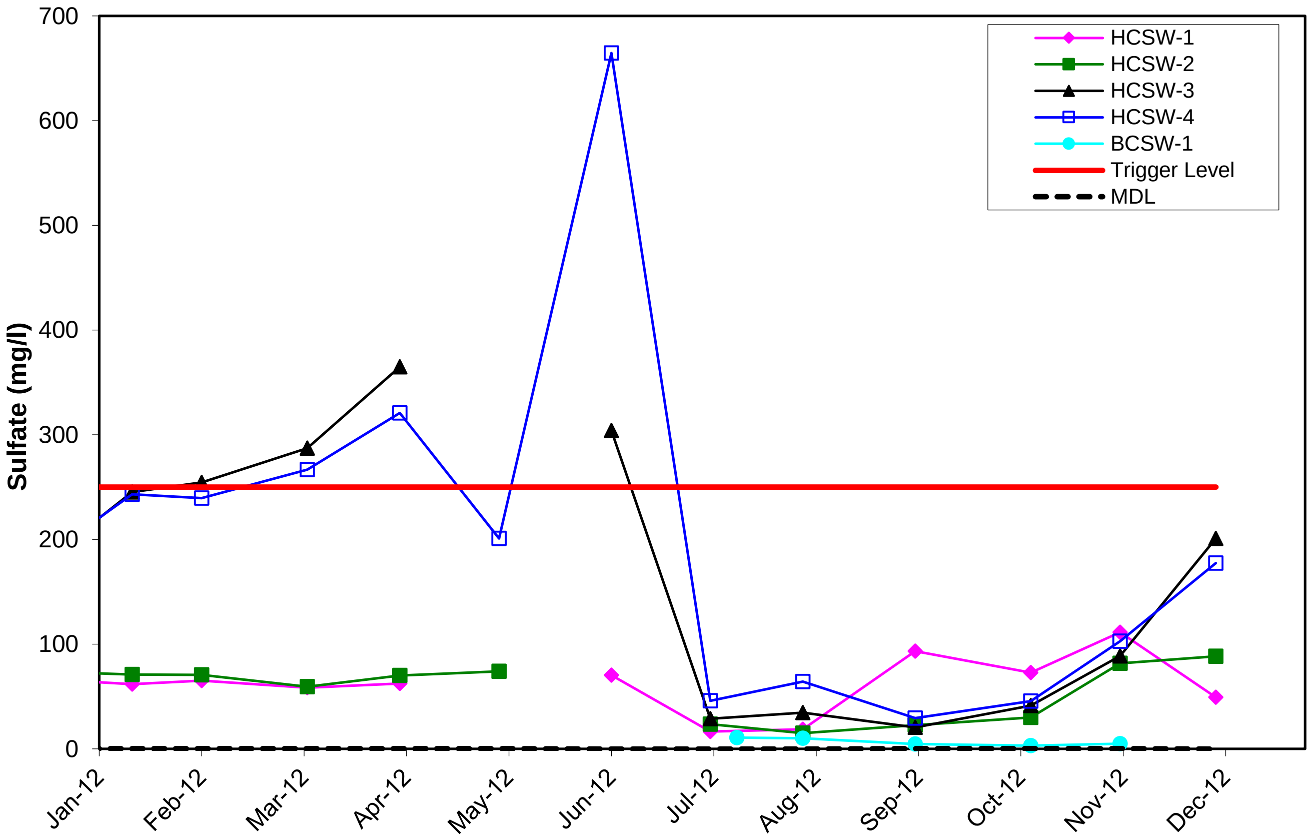

Figure 68. Sulfate Concentrations Obtained During Monthly HCSP Water Quality Sampling in 2012.

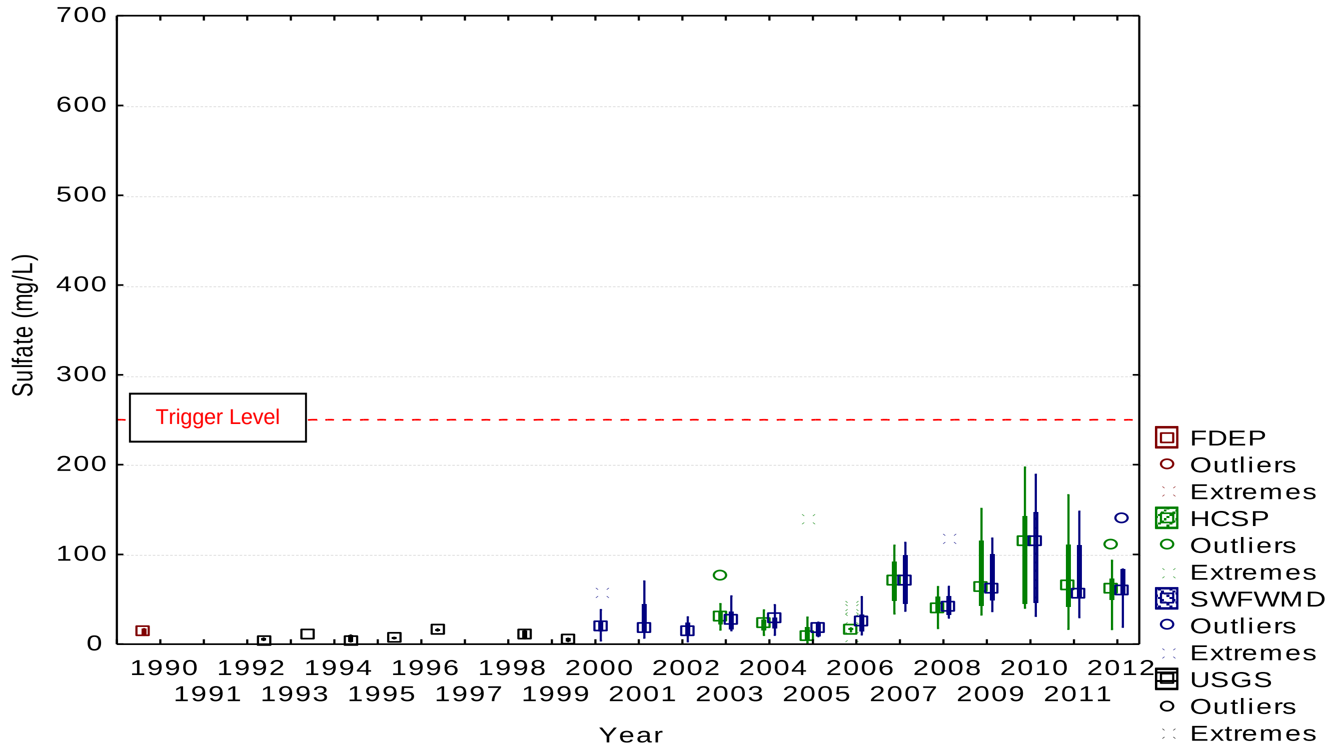

Figure 69. HCSW-1 Values of Sulfate Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

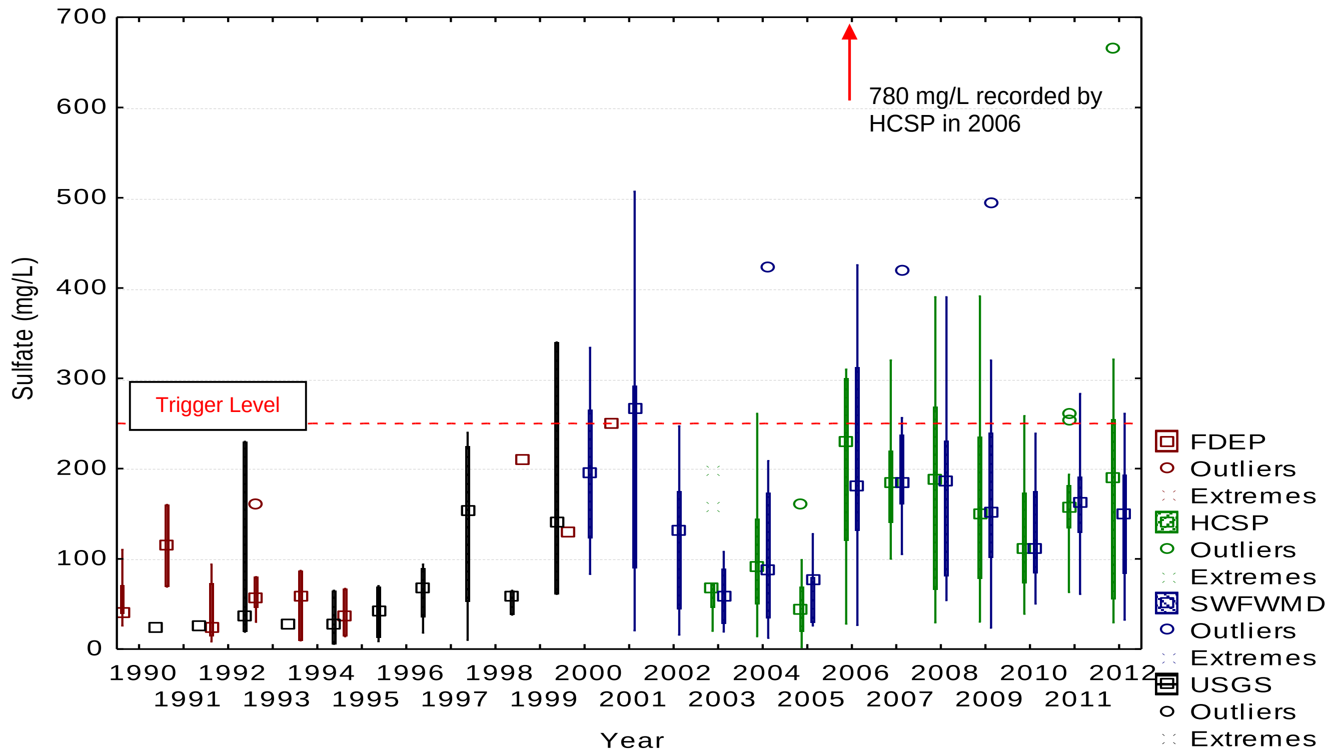

Figure 70. HCSW-4 Values of Sulfate Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.

Total Dissolved Solids

Total dissolved solids levels were below the trigger level of 500 mg/l during all sampling events at HCSW-1 and HCSW-2 in 2012; the trigger level was exceeded at HCSW-3 from January to April and June and at HCSW-4 from January to June 2012 (Figure 71). The TDS concentrations at HCSW-4 measured by the HCSP are consistent with other water quality data sources and exhibited no monotonic trend since 2003 (Seasonal Kendall Tau with LOWESS, $p > 0.05$); however, HCSW-1 exhibited an increasing trend since 2003 (Seasonal Kendall Tau with LOWESS, Tau = 0.35, $p = 0.02$, Sen slope = 6.64 mg/L per year, Figures 72 and 73, Table 15). The trend for TDS and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful (Appendix I).

As with sulfate concentrations, total dissolved solids levels over the course of the entire period of record were lowest at HCSW-1 and HCSW-2 and highest at HCSW-4 (ANOVA, $F = 41.08$, $p < 0.001$, Table 16; Duncan's multiple range test, $p < 0.05$, Figure 71). Total dissolved solid levels were negatively correlated with streamflow, rainfall, and NPDES discharge at HCSW-4 (Spearman's rank correlation $-0.21 > r > -0.71$, $p < 0.05$), but positively correlated with NPDES discharge at HCSW-1 (r_s =0.34, Table 17). Both sulfate and total dissolved solids at downstream stations are probably affected by agricultural irrigation return flows and groundwater seepage downstream in the same manner as discussed above for conductivity and calcium. Brushy Creek concentrations were lower than at Horse Creek stations. Dissolved ion concentrations at HCSW-1 have been affected by recent changes in mining practices, but the new levels appear to be stable and not biologically harmful (Appendix I).

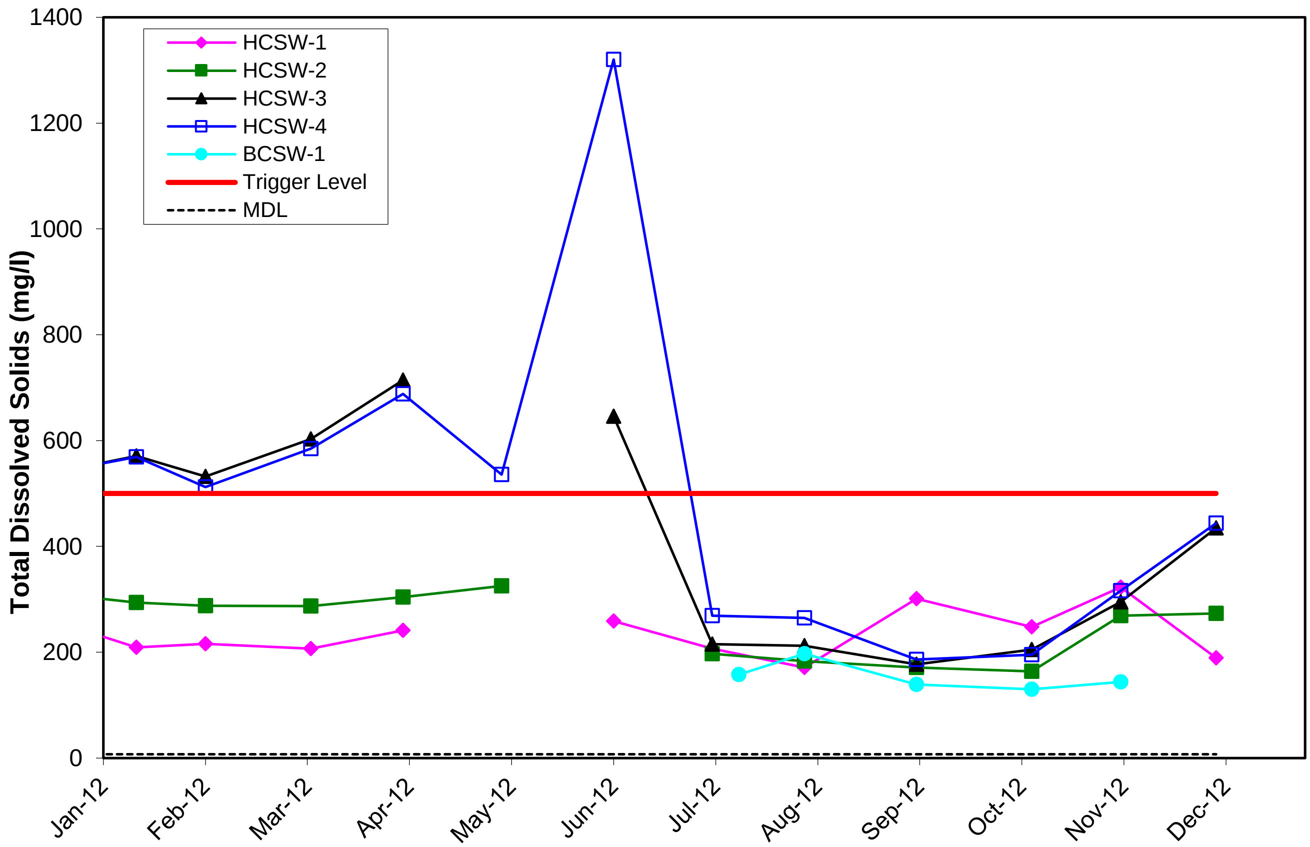

Figure 71. **Levels of Total Dissolved Solids Obtained During Monthly HCSP Water Quality Sampling in 2012.**

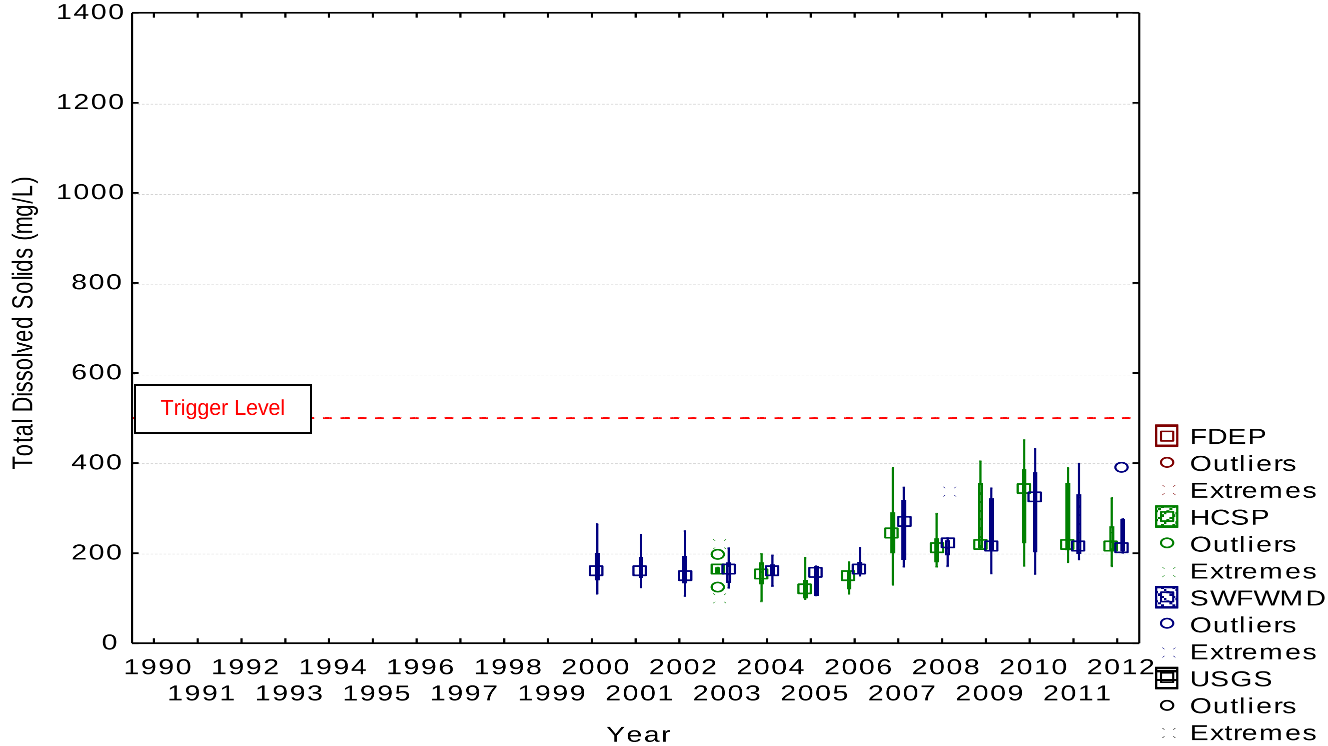

Figure 72. **HCSW-1 Values of Total Dissolved Solids Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

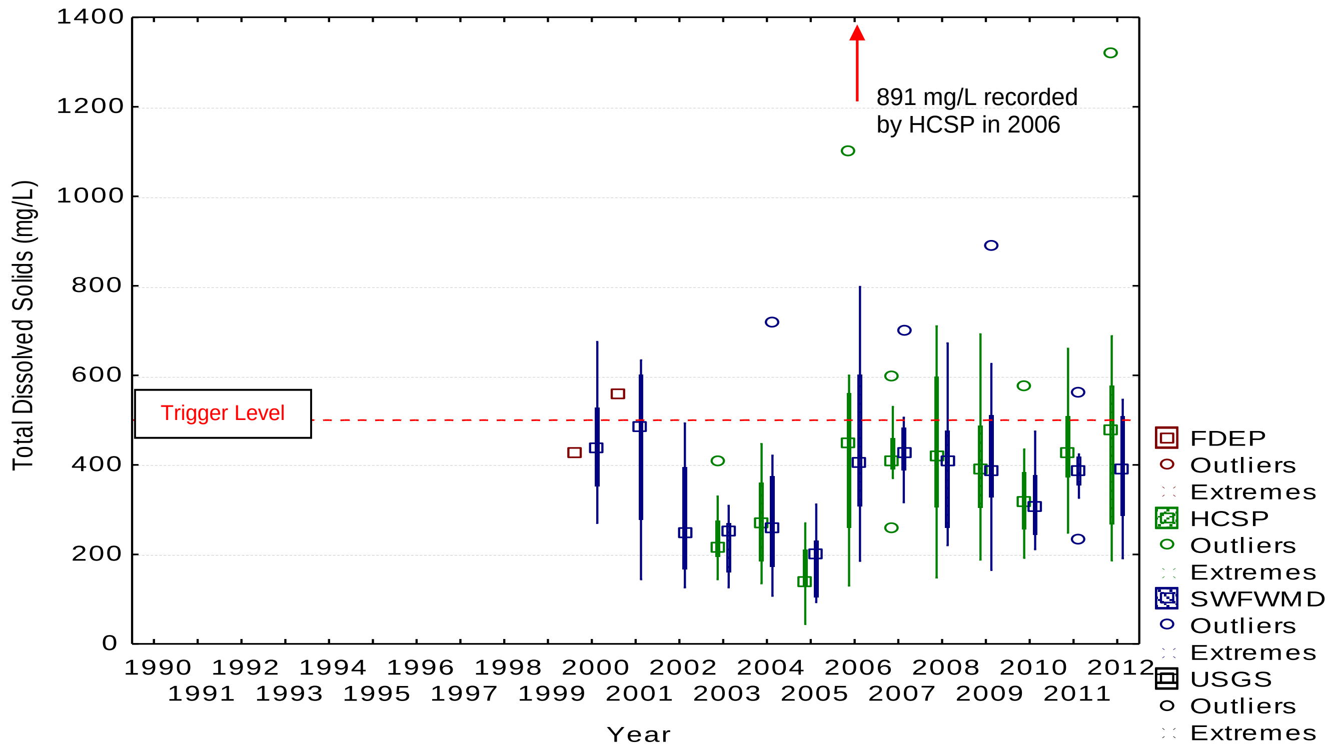

Figure 73. **HCSW-4 Values of Total Dissolved Solids Obtained from Various Data Sources (FDEP, SWFWMD, and USGS Data from EPA STORET and HCSP) for Years 1990 – 2012.**

Total Radium

Phosphate ore is a source of radioactivity as naturally occurring uranium-238 disintegrates into isotopes of radium and radon, which emit alpha particles in water. A water quality study of unmined and reclaimed basins in phosphate-mining areas found that radium concentrations of surface waters were slightly higher in unmined areas than in reclaimed basins, probably because of undisturbed phosphate deposits near the surface of unmined lands (Lewelling and Wylie 1993). Clay-settling areas may trap radioactive chemicals associated with clay slurry, but release only small amounts of radioactive chemicals into surface waters (Lewelling and Wylie 1993). The study also found that general radiochemical concentrations in groundwater from the surficial aquifer were greatly lower on unmined lands than reclaimed lands.

In Horse Creek during 2012, total radium[14] levels were below the trigger level of 5 pCi/L (Figure 74) at all stations during all sampling events. There were no monotonic trends observed since 2003 for total radium at HCSW-1 or HCSW-4 (Seasonal Kendall Tau, $p > 0.05$, Table 15). Total radium levels during 2003-2012 were significantly different among stations (ANOVA, $F = 3.12$, $p < 0.05$) with lowest levels at HCSW-2 (Duncan's multiple range test, $p < 0.05$, Table 16). Total radium was negatively correlated with NPDES discharge at HCSW-1 and HCSW-4 (Spearman's correlations $-0.35 > r > -0.38$, $p < 0.05$), but not correlated with streamflow or rainfall at either station (Spearman's correlations, $p > 0.05$, Table 17); indicating that radium was higher when NPDES discharge was low. Brushy Creek concentrations were similar to Horse Creek stations.

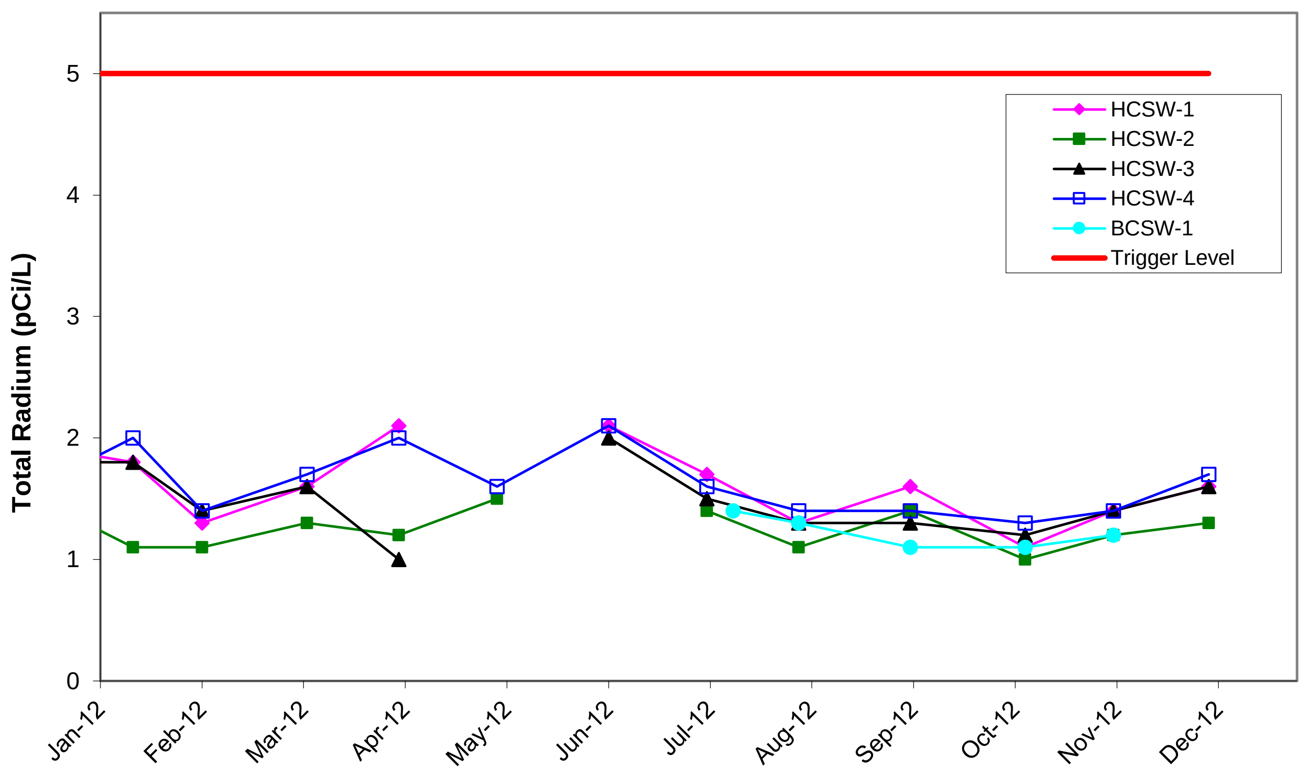

Figure 74. **Levels of Total Radium Obtained During Monthly HCSP Water Quality Sampling in 2012. (All of the samples were undetected for Radium 228 except HCSW-4 in December 2012.)**

[14] The HCSP methodology specifies that "Radium 226 + 228" be analyzed as part of the monthly sampling. This data has been reported as both individual constituents and as a total. The data in Appendix E reflects these changes. Starting in December 2003 and continuing through the present, the data has been analyzed and reported as Radium 226 and Radium 228 separately and an arithmetic sum of the two numbers ("Radium 226 + 228"). As requested by the PRMRWSA, if either Radium 226 or Radium 228 is undetected, the MDL of the undetected constituent will be used as part of the "Radium 226 + 228." This use of MDL for undetected constituents as part of a calculated constituent is contrary to both laboratory and DEP SOPs.

Total Fatty Acids, FL-PRO, and Total Amines

The phosphate beneficiation process that refines the mined phosphate ore uses several chemicals as reagents in the physio-chemical separation process. Three of these chemicals (fuel oil, fatty acids, and fatty amines) were selected for testing in the water-quality sampling program as potential indicator parameters of specific mining wastewater impacts. The FDEP Petroleum Range Organics (FL-PRO) test was selected as a test for fuel oil. Specific test methods were developed for fatty acids (obtained by Mosaic as a by-product of the paper industry and largely composed of oleic and linoleic acids) and fatty amines (fatty acids reacted with ammonia). FL-PRO, fatty acid, and amines all degrade biologically and/or photochemically within mine recirculation waters and clay settling areas (Patel and Schreiber 2001). These organic parameters were added to the HCSP monitoring list as an extra safeguard, although it was Mosaic's position that they would never be present at detectable limits in any waters discharged from mining areas.

In September 2009, these three parameters were removed from the HCSP based on recommendations from Mosaic, the TAG, and the PRMRWSA because they were seldom detected but may be added later should conditions warrant the addition.

5.2.5 **Summary of Water Quality Results**

Water quality parameters in 2012 were almost always within the desirable range relative to trigger levels and water quality standards at the station closest to mining (HCSW-1, Table 18). Trigger levels were exceeded only once at HCSW-1 in 2012 with the alkalinity exceedance in November. At HCSW-2, trigger levels were exceeded for dissolved oxygen during half of the year (Table 18). Based upon historical conditions in Horse Creek (Durbin and Raymond 2006), the reported values for dissolved oxygen at HCSW-2 are the result of natural conditions (proximity to hypoxic segment of stream – Horse Creek Prairie) and are not related to mining activities. The chlorophyll *a* trigger level was exceeded during low-flow periods at HCSW-2 in February, April, May, and December, and the lower pH trigger level was exceeded in October. HCSW-3 exceeded trigger levels for dissolved oxygen (June-September), calcium (April), sulfate (February-April and June), and TDS (January-April and June). HCSW-4 exceeded trigger levels for specific conductivity (June), dissolved oxygen (July and September), calcium (April-June), iron (July-October), alkalinity (May), sulfate (March-April and June), and TDS (January-June). Dissolved oxygen triggers were exceeded during summer wet months of 2012, when high temperatures reduce the oxygen carrying capacity of the stream. Sulfate, calcium, TDS, and other ions were exceeded in the dry season, when low rainfall and streamflow likely led to increased groundwater inputs from baseflow and agricultural runoff. Dissolved iron concentrations consistently exceeded the trigger value set at HCSW-4, but Mosaic and the PRMRWSA agree that the trigger value at that station has been set too low given historical and upstream concentrations of dissolved iron. Based on impact assessments already completed, none of the observed exceedances pose a significant adverse ecological impact to Horse Creek that would be attributable to mining.

Several of the trends detected during the statistical analysis either have an estimated slope that 1) was not in the direction of an adverse trend (color, ammonia, and dissolved iron) or 2) was very small compared to limits in laboratory prediction or the observed differences between primary and duplicate field samples (pH, ammonia, orthophosphate) (Table 19). For the trends with higher estimated rates of change (specific conductivity and various dissolved ions), the potential trends are discussed in Appendix I. Orthophosphate was also discussed further in this impact assessment as a follow-up to the trend impact assessment of the 2010 Annual Report (Robbins et al. 2014). Appendix I shows that the apparent trend in orthophosphate from 2003-2012 is caused by a data bias, and that extending the period of record into pre-2003 data eliminates this trend. Specific conductivity and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful.

A significant increasing trend in specific conductivity was detected at HCSW-1 even when the period of record was expanded, but the inclusion of data through 2012 (Appendix I) makes it clear that the most visible increase is a single step-change rather than a monotonic, continuing increase. In addition, at least part of the increase may be caused by regional influences, as Charlie Creek also shows a slow increase in conductivity over time. In addition, specific conductivity at Horse Creek and West Fork Horse Creek stations upstream of the NPDES outfalls has also been increasing over the same time period (data found in Mosaic DRI reports to Manatee County). Specific conductivity at HCSW-1 began to rise during a very dry period in 2006-2008 when Mosaic had very little NPDES discharge into Horse Creek, and the stream was dominated by baseflow (surficial aquifer) contributions. After the Horse Creek outfalls began to receive water decanting from Wingate Mine clays, the specific conductivity remained at the higher levels, presumably because of the greater proportion of groundwater involved in the dredge mining. If groundwater influence is the main cause of increased specific conductivity at HCSW-1, then concentrations have reached a threshold and should not continue to increase. This theory has been validated considering that the range and median specific conductivity from 2009 to 2012 has been very consistent with no expectation that it will increase from the recently observed range in the future.

Dissolved ion concentrations at HCSW-1 in recent years still meet all applicable state drinking water and Class III surface water standards, except where otherwise noted in previous monthly impact assessments. In addition, the current specific conductivity concentrations at HCSW-1 are within the preferred range of Horse Creek freshwater fish species, and the diversity and composition of fish populations have remained steady over the HCSP study period with no correlation with specific conductivity. SCI Scores, an indicator of benthic Macroinvertebrate community health, have remained steady over the HCSP study period and show no relationship with specific conductivity.

Significant differences between stations were evident for several parameters. Overall, HCSW-2 was the most dissimilar from the other three stations, especially in pH, dissolved oxygen, and some dissolved ions. Some nutrients (nitrate + nitrite) and dissolved ions (specific conductivity, calcium, chloride, sulfate) had higher concentrations downstream in Horse Creek, probably because of increased groundwater seepage and agricultural runoff in the lower Horse Creek basin. Differences in topography, geology, and land use that could account for these trends in Horse Creek are examined in the Horse Creek Stewardship Program Historical Report (Durbin and Raymond 2006).

Water quality parameters were compared with water quantity variables recorded during the same month: average daily streamflow for the month, average daily NPDES discharge for the month, and total monthly rainfall (Figure 75). In general, pH, dissolved oxygen, and most dissolved ions are higher when the overall quantity of water in the Horse Creek system is low. (In this report, specific conductivity, calcium, alkalinity, sulfate, and TDS showed the opposite pattern with NPDES discharge at HCSW-1; possible reasons behind this are discussed as part of the trend analysis in Appendix I.) Conversely, turbidity, color, iron, and nitrogen are high when the water quantity is also high. When water quantity in Horse Creek is low, the stream may be pooled or slow-moving, leading to algal blooms that may increase pH and chlorophyll a. In addition, the majority of water in the stream during low quantity periods may be from groundwater (seepage or agricultural runoff); groundwater has a higher concentration of dissolved ions than surface water. When water quantity is high, an increased amount of sediment and organic debris is washed into the stream, leading to increases in turbidity, color, iron, and nitrogen.

Table 18. Instances of Trigger Level Exceedance Observed in 2012 HCSP Monthly Monitoring.

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at Goose Pond Road	HCSW-2	10/10/2012	pH (SU)	5.96	6
Horse Creek at State Road 72	HCSW-4	6/5/2012	Specific Conductance (µmhos/cm)	1,425	1,275
Horse Creek at Goose Pond Road	HCSW-2	3/5/2012	Dissolved Oxygen (mg/l)	4.55	5
Horse Creek at Goose Pond Road	HCSW-2	5/2/2012	Dissolved Oxygen (mg/l)	3.32	5
Horse Creek at Goose Pond Road	HCSW-2	10/10/2012	Dissolved Oxygen (mg/l)	2.92	5
Horse Creek at Goose Pond Road	HCSW-2	11/6/2012	Dissolved Oxygen (mg/l)	3.95	5
Horse Creek at Goose Pond Road	HCSW-2	12/5/2012	Dissolved Oxygen (mg/l)	4.74	5
Horse Creek at State Road 70	HCSW-3	6/5/2012	Dissolved Oxygen (mg/l)	4.64	5
Horse Creek at State Road 70	HCSW-3	7/5/2012	Dissolved Oxygen (mg/l)	3.28	5
Horse Creek at State Road 70	HCSW-3	8/2/2012	Dissolved Oxygen (mg/l)	3.05	5
Horse Creek at State Road 70	HCSW-3	9/5/2012	Dissolved Oxygen (mg/l)	3.8	5
Horse Creek at State Road 70	HCSW-3	10/10/2012	Dissolved Oxygen (mg/l)	4.66	5
Horse Creek at State Road 72	HCSW-4	7/5/2012	Dissolved Oxygen (mg/l)	2.23	5
Horse Creek at State Road 72	HCSW-4	9/5/2012	Dissolved Oxygen (mg/l)	4.12	5
Horse Creek at State Road 70	HCSW-3	4/2/2012	Dissolved Calcium (mg/L)	114	100
Horse Creek at State Road 72	HCSW-4	4/2/2012	Dissolved Calcium (mg/L)	117	100
Horse Creek at State Road 72	HCSW-4	5/2/2012	Dissolved Calcium (mg/L)	109	100
Horse Creek at State Road 72	HCSW-4	6/5/2012	Dissolved Calcium (mg/L)	182	100
Horse Creek at Goose Pond Road	HCSW-2	2/2/2012	Chlorophyll a (mg/m3)	75.1	15
Horse Creek at Goose Pond Road	HCSW-2	4/2/2012	Chlorophyll a (mg/m3)	35.9	15
Horse Creek at Goose Pond Road	HCSW-2	5/2/2012	Chlorophyll a (mg/m3)	34.1	15
Horse Creek at Goose Pond Road	HCSW-2	12/5/2012	Chlorophyll a (mg/m3)	17.9	15
Horse Creek at State Road 72	HCSW-4	7/5/2012	Dissolved Iron (mg/L)	0.779	0.3
Horse Creek at State Road 72	HCSW-4	8/2/2012	Dissolved Iron (mg/L)	0.531	0.3
Horse Creek at State Road 72	HCSW-4	9/5/2012	Dissolved Iron (mg/L)	0.604	0.3
Horse Creek at State Road 72	HCSW-4	10/10/2012	Dissolved Iron (mg/L)	0.508	0.3
Horse Creek at State Road 64	HCSW-1	11/6/2012	Alkalinity (mg/L)	100.3	100
Horse Creek at State Road 72	HCSW-4	5/2/2012	Alkalinity (mg/L)	147.5	100
Horse Creek at State Road 70	HCSW-3	2/2/2012	Sulfate (mg/L)	254	250
Horse Creek at State Road 70	HCSW-3	3/5/2012	Sulfate (mg/L)	287	250
Horse Creek at State Road 70	HCSW-3	4/2/2012	Sulfate (mg/L)	365	250
Horse Creek at State Road 70	HCSW-3	6/5/2012	Sulfate (mg/L)	304	250
Horse Creek at State Road 72	HCSW-4	3/5/2012	Sulfate (mg/L)	267	250

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at State Road 72	HCSW-4	4/2/2012	Sulfate (mg/L)	321	250
Horse Creek at State Road 72	HCSW-4	6/5/2012	Sulfate (mg/L)	665	250
Horse Creek at State Road 70	HCSW-3	1/12/2012	TDS (mg/L)	571	500
Horse Creek at State Road 70	HCSW-3	2/2/2012	TDS (mg/L)	532	500
Horse Creek at State Road 70	HCSW-3	3/5/2012	TDS (mg/L)	603	500
Horse Creek at State Road 70	HCSW-3	4/2/2012	TDS (mg/L)	714	500
Horse Creek at State Road 70	HCSW-3	6/5/2012	TDS (mg/L)	646	500
Horse Creek at State Road 72	HCSW-4	1/12/2012	TDS (mg/L)	569	500
Horse Creek at State Road 72	HCSW-4	2/2/2012	TDS (mg/L)	512	500
Horse Creek at State Road 72	HCSW-4	3/5/2012	TDS (mg/L)	585	500
Horse Creek at State Road 72	HCSW-4	4/2/2012	TDS (mg/L)	688	500
Horse Creek at State Road 72	HCSW-4	5/2/2012	TDS (mg/L)	536	500
Horse Creek at State Road 72	HCSW-4	6/5/2012	TDS (mg/L)	1,320	500

Table 19. Summary of Trends over Time (2003-2012) from Seasonal Kendall Tau Analysis.

Parameter	HCSW-1 Slope	HCSW-4 Slope	Discussion
pH	0.05 SU/yr		Slope is within error range of pH probe; not ecologically significant
Color	5 PCU/yr	11 PCU/yr	Not an adverse trend
Ammonia*	-0.0003 mg/L/yr		Not an adverse trend
Iron	-0.02 mg/L/yr	-0.01 mg/L/yr	Not an adverse trend
Orthophosphate	0.02 mg/L/yr		Similar to pre-HCSP years; no trend over longer term; see further discussion in Appendix I
Conductance	11 μmhos/cm/yr		See further discussion in Appendix I
Calcium	1.1 mg/L/yr		Related to Conductance Trend Discussion (See Appendix I)
Alkalinity	3.0 mg/L/yr	1.7 mg/L/yr	
Sulfate	2.3 mg/L/yr		
TDS	6.6 mg/L/yr		

* Annual trends of SWFWMD data.

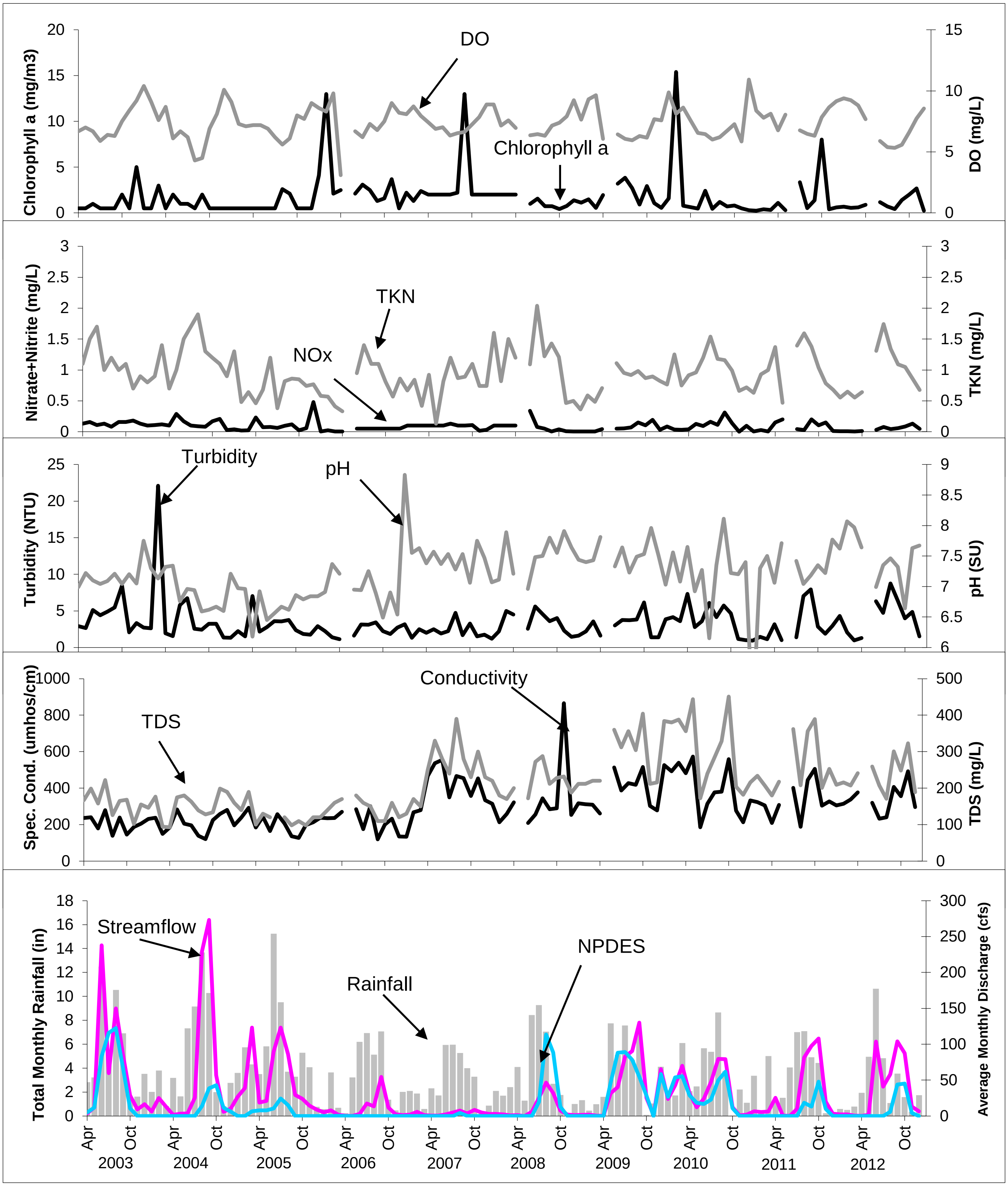

Figure 75. **HCSP Water Quality Correlations With Average Monthly NPDES Discharge, Average Monthly Streamflow, and Total Monthly Rainfall at HCSW-1 in 2003 – 2012.**

5.3 Benthic Macroinvertebrates

Benthic macroinvertebrate sampling was conducted at HCSW-4 on 30 March 2012, at all stations but HCSW-2 on 26 October 2012, and all stations but HCSW-2 on 12 December 2012. Only HCSW-4 was sampled in March 2012 because the other stations had little to no water or flow. The summer sampling event was delayed until the end of October to ensure that there were at least 90 days of flow after HCSW-1 and HCSW-3 were observed as "dry" in May 2012. No samples were collected at HCSW-2 in 2012 either due to dry/no flow (March and December) or flooded conditions (October).

As discussed in Section 4.4, the calculation methodology for the SCI was revised by DEP in June 2004, and sampling conducted in 2003–2006 uses that methodology. This change requires a departure from the specific metrics listed for benthic macroinvertebrates in the HCSP plan; however, the plan contemplated such changes in methodology and the use of the revised protocol is acceptable with the plan. Between the 2004 and 2005 biological sampling, individuals conducting biological sampling were trained and audited by the DEP in SCI and Stream Habitat Assessment techniques.

In 2007, the FDEP SCI protocol was again revised, this time to include provisions for the taxonomic analysis of two aliquots for every SCI sample collected to account for variation in sample sorting (DEP-SOP-002/01 LT 7200). Under the new protocol, the SCI score for each sample is the arithmetic average of the SCI scores of the two aliquots. In addition to the change in aliquots in 2007, the new protocol also gives a slightly different ecological interpretation of SCI scores (Table 7). Scores from the 2004 SCI (2003-2006) and the 2007 SCI (2007-2011) methods may not be directly comparable, given the differences in how they were collected.

5.3.1 Stream Habitat Assessment

The majority of the habitat assessment parameters evaluated through the DEP procedure are not directly related to mining, but are generally related to the nature of the system being examined and its surroundings (e.g., substrate diversity and availability, artificial channelization, bank stability, buffer width, and vegetation quality). Parameters that might be hypothesized to have some linkage to mining are water velocity and habitat smothering, primarily as a result of NPDES discharges to a stream. Since the turbidity of the NPDES discharge in 2012 was low, it is unlikely that suspended particles within the discharge made a significant contribution to sediment deposition in the stream. Habitat smothering in 2012 was low at HCSW-1 and fairly high at HCSW-3 and HCSW-4, because stream velocity was very low during the dry season followed by high runoff during the wet season.

For the habitat assessment metric on smothering along with the definition of smothering for the SCI it looks at the productive habitats and the degree to which they are smothered. HCSW-1 is higher up in the basin and would receive less sediment load that could smother the various habitats (roots, snags, and rock) from upstream sources. The more downstream locations have a larger basin area that contributes both sediment and flowing water. HCSW-3 and HCSW-4 have higher smothering that occurs in the productive habitats (roots, snags, and aquatic vegetation) usually after high flows when sediment settles out after flow decreases.

Overall, the habitat quality of Horse Creek was within the optimal or sub-optimal range during all sampling events in 2012 (Table 20), as it was for 2003 to 2011 (Figures 76 - 79[15]). Some of the minor variation among the sampling events for a given station primarily reflects differences in habitat quality and quantity caused by changes in stream stage, which affects the availability and ratios of in-stream habitats, and also the inherent variability in the habitat scoring protocol itself. The fall sampling event is usually immediately following summer high flows where the banks are scoured (habitat stability) and there may not be any vegetation in the water to sample as a productive habitat (substrate diversity and availability). For those reasons, the overall habitat assessment score tends to be lower in the summer or fall.

[15] An asterisk (*) represents a sampling event that did not meet the required SOP. See Appendix J for more details.

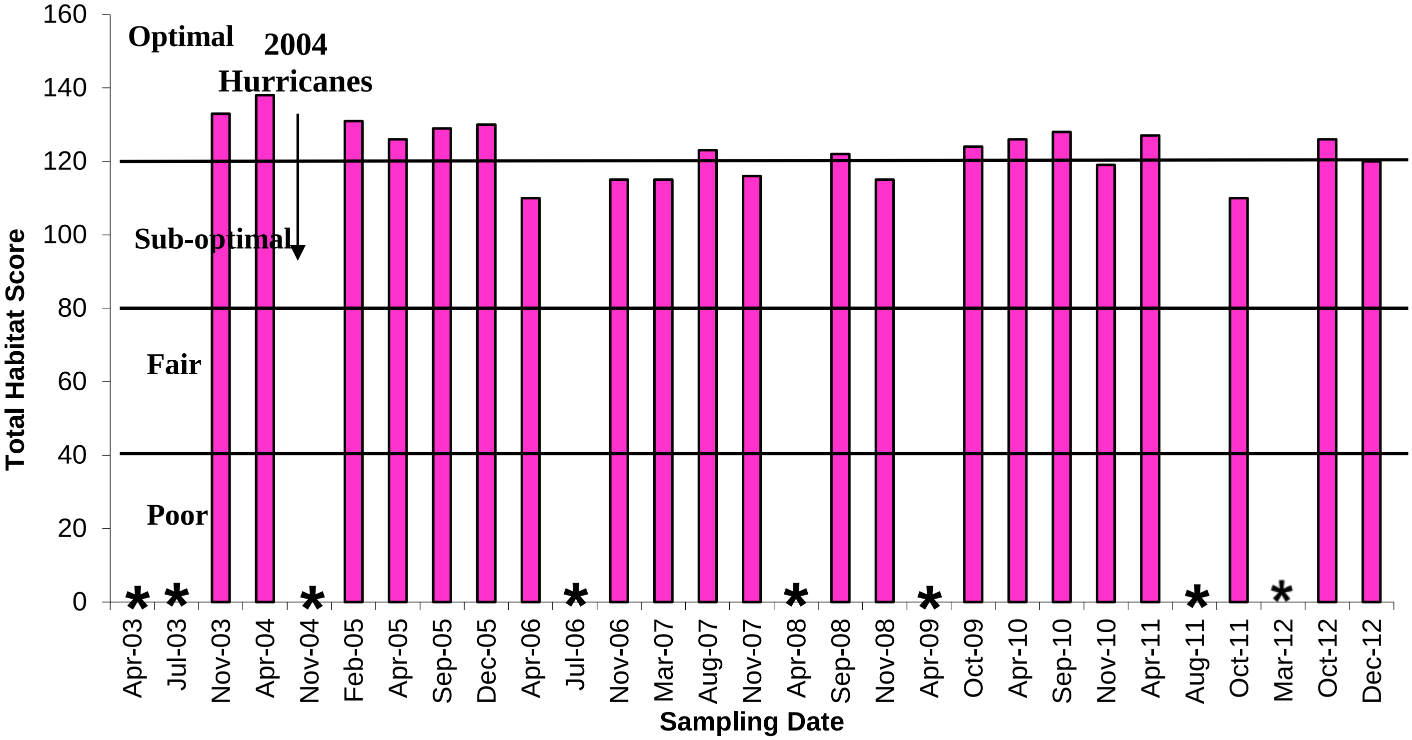

Figure 76. **Total Habitat Scores Obtained During HCSP Biological Sampling Events at HCSW-1 from 2003-2012. (HCSW-1 November 2004 score omitted because of sampler oversight.)**

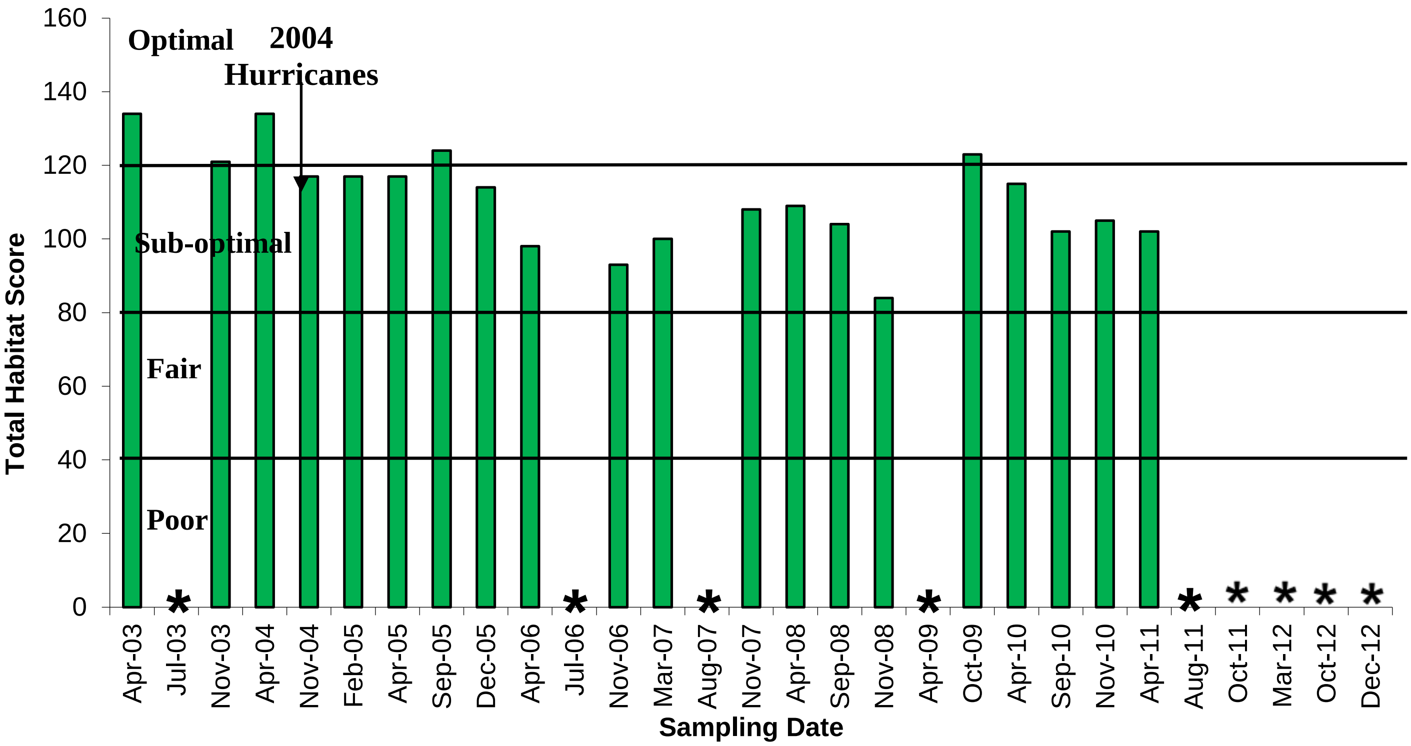

Figure 77. **Total Habitat Scores Obtained During HCSP Biological Sampling Events at HCSW-2 from 2003-2012.**

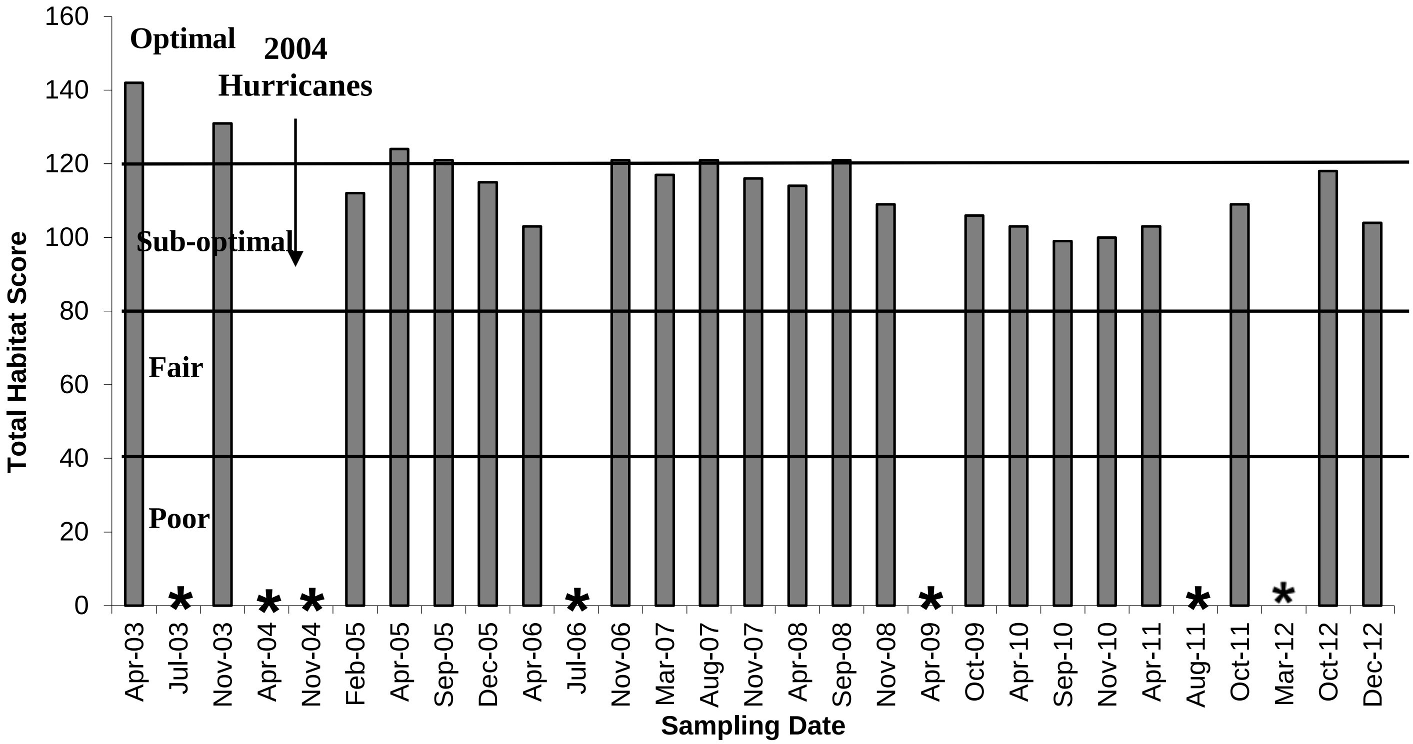

Figure 18. **Total Habitat Scores Obtained During HCSP Biological Sampling Events at HCSW-3 from 2003-2012.**

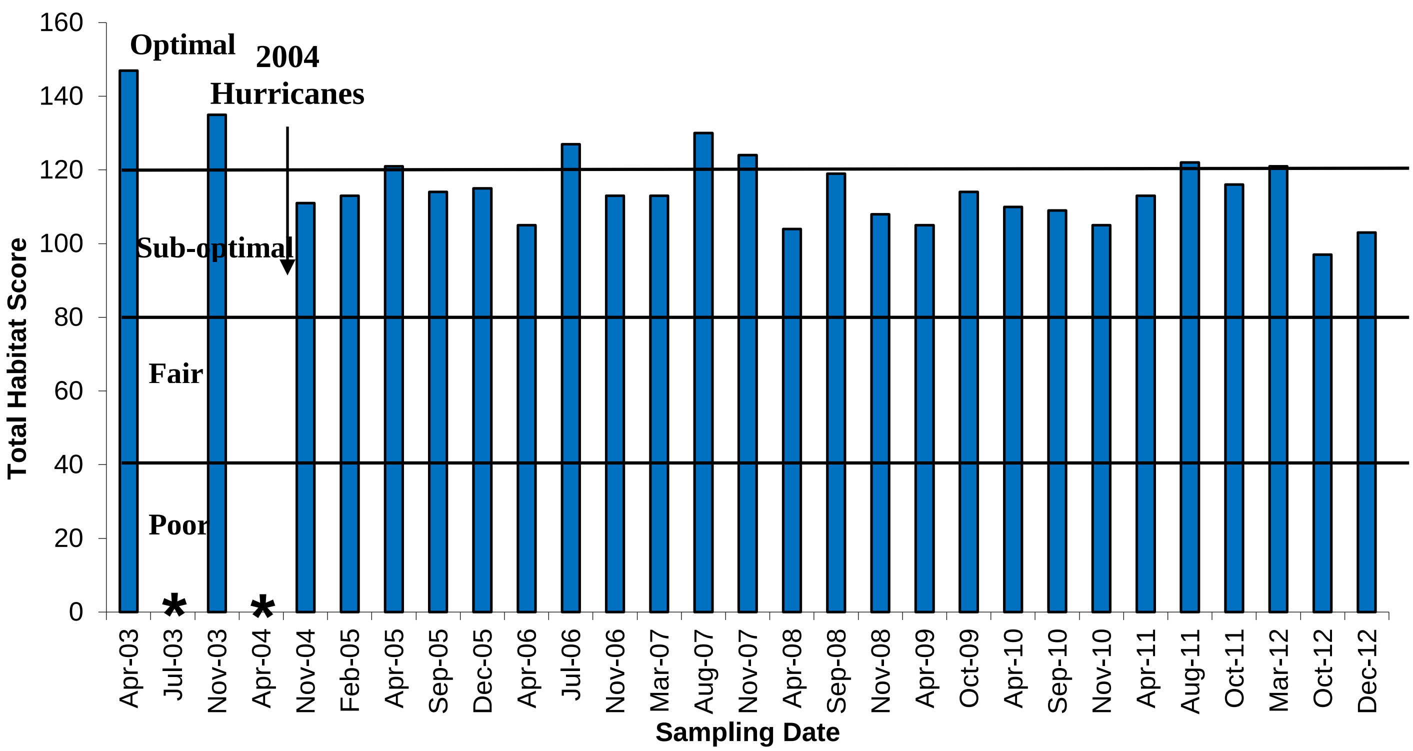

Figure 79. **Total Habitat Scores Obtained During HCSP Biological Sampling Events at HCSW-4 from 2003-2012.**

Table 20. Habitat Scores Obtained During HCSP Biological Sampling Events in 2012.

	HCSW-1			HCSW-2			HCSW-3			HCSW-4		
	30 March 2012	26 October 2012	12 December 2012	30 March 2012	26 October 2012	12 December 2012	30 March 2012	26 October 2012	12 December 2012	30 March 2012	26 October 2012	12 December 2012
Substrate Diversity	Low/No Flow – Not Sampled	13	11	Dry – Not Sampled	Flooded – Could Not Sample	Low/No Flow – Not Sampled	Low/No Flow – Not Sampled	12	7	9	6	6
Substrate Availability		9	13					5	3	16	3	4
Water Velocity		20	14					20	14	12	18	18
Habitat Smothering		18	18					13	16	17	11	16
Artificial Channelization		20	20					20	20	20	20	20
Bank Stability												
Right Bank		5	5					7	5	7	5	5
Left Bank		5	5					5	5	7	5	5
Riparian Buffer Zone Width												
Right Bank		10	10					10	10	10	10	10
Left Bank		10	10					10	10	5	5	5
Riparian Zone Vegetation Quality												
Right Bank		8	8					8	7	9	7	7
Left Bank		8	8					8	7	9	7	7
Total Score*		**126**	**120**					**118**	**104**	**121**	**97**	**103**
Habitat Descriptor		Optimal	Optimal					Sub-Optimal	Sub-Optimal	Optimal	Sub-Optimal	Sub-Optimal

* - The maximum possible score under this protocol is 160 (120-160 Optimal, 80-119 Suboptimal, 40-79 Marginal, <40 Poor).

Table 21. SCI Metrics Calculated for Benthic Macroinvertebrates Collected at Four Locations on Horse Creek for the HCSP During 2012.

SCI Metric	HCSW-1						HCSW-2					
	30 March 2012		26 October 2012		12 December 2012		30 March 2012		26 October 2012		12 December 2012	
	Raw	SCI	Raw	SCI	Raw	SCI	Raw	SCI	Raw	SCI	Raw	SCI
Total Taxa			20	**1.6**	27	**4.2**						
Ephemeropteran Taxa			2.0	**4.0**	1.5	**3.0**						
Trichopteran Taxa			3.0	**4.3**	5.0	**7.1**						
Percent Filterer Taxa			19	**4.5**	9.0	**2.1**						
Long-lived Taxa			1.0	**2.5**	0.5	**1.3**						
Clinger Taxa			6.5	**8.1**	6.5	**8.1**						
Percent Dominant			26	**6.3**	40	**3.1**						
Percent Tanytarsini			6.7	**6.2**	8.4	**6.7**						
Sensitive Taxa			3.0	**3.3**	3.0	**3.3**						
Percent Very Tolerant			1.7	**7.6**	2.2	**7.3**						
Total SCI Score	**No SCI – Low/No Flow**		**53.8**		**51.4**		**No SCI – Dry**		**No SCI – Flooded**		**No SCI – Low/No Flow**	
Interpretation			**Healthy**		**Healthy**							
Total Number of Individuals			**150**		**161**							

SCI Metric	HCSW-3						HCSW-4					
	30 March 2012		26 October 2012		12 December 2012		30 March 2012		26 October 2012		12 December 2012	
	Raw	SCI	Raw	SCI	Raw	SCI	Raw	SCI	Raw	SCI	Raw	SCI
Total Taxa			29	**5.2**	4.1	**9.4**	46	**10**	23	**2.8**	36	**7.8**
Ephemeropteran Taxa			5.0	**10**	4.5	**9.0**	6.5	**10**	5.5	**10**	3.5	**7.0**
Trichopteran Taxa			2.5	**3.6**	3.5	**5.0**	4.5	**6.4**	2.5	**3.6**	3.5	**5.0**
Percent Filterer Taxa			42	**10**	15	**3.5**	23	**5.7**	38	**9.6**	21	**5.1**
Long-lived Taxa			1.0	**2.5**	2.0	**5.0**	2.0	**5.0**	1.5	**3.8**	1.5	**3.8**
Clinger Taxa			6.5	**8.1**	7.5	**9.4**	4.5	**5.6**	7.0	**8.8**	7.5	**8.8**
Percent Dominant			27	**6.3**	14	**9.0**	25	**6.6**	25	**6.6**	28	**6.0**
Percent Tanytarsini			0.7	**1.55**	9.8	**7.2**	37	**10**	1.3	**2.4**	1.8	**3.1**
Sensitive Taxa			3.5	**3.9**	4.0	**4.4**	3.0	**3.3**	4.5	**5.0**	4.0	**4.4**
Percent Very Tolerant			13	**3.8**	17	**3.0**	18	**2.9**	9.0	**4.5**	6.7	**5.0**
Total SCI Score	**No SCI – Low/No Flow**		**61.0**		**72.1**		**72.9**		**63.3**		**62.1**	
Interpretation			**Healthy**		**Exceptional**		**Exceptional**		**Healthy**		**Healthy**	
Total Number of Individuals			**150**		**157**		**153**		**151**		**142**	

5.3.2 Stream Condition Index

A database containing a list of the benthic macroinvertebrate taxa collected during 2003 - 2012 is on the attached CD-ROM[16]. Table 21 provides the SCI metrics, resulting SCI values, and total SCI scores calculated as a vial average for the benthic macroinvertebrates collected at the four stations during each sampling event in 2012. The numbers of individuals included in Table 21 represent the number extracted from the whole sample for identification (i.e., all 20 dipnet sweeps), which were analyzed by the taxonomist (only a portion of each sample is sorted and processed, per the SOP). The various components of the SCI calculations are briefly described in the subsections below.

5.3.2.1 *Total Taxa*

In general, a healthy stream system will support colonization by a diverse number of taxa. Therefore, the more taxa a station is shown to have, the healthier that system is regarded. Figures 80-83[17] illustrate the number of taxa collected at each of the HCSP stations during the monitoring events. Differences in taxa numbers among samples are expected, both spatially and temporally, as a result of natural variability, as well as differences in sampling conditions and sample processing, even when the invertebrate communities are very similar. The number of invertebrate taxa collected in each sample was similar to historic sampling in the basin (Durbin and Raymond 2006). When considered over time from 2003 to 2012, total taxa were variable over time at each station, with their scores increasing over time (Kendall Tau = 0.25, $p < 0.05$). Total taxa scores also increased at HCSW-3 and HCSW-4 from 2003 to 2012 (Kendall Tau = 0.61 and 0.47, respectively, $p < 0.05$). The total taxa scores were not significantly different between stations (ANOVA: F = 2.32, p = 0.08).

[16] For the 2010 annual report and subsequent reports, we have reevaluated the HCSP SCI data with strict interpretation of FDEP SOP guidance (Appendix J), including the upper and lower limits of the SOP target number of individuals, the SOP target of 90 days of previous flow, and the SOP target of less than a 0.5 m water level increase in the previous 30 days. As a result of this evaluation, some SCI scores have been removed from this analysis (Appendix J, *red italics*). In addition, some SCI results with more than the target number of individuals were randomly resampled to provide an unbiased result. In future reports, those samples with less than the SOP target range of individuals may be revaluated if sufficient material can be resorted from stored samples.

[17] An asterisk (*) represents a sampling event that did not meet the required SOP. See Appendix J for more details.

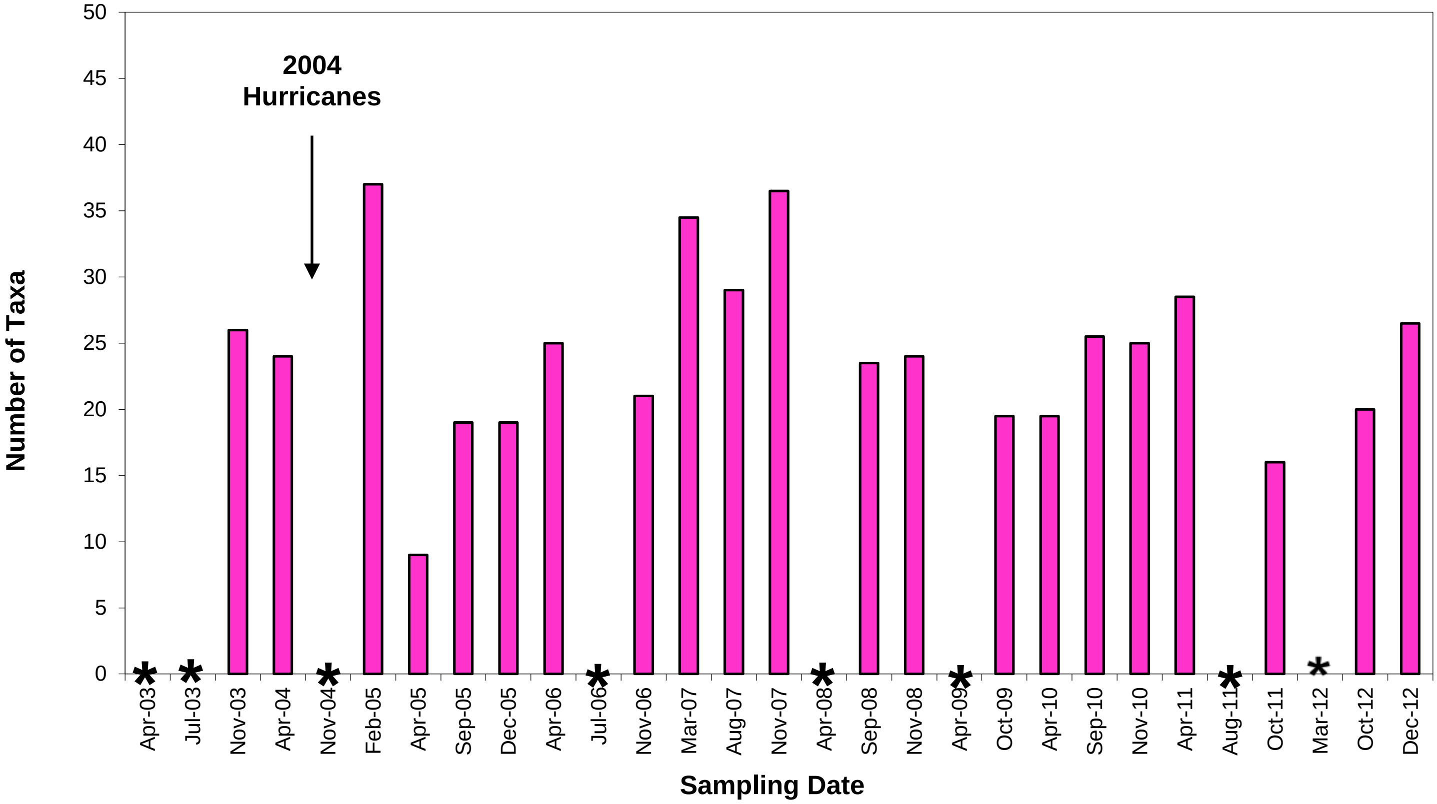

Figure 80. **Number of Invertebrate Taxa Collected from HCSW-1 for the HCSP in 2003 – 2012.**

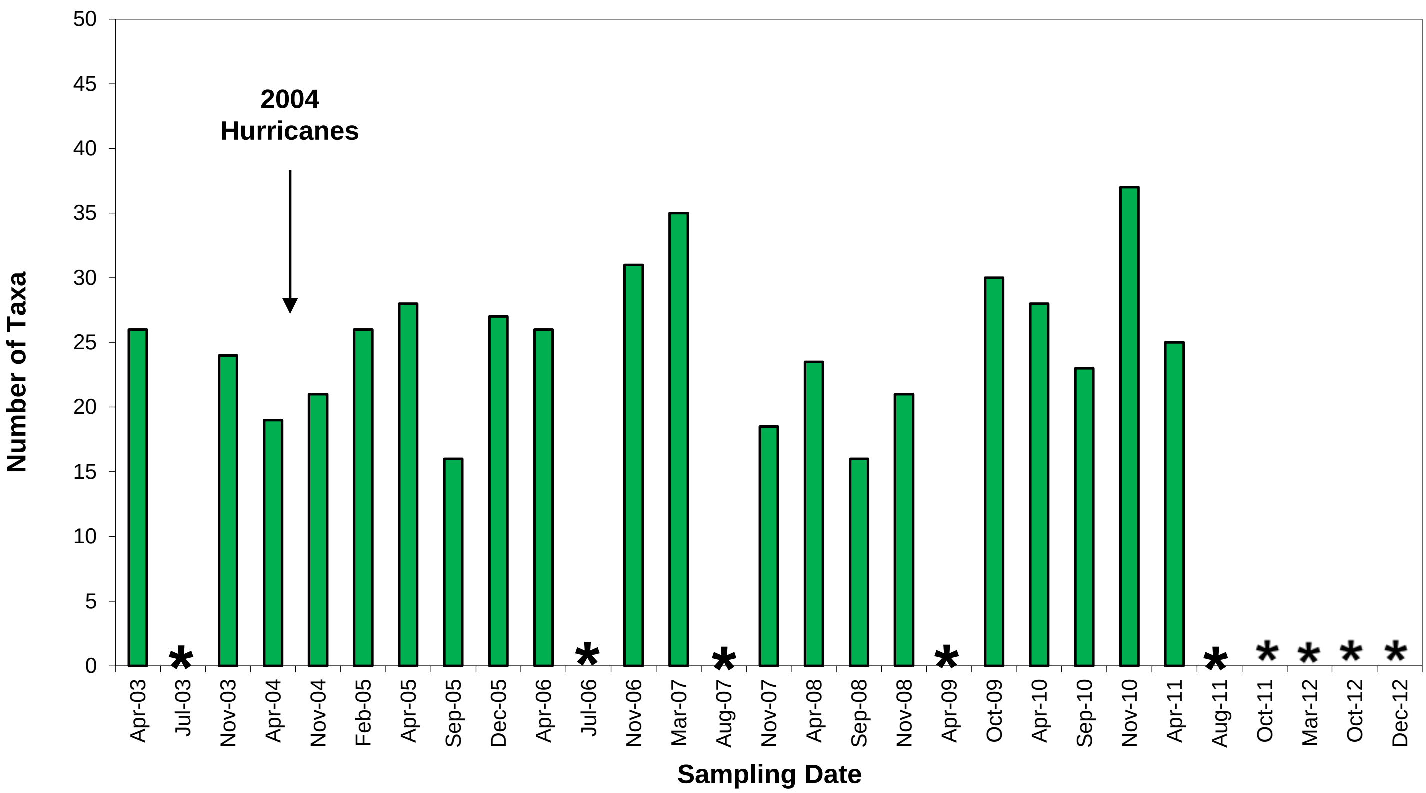

Figure 81. **Number of Invertebrate Taxa Collected from HCSW-2 for the HCSP in 2003 - 2012.**

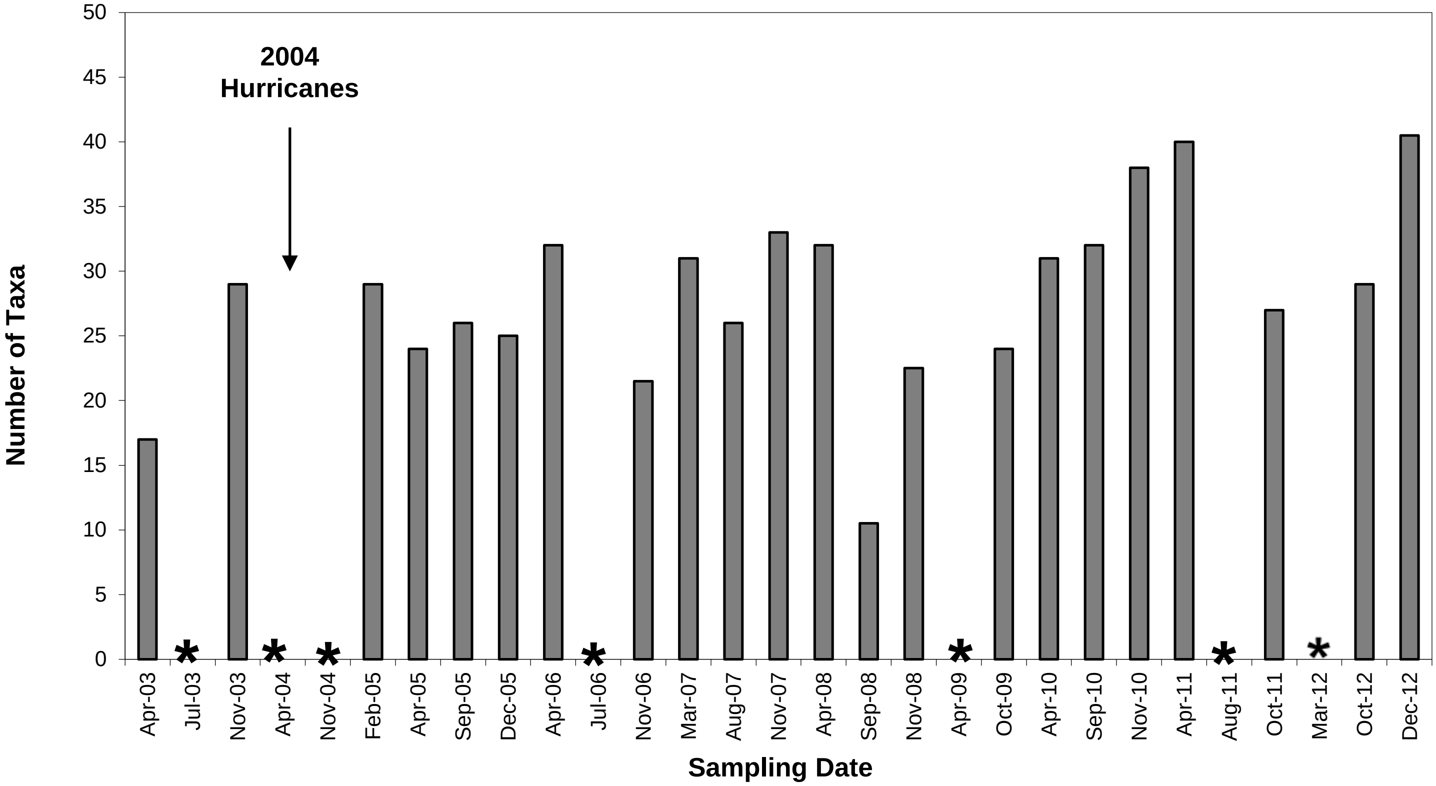

Figure 82. **Number of Invertebrate Taxa Collected from HCSW-3 for the HCSP in 2003 – 2012.**

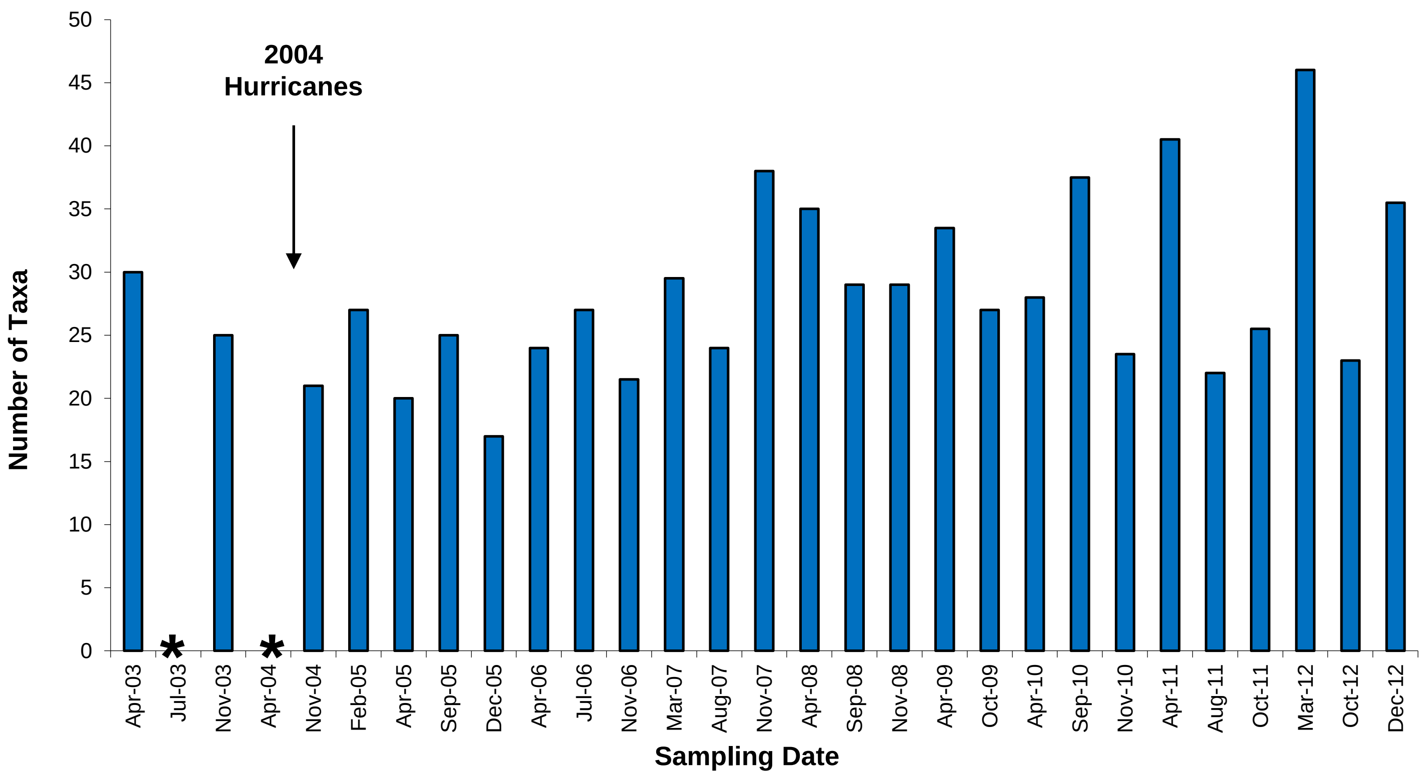

Figure 83. **Number of Invertebrate Taxa Collected from HCSW-4 for the HCSP in 2003 – 2012.**

5.3.2.2 *Ephemeroptera Taxa*

Ephemeropterans (mayflies) are typically associated with more pristine waters and better habitat conditions. A higher taxa count for this group is associated with better habitat value. At least one mayfly taxon must be present to score a SCI metric above zero. This metric was never zero in 2012. The greatest number of mayfly taxa collected at any station during any event in 2012 was seven (in March at HCSW-4). Although the number of Ephemeroptera taxa was as high as six at some sites used in developing the SCI calculation protocols, typical samples produce only 0-2 taxa (Fore 2004). This is consistent with the findings from the Horse Creek stations (Table 21). When considered over time from 2003 to 2012, Ephemeroptera taxa were variable over time at each station, but their scores did not increase or decrease over time (Kendall Tau, $p > 0.05$); scores were significantly lower at HCSW-2 followed by HCSW-1 (ANOVA: $F = 18.63$, $p < 0.0001$, Duncan's multiple range test: $p < 0.05$). Examples of common mayfly species collected in 2012 were *Maccaffertium exiguum*, *Caenis hilaris*, *Caenis diminuta*, and *Tricorythodes albilineatus.*

5.3.2.3 *Trichoptera Taxa*

Trichopterans (caddisflies) are also associated with more pristine waters and better habitats, so higher counts of caddisflies are associated with better ecological conditions. At least one taxon must be collected in order for the SCI metric to be above zero. This metric was not zero in 2012. The greatest number of caddisfly taxa in any sample in 2012 was six (in December at HCSW-1). According to Fore (2004), caddisfly taxa ranged from zero to eight in samples used for calibrating the SCI protocol, with most samples having four or fewer taxa. This is quite comparable to the observed pattern from Horse Creek in 2012 (Table 21). When considered over time from 2003 to 2012, Trichoptera taxa were variable over time at each station, with their scores increasing over time (Kendall Tau = 0.37, $p < 0.05$). Trichoptera scores also increased at HCSW-1, HCSW-3 and HCSW-4 from 2003 to 2012 (Kendall Tau = 0.58, 0.59, and 0.66, respectively, $p < 0.05$). Overall, scores were significantly lower at HCSW-2 than other stations (ANOVA: $F = 15.30$, $p < 0.0001$, Duncan's multiple range test: $p < 0.05$). Examples of common caddisfly species found in Horse Creek in 2012 were *Cheumatopsyche* sp., *Neotrichia* sp., and *Hydropsyche rossi*.

5.3.2.4 *Percent Collector-Filterer Taxa*

Taxa whose functional feeding group is "collector-filterer" are often more prolific in pristine natural waters. A reduction in the collector-filterer community can indicate a water quality problem. The SCI metric increases as the percentage of a sample comprised by these taxa increases. To score above zero for this metric, more than one percent of the sample must be composed of collector-filterers. Samples at each station during each 2012 event were composed of seven to just over forty-two percent collector-filterers (Table 21). The percentages were lowest during the December sampling event at all stations. This is within the range reported by Fore (2004) in developing the SCI calculation protocol. When considered over time from 2003 to 2012, collector-filterers taxa were variable over time at each station, and their scores had a slight decrease over time at HCSW-2 (Kendall Tau = -0.50, $p < 0.05$); scores were not significantly different among stations (ANOVA: $F = 1.67$, $p = 0.18$). Examples of filter feeder species collected in 2012 were *Cheumatopsyche* sp., *Tanytarsus* species, and *Hydropsyche rossi*.

5.3.2.5 *Long-lived Taxa*

Long-lived taxa are those that require more than one year to complete their life cycles (Fore, 2004), so they would not be expected in great numbers in intermittent streams or tributaries that go dry before their life cycle can be completed. Some long-lived taxa might also be less frequently encountered in less pristine waters, where these taxa could be exposed to potential contaminants for longer than their short-lived counterparts. To score above zero for this SCI metric, at least one long-lived taxon must be present in a sample. In 2012, the number of long-lived taxa ranged from zero to two (Table 21). The observed range of long-lived taxa (0 - 5 taxa) in samples collected from Horse Creek corresponds with the range used to develop the SCI methodology (Fore 2004). When considered over time from 2003 to 2012, long-lived taxa were variable over time at each station, but their scores did not increase or decrease (Kendall Tau, $p > 0.05$); scores were not significantly different among stations (ANOVA: $F = 2.54$, $p = 0.08$). Examples of long-lived species collected in 2012 were Palaemonetes *paludosus Corydalus cornutus*, and *Corbicula fluminea*.

5.3.2.6 *Clinger Taxa*

Taxa whose mode of existence is identified as clinging by Merritt and Cummins (1996) are defined as "having behavioral (e.g., fixed retreat construction) and morphological adaptations for attachment to surfaces in stream riffles." The SCI metric increases as the number of clinger taxa increases within a sample. To score above zero for this SCI metric, at least one clinger taxon must be present in a sample. Clinger taxa were found at the three sampled stations during all events in 2012, with the most in any sample being nine at HCSW-4 in December (Table 21). While Fore (2004) reported more than ten clinger taxa in some cases, most samples used to develop the SCI protocol had less than five taxa. When considered over time from 2003 to 2012, clinger taxa were variable over time at each station with an overall slight increase (Kendall Tau = 0.31, $p < 0.05$) along with scores increasing over time at HCSW-1 (Kendall Tau = 0.60, $p < 0.05$); scores were significantly lower at HCSW-2 than other stations (ANOVA: F = 34.38, $p < 0.0001$, Duncan's multiple range test: $p < 0.05$). Common clinger species found in Horse Creek in 2012 were *Cheumatopsyche* sp., *Stenelmis hungerfordi*, *Hydropsyche rossi*, and *Neotrichia sp*.

5.3.2.7 *Percent Dominant Taxon*

As the contribution of the dominant taxon increases, the diversity of taxa within a system generally decreases. Therefore, higher percent contribution by one taxon is interpreted as less ecologically desirable, and lowers the numerical value associated with this metric. The SCI score is zero if the percentage contribution of the dominant taxon is at or above 54 percent. Overall, six of the seven samples in 2012 had a single taxon representing more than one fourth of the invertebrate community (Table 21). Even though the amphipod *Hyalella azteca* complex was present at all stations and has been a dominant species in the past, it was not the dominant species for all stations in 2012; instead each station was dominated by a different order of invertebrates. The coleopterans (beetles) and dipterans (flies) dominated HCSW-1, with *Microcylloepus pusillus* and *Polypedilum flavum*. (No samples were collected from HCSW-2 in 2012.) Trichopterans (caddisfly) dominated HCSW-3, with *Cheumatopsyche* sp., *Neotrichia*, and *Hydropsyche rossi*. At HCSW-4, dipterans dominated the March (*Tanytarsus sp. C*) and December (*Polypedilum flavum*) samples, while a trichopteran was the dominant species in October (*Cheumatopsyche* sp.). The dominant taxa vary from year to year, with the 2012 samples dominated by dipterans, trichopterans, and coleopterans. When considered over time from 2003 to 2012, percent dominant taxa were variable over time at each station, but their scores did not increase or decrease over time (Kendall Tau, $p > 0.05$); scores were not significantly different between stations (ANOVA: F = 2.23, $p > 0.05$).

5.3.2.8 *Percent Tanytarsini*

Species in the chironomid tribe Tanytarsini (comprising several genera found in Florida) are commonly associated with less disturbed sites. Therefore, as the percentage of Tanytarsini increases for a sampling site, the SCI metric score also increases. If no Tanytarsini individuals are collected in a sample, this SCI metric score is zero; the percent Tanytarsini ranged from 0.7% to 37.4% in 2012. The percentages were lowest during the October sampling event at all three stations. When considered over time from 2003 to 2012, Tanytarsini taxa were variable over time at each station, but their scores did not increase or decrease over time (Kendall Tau, $p > 0.05$). Tanytarsini taxa were not significantly different between stations (ANOVA: F = 0.45, p = 0.71). Common chironomids found in 2012 were various *Tanytarsus* and *Rheotanytarsus* species.

5.3.2.9 *Sensitive Taxa*

Sensitive taxa are those that have been identified as sensitive to human disturbance (Fore, 2004). Using this definition, one would expect to find more sensitive taxa in undeveloped "natural" areas as opposed to developed watersheds. At least one sensitive taxon must be collected to raise this SCI metric score above zero. The number of sensitive taxa collected at Horse Creek stations in 2012 ranged from two to five (Table 21). When considered over time from 2003 to 2012, sensitive taxa were variable over time, and had a decreasing trend at HCSW-2 (Kendall Tau = -0.55, $p < 0.05$); scores were significantly lower at HCSW-2 (ANOVA: F = 17.88, $p < 0.0001$, Duncan's multiple range test: $p < 0.05$). Examples of common sensitive species found in 2012 were *Hydropsyche rossi*, *Maccaffertium exiguum*, *Tricorythodes albilineatus* and *Simulium* species.

5.3.2.10 *Percent Very Tolerant Taxa*

Fore (2004), classified a number of taxa as "very tolerant", meaning they are commonly present in areas with marked human disturbance (although they may also be found in undisturbed sites). More disturbed and/or developed areas, therefore, would be expected to have a higher percentage of tolerant taxa in comparison to areas that have not experienced human disturbance. This SCI metric is similar to the percent contribution of dominant taxa in that, as the fraction of a sample comprised by tolerant taxa increases, the calculated metric decreases. If the percentage of very tolerant taxa reaches or surpasses fifty-nine percent, the SCI metric is zero. Very tolerant taxa scores were never zero in 2012 but were fairly high ranging from 2.9 to 7.6 (Table 21). When considered over time from 2003 to 2012, percent very tolerant taxa were variable over time at each station but did not increase or decrease over time (Kendall Tau, $p > 0.05$). Scores were significantly lower at HCSW-2 and higher at HCSW-1 than other stations (ANOVA: $F = 13.60$, $p < 0.0001$, Duncan's multiple range test: $p < 0.05$). Common very tolerant taxa found in Horse Creek in 2012 included *Pyrgophorus platyrachis, Enallagma coecum*, and *Polypedilum illinoense grp*.

5.3.2.11 *SCI Overall Score*

Final SCI scores for the samples (with the recommended range of individuals in the sorted portion) ranged from 51 to 73 in 2012, similar to other years (Table 21 and Figures 84 - 87[18]). When considered over time from 2003 to 2012, the overall SCI scores were variable at each station but did not increase or decrease over time (Kendall Tau, $p > 0.05$); however, there was an increasing trend at HCSW-3 for the SCI score over time (Kendall Tau = 0.56, $p < 0.05$). The SCI scores were also significantly lower at HCSW-2 than other stations (ANOVA: $F = 15.60$, $p < 0.0001$, Duncan's multiple range test: $p < 0.05$).

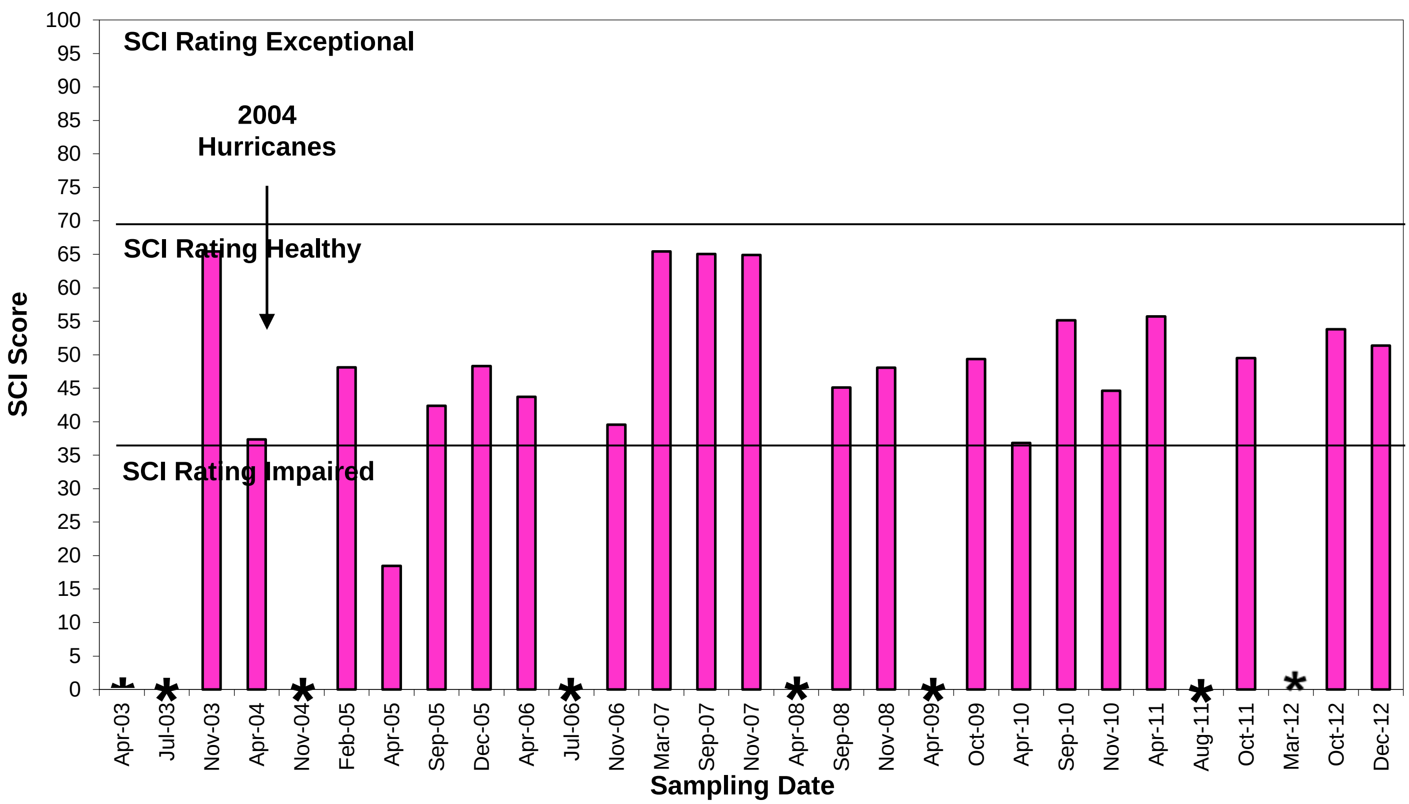

Figure 84. SCI Scores for Samples Collected at HCSW-1, 2003 - 2012.

[18] An asterisk (*) represents a sampling event that did not meet the required SOP. See Appendix J for more details.

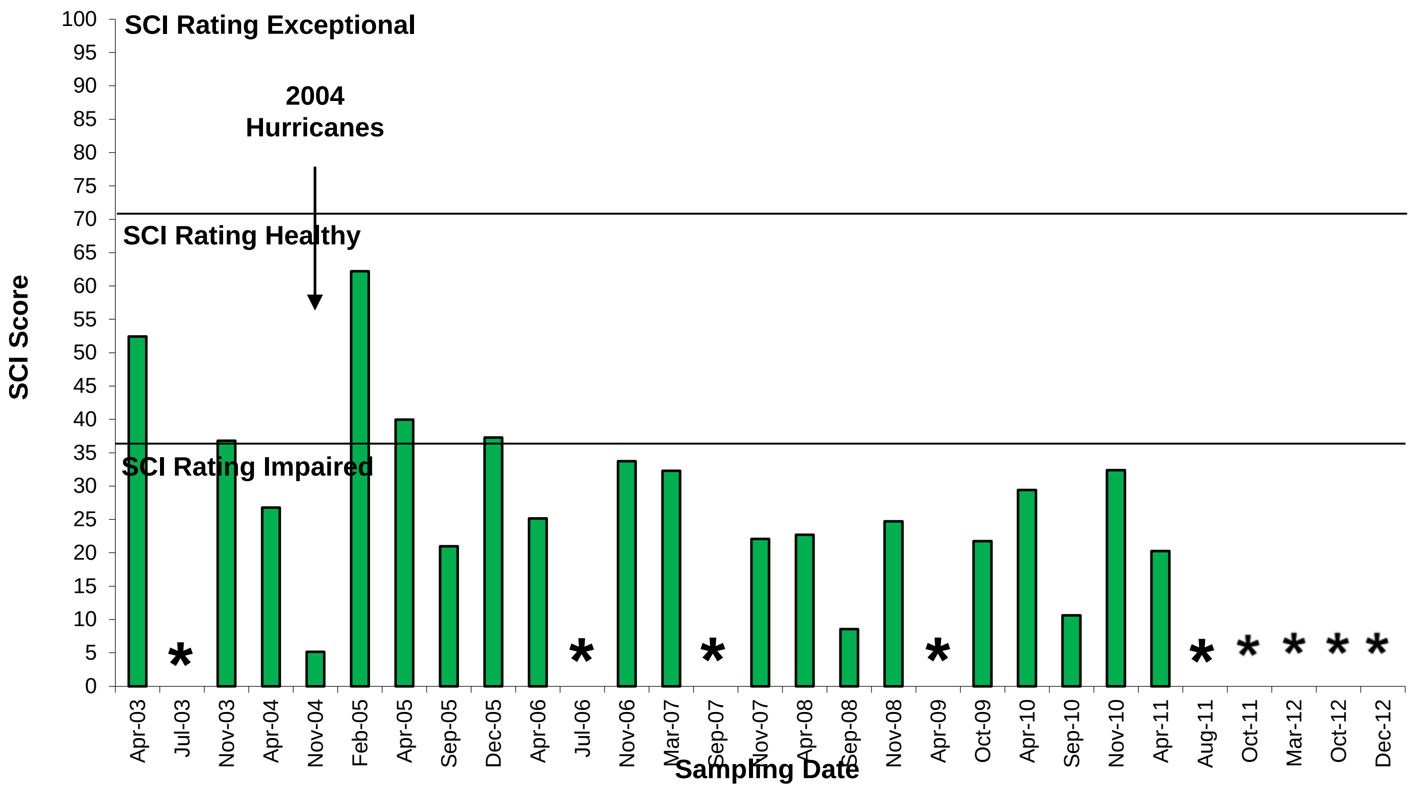

Figure 85. SCI Scores for Samples Collected at HCSW-2, 2003 - 2012.

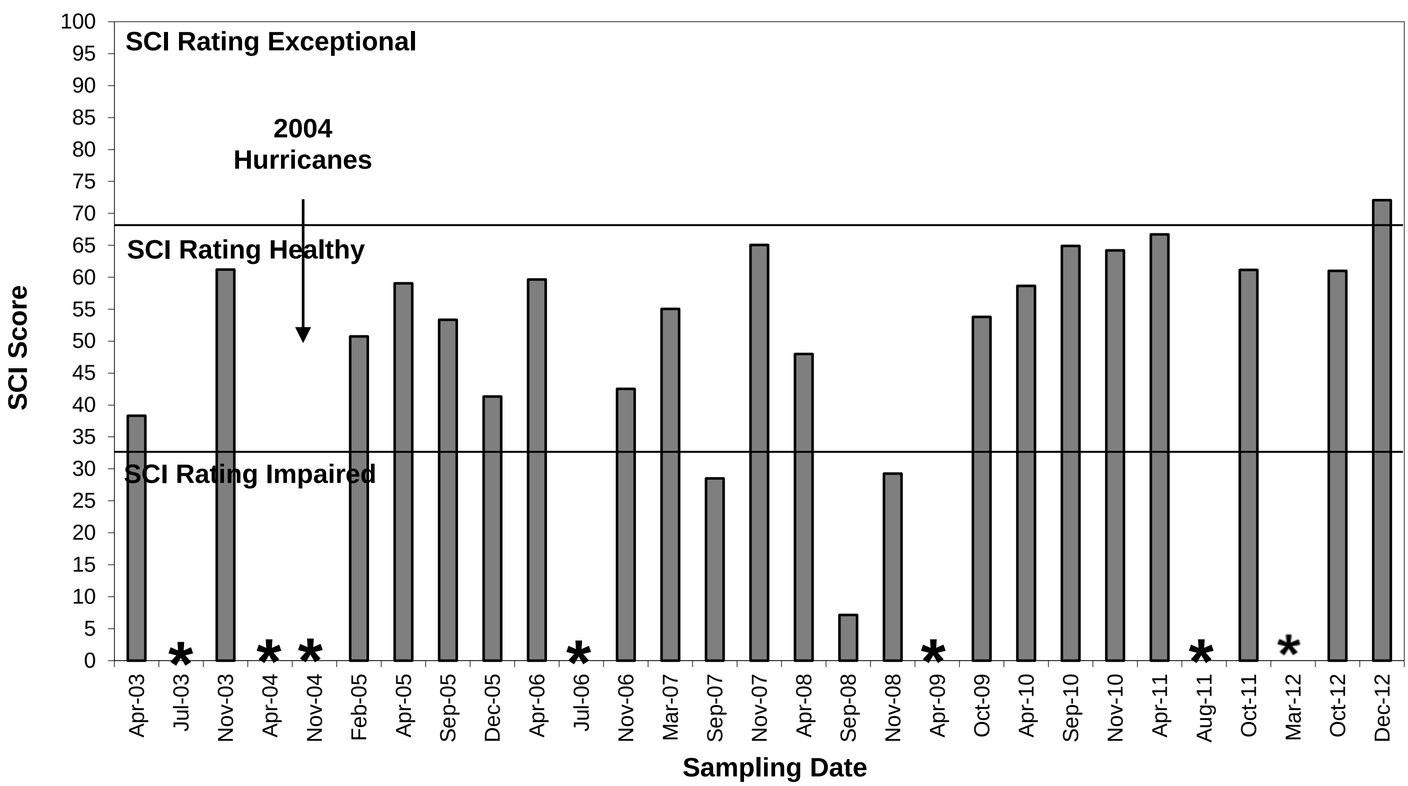

Figure 86. SCI Scores for Samples Collected at HCSW-3, 2003 - 2012.

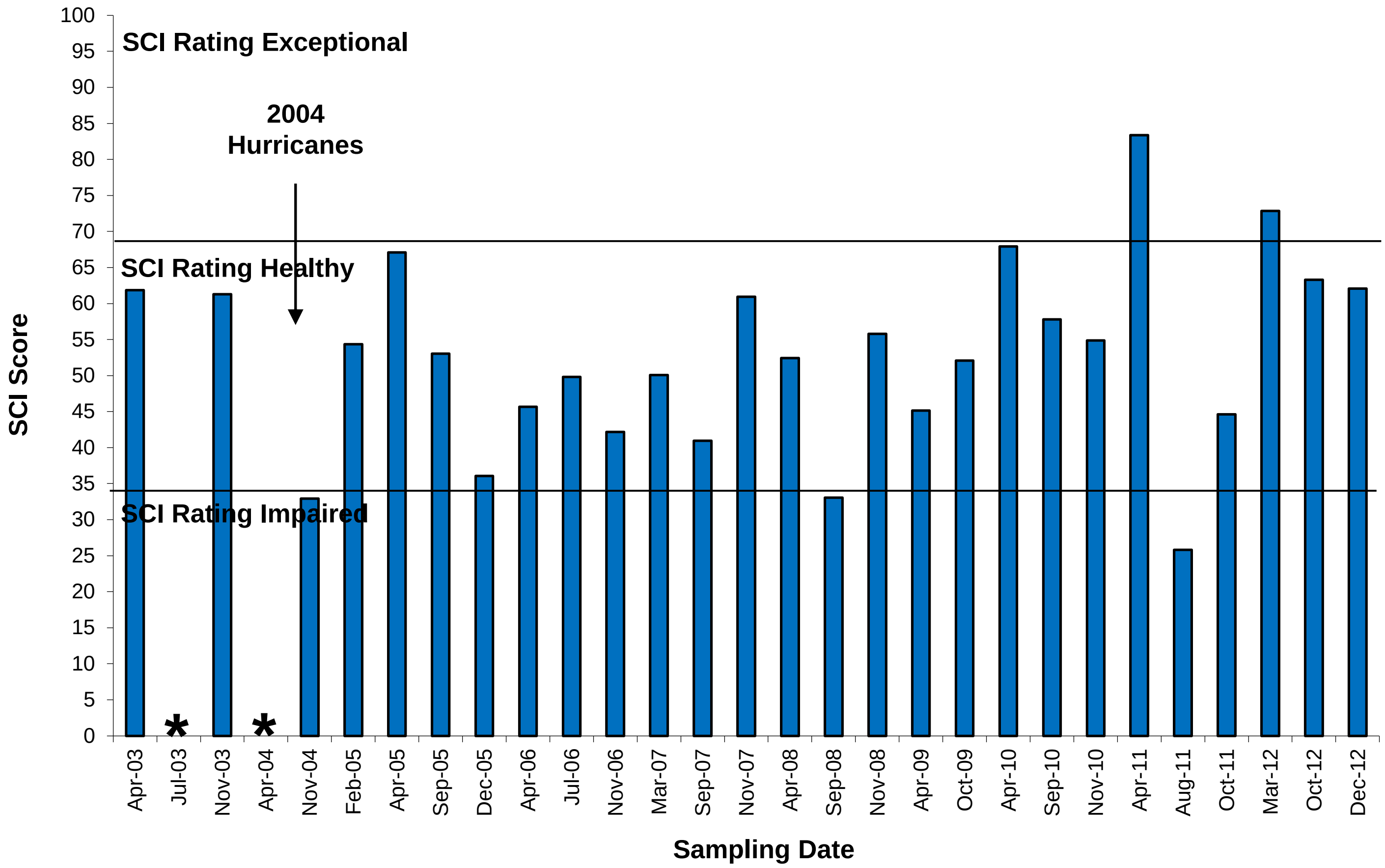

Figure 87. SCI Scores for Samples Collected at HCSW-4, 2003 - 2012.

5.3.3 Shannon-Wiener Diversity Index

Although not a component of the SCI protocol, the Shannon-Wiener Diversity Index is calculated for generic diversity for each benthic macroinvertebrate sampling event at each location. This index, one of the most popular measures of diversity, is based on information theory and is a measure of the degree of uncertainty in predicting what taxa would be drawn at random from a collection of taxa and individuals (Ludwig and Reynolds 1988). The Shannon-Wiener Index assumes that all taxa are represented in a sample and that the sample was obtained randomly:

$$H' = \sum_{i=1}^{S} (p_i)(\log_2 p_i)$$

where, H' = Information content of sample (bits/individual), index of taxa diversity,

S = Number of taxa, and

p_i = Proportion of total sample belonging to i^{th} taxa.

The Shannon-Wiener Index, H', increases with the number of taxa in the community and theoretically can reach very large values (Krebs 1998). In practice, however, H' does not generally exceed 5.0 for biological communities. The index is affected both by the number of taxa and their relative abundance; a greater number of taxa and a more even distribution of individuals across taxa both increase diversity as measured by H'. For example, consider two communities, each with 100 individuals of 10 taxa captured. Community A is dominated by one taxa (91 of 100 individuals), while only one individual was captured for each of the other nine taxa. Community B, however, is even, with 10 individuals captured for each of the ten taxa. While taxa richness is the same for both communities, the Shannon-Wiener Diversity Index shows that Community B is much more diverse than Community A (H' = 3.3 and 0.7, respectively), because Community A is dominated by only one taxa.

For the Horse Creek data, generic diversity[19], rather than species diversity, was used to account for the high variability of species present from year to year. In Horse Creek in 2012, the Shannon-Wiener Diversity Index ranged from 3.15 (October, HCSW-1) to 4.32 (December, HCSW-3, Figures 88-91[20]). When considered over time from 2003 to 2012, the diversity was variable at each station with a significant increase in diversity over time at HCSW-3 (annual median Kendall Tau = 0.56, $p < 0.05$). When stations and dates within years were combined, diversity was statistically different among years (ANOVA: $F = 2.37$, $p < 0.05$, Figure 92) with 2012 having higher diversity and 2004 having lower (Duncan's multiple range test, $p < 0.05$). When results from all events in 2003 - 2012 were combined by station (Figure 93), there was a significant difference between stations (ANOVA: $F = 3.64$, $p < 0.05$), where HCSW-4 was higher than all but HCSW-3, with HCSW-1, HCSW-2, and HCSW-3 being similar (Duncan's multiple range test, $p < 0.05$).

[19] After a conversation with Dr. John Epler (entomologist) about updates to the accuracy of the species identification of a few Tanytarsini spp. an overall review of the data was performed. Some of the taxonomic classifications of older data (prior to 2006) had changed, so the database had multiple names for the class, family, or genus of some individuals. Taxonomic names were updated and consolidated where appropriate, which changed the number of individual genera counted for each sampling event. The richness and diversity stats were rerun for each sampling event (2003-2012), along with the combined diversity measures for the year and sampling location. All graphs and tables represent the updated generic diversity scores after data review and consolidation.

[20] An asterisk (*) represents a sampling event that did not meet the required SOP. See Appendix J for more details.

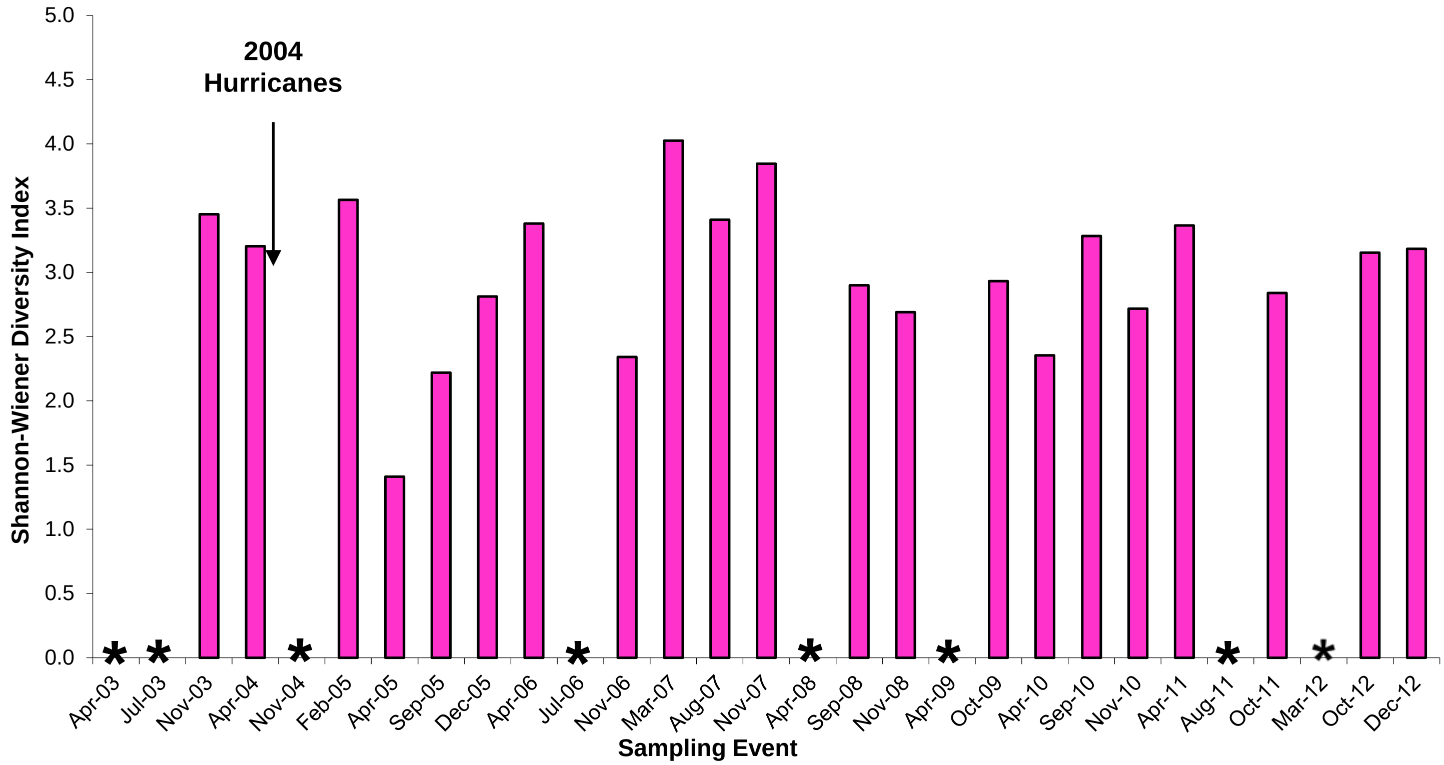

Figure 88. **Shannon-Wiener Diversity Indices for Benthic Macroinvertebrate Genera from HCSW-1 on Horse Creek from 2003 - 2012.**

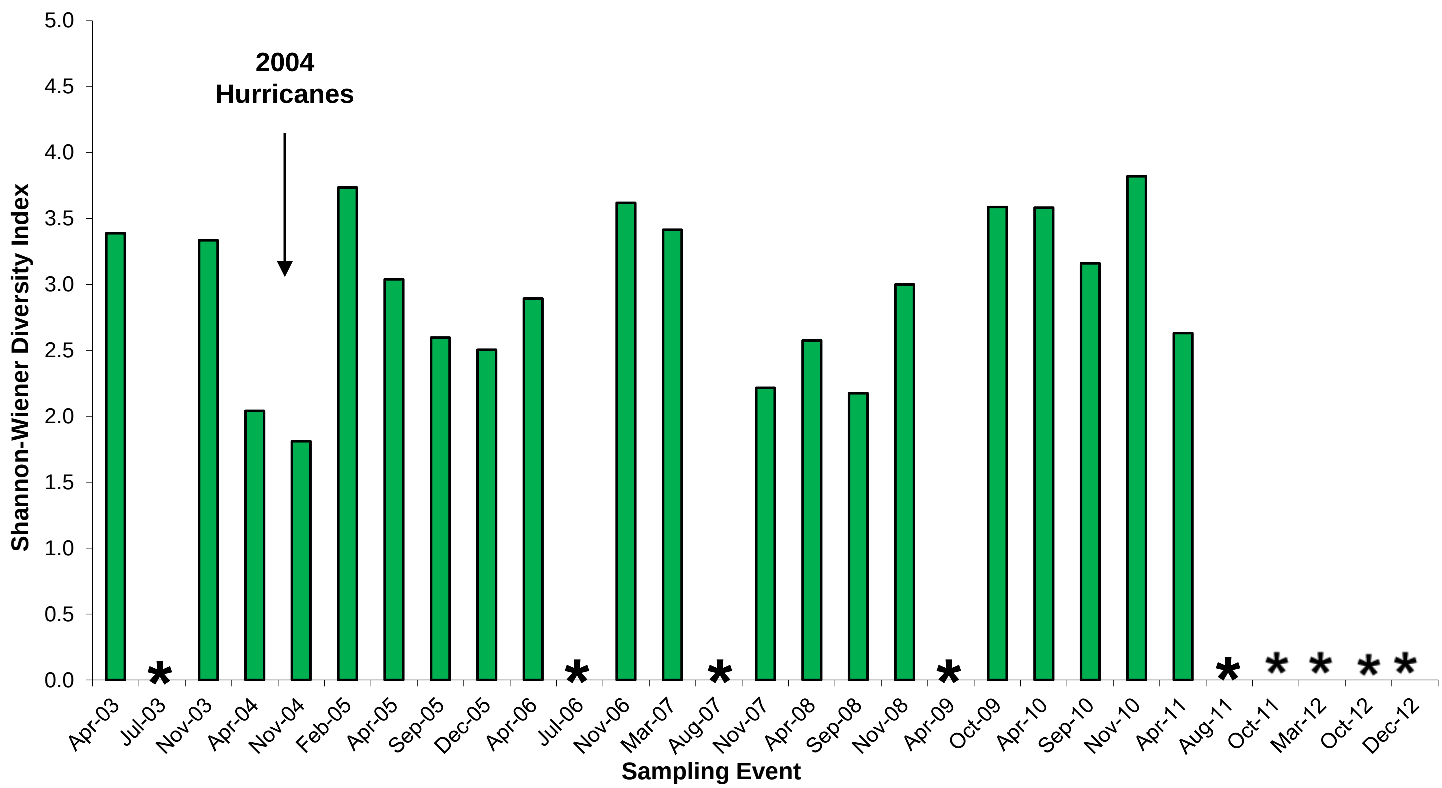

Figure 89. **Shannon-Wiener Diversity Indices for Benthic Macroinvertebrate Genera from HCSW-2 on Horse Creek from 2003 - 2012.**

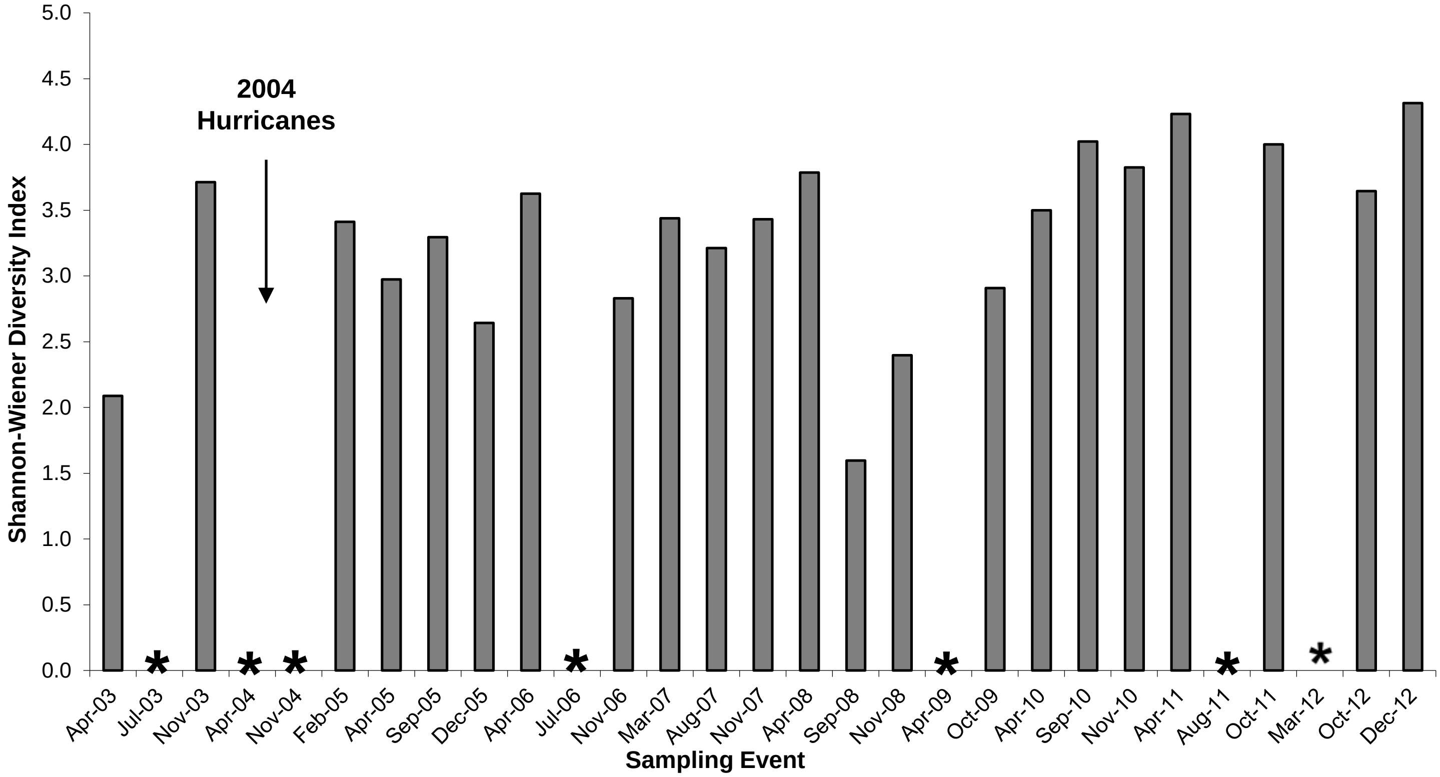

Figure 90. **Shannon-Wiener Diversity Indices for Benthic Macroinvertebrate Genera from HCSW-3 on Horse Creek from 2003 - 2012.**

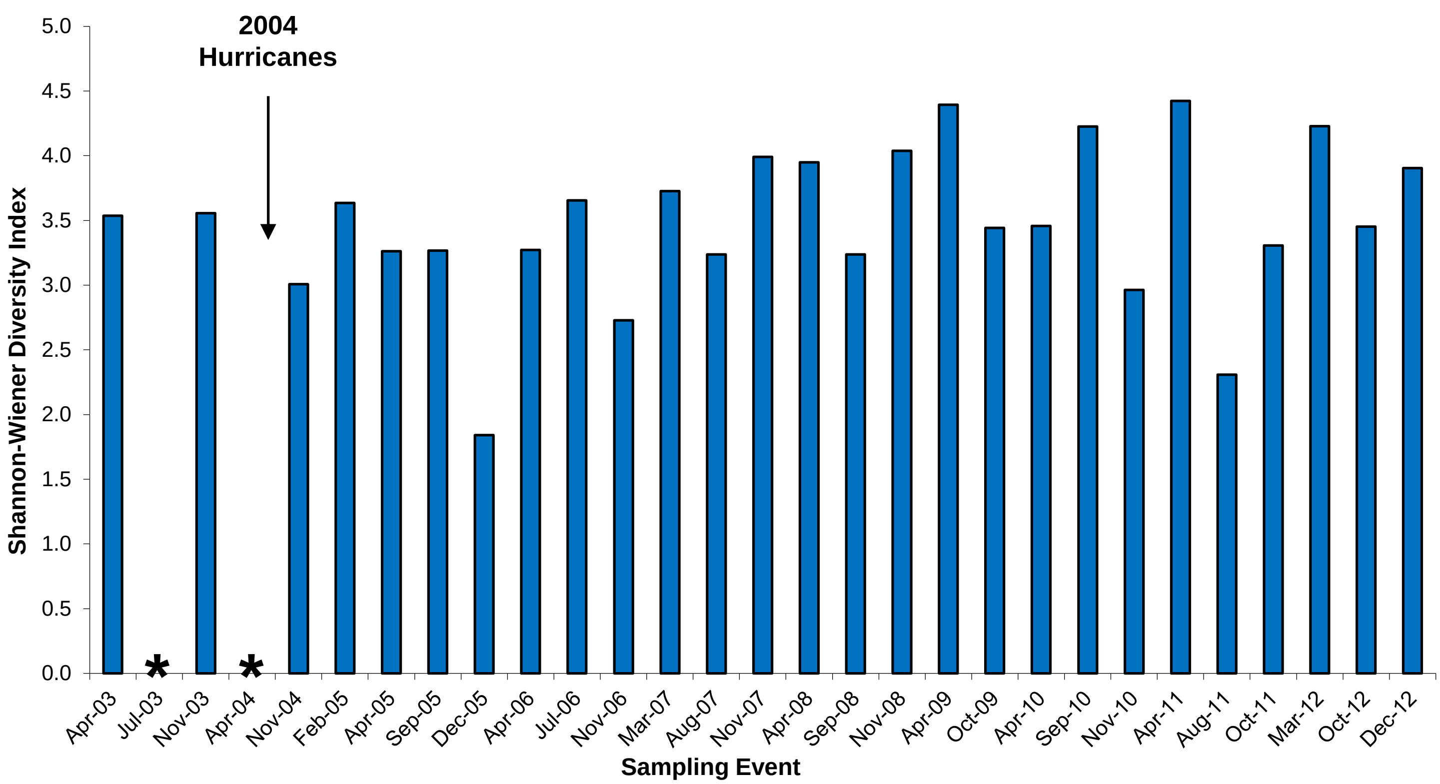

Figure 91. **Shannon-Wiener Diversity Indices for Benthic Macroinvertebrate Genera from HCSW-4 on Horse Creek from 2003 - 2012.**

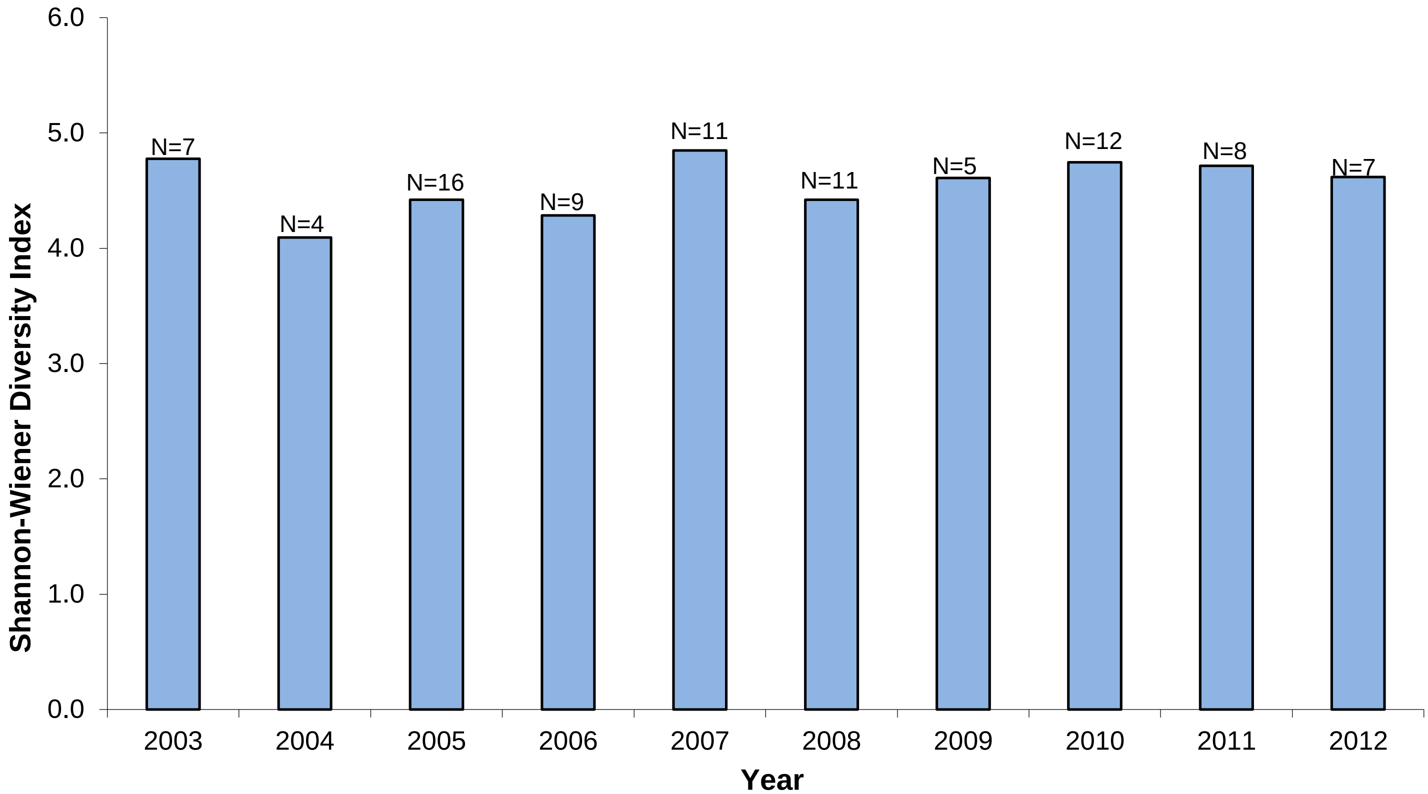

Figure 92. **Shannon-Wiener Diversity Indices for Benthic Macroinvertebrates Genera per year from Horse Creek for combined sample dates and stations.**

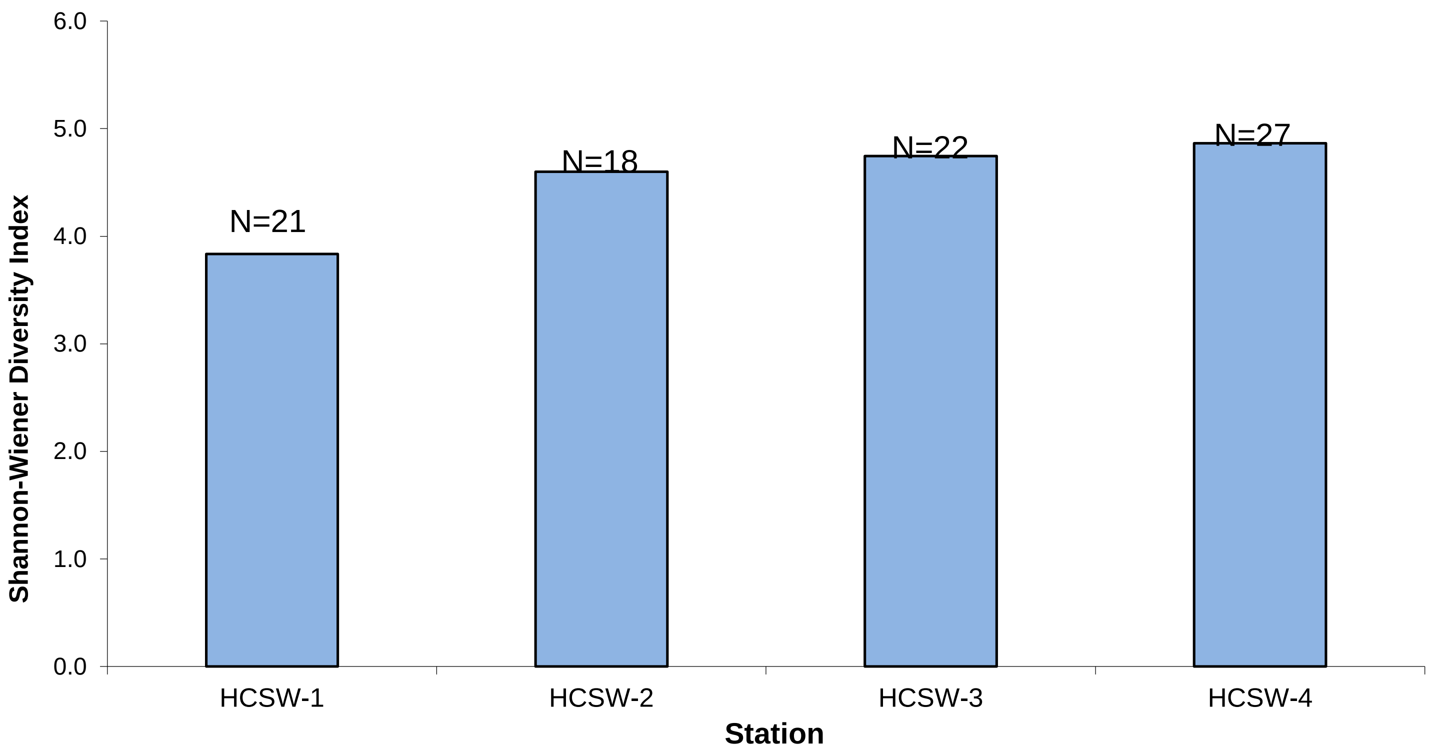

Figure 93. **Shannon-Wiener Diversity Indices for Benthic Macroinvertebrates Genera per Station on Horse Creek for combined sample dates.**

5.3.4 Summary of Benthic Macroinvertebrate Results

The brief discussion of each of the SCI parameters above conveys two important aspects of this particular ecological metric. First, there can be a large degree of variability among stations and among samples from the same station for a given calculated metric. Second, the actual range over which many of the measured parameters fluctuates can be very small, particularly for the parameters relying on integer counts of taxa (e.g., Ephemeroptera taxa generally range between 0 and about 4 across the various stream types evaluated in developing the SCI). These considerations suggest that care should be exercised in using any individual metric of the SCI as a separate indicator of stream habitat quality. This is the justification for combining all the parameters into a composite index that presumably has a stronger correlation to stream conditions than the separate metrics themselves.

The general quality of the macroinvertebrate community at the Horse Creek stations was within the range commonly observed by Cardno in similarly-sized natural streams in this region of Florida. Recent SCI scores at three of the four stations are consistently rated as Healthy, with station HCSW-2 previously having Impaired SCI scores because of unique, natural upstream conditions (HCSW-2 was not sampled in 2012 due to dry/low flow conditions or flooding). In 2012, SCI scores were Exceptional at HCSW-3 in December and HCSW-4 in March; scores were Healthy all other stations and events.

It may appear inconsistent that when the Habitat Assessment scores indicated optimal or sub-optimal conditions, the total SCI scores indicated that the benthic communities were Impaired. However, this is essentially a matter of semantics resulting from the assignment of qualitative categories under the two different assessment protocols (which were developed independently and not necessarily designed to provide matching qualitative assignments for a given location). Following the adoption of the revised SCI calculation procedure in 2007, FDEP found that the majority of the reference/background stations it had sampled fell into the Healthy category when calculated under the new SCI (R. Frydenbourg, pers. comm.). This indicates that the sampled segments of Horse Creek are considered healthy and thus comparable in quality (as determined via the SCI) to other reference streams in Florida.

Benthic invertebrate taxa diversity and SCI metrics in Horse Creek exhibit both seasonal and year-to-year variation. Overall, taxa diversity indices and SCI metrics show few monotonic trends over time and are very similar between sampling events and stations. However, HCSW-2 has slightly lower diversity and significantly lower SCI metrics than other stations. Habitat conditions at HCSW-2 are consistently poor, with lower streamflow, dissolved oxygen, and pH than other Horse Creek stations. The source of the poor conditions are related to the lower than average streamflow and rainfall of the previous few years and the presence of Horse Creek Prairie, the large marsh located upstream of the biological sampling station.

5.4 Fish

Fish sampling was conducted at HCSW-4 on 30 March 2012, at all stations but HCSW-2 on 26 October 2012, and all stations but HCSW-2 on 12 December 2012. No fish sampling occurred at HCSW-2 in in 2012 either due to dry/no flow (March and December) or flooded conditions (October). Fish sampling did not occur at HCSW-1 or HCSW-3 in March since no SCI was conducted due to not meeting SOP criteria (dry/no flow).

During 2012, 25 species of fish were collected from the four Horse Creek sampling stations; they are listed in Table 22 (the attached CD-ROM provides all data). In 2012, one new fish species was collected at HCSW-4, the Orinoco sailfin catfish (*Pterygoplichthys multiradiatus*). However, fifteen species of fish that had been collected in 2003-2011[21] were not collected in 2012.

Of the native species collected, most are quite common regionally, and none were unexpected for this portion of Florida. Catfishes, killifishes, shiners and sunfishes were the most commonly collected groups. Nine of the 41 species collected from 2003 to 2012 are not native to Florida: the walking catfish (*Clarias batrachus*), African jewelfish (*Hemichromis letourneuxi*), brown hoplo (*Hoplosternum littorale*), oriental weatherfish (*Misgurnus anguillicaudatus*), blue tilapia (*Oreochromis aureus*), vermiculated sailfin catfish[22] (*Pterygoplichthys disjunctivus*), Orinoco sailfin catfish, sailfin catfish[6] (*Pterygoplichthys pardalis*), and *Pterygoplichthys gibbiceps*.

5.4.1 Taxa Richness and Abundance

Most of the individuals collected at each sampling station consisted of eastern mosquitofish, sailfin molly, or coastal shiners (*Notropis petersoni*). This can generally be attributed to conditions that are conducive to seining for small species. Eastern mosquitofish were collected at all sampling stations during all the 2012 sampling events. Seminole killifish (*Fuldulus seminolis*), spotted sunfish (*Lepomis punctatus*), bluefin killifish (*Lucania* goodei), least killifish, coastal shiners, sailfin mollies, and hogchokers (*Trinectes maculatus*) were collected at all sampling stations the majority of the time in 2012. Small numbers (as few as one) of individual fish were collected for some of the species found in 2012 (Table 22). Fewer fish species were collected at HCSW-1 during two sampling events in 2012 than the other stations because of the unique characteristics of that sampling location (Table 22, Figures 94 - 97). In addition, water levels and streamflow were fairly high during biological sampling at all stations in October which led to HCSW-2 not being sampled; higher water levels did not allow for some habitats to be reached by our sampling equipment. Taxa richness showed no monotonic trend over time at any station (Kendall Tau of annual median, $p > 0.05$).

[21] HCSP fish samples have been periodically sent to the fish collection of Florida Museum of Natural History. Fish species identifications from the museum collection were used to update the HCSP database and all diversity and richness calculations.

[22] Previously identified in 2004 Annual Report as *Hypostomus plecostomus* (suckermouth catfish). Confirmation identification as *P. disjunctivus* by Florida Museum of Natural History (FLMNH).

Table 22. Fish Collected from Horse Creek during HCSP Sampling in 2012.

Scientific Name	Common Name	HCSW-1			HCSW-2			HCSW-3			HCSW-4		
		30 Mar. 2012	26 Oct. 2012	12 Dec. 2012	30 Mar. 2012	26 Oct. 2012	12 Dec. 2012	30 Mar. 2012	26 Oct. 2012	12 Dec. 2012	30 Mar. 2012	26 Oct. 2012	12 Dec. 2012
Hemichromis letourneuxi	African jewelfish*		3	2					73	5		6	12
Oreochromis aureus	Blue tilapia*								1				
Lucania goodei	Bluefin killifish		6	4					33		20	2	2
Lepomis macrochirus	Bluegill								3			2	1
Labidesthes sicculus	Brook silverside		1	29					44		30	17	15
Hoplosternum littorale	Brown hoplo*			1								1	
Notropis petersoni	Coastal shiner		74	38					16	10	196	22	81
Lepomis marginatus	Dollar sunfish			2								1	1
Gambusia holbrooki	Eastern mosquitofish		114	99					1561	130	158	393	221
Jordanella floridae	Flagfish								123		1		
Lepisosteus platyrhincus	Florida gar		3						1				
Fundulus chrysotus	Golden topminnow								1				
Trinectes maculatus	Hogchoker								14	2	3	7	
Heterandria formosa	Least killifish								29		7	14	1
Pterygoplichthys multiradiatus	Orinoco sailfin catfish *												1
Lepomis microlophus	Redear sunfish			1						1		1	
Poecilia latipinna	Sailfin molly								341	31	34	136	34
Fundulus seminolis	Seminole killifish								2		16	8	2
Lepomis punctatus	Spotted sunfish		2	2					16	8	13	16	33
Etheostoma fusiforme	Swamp darter										1		1
Noturus gyrinus	Tadpole madtom								1	1	1		
Notropis maculates	Taillight shiner								53			1	
Clarias batrachus	Walking catfish*								1				
Lepomis gulosus	Warmouth		1	1							1	2	3
Ameiurus natalis	Yellow bullhead		1										
	Total Taxa		**9**	**10**					**18**	**8**	**13**	**16**	**14**
	Total Individuals	**	**205**	**179**	**	**	**	**	**2,313**	**188**	**481**	**629**	**408**

* - Non-native species

** - Samples not collected due to unfavorable conditions (dry/low flow or flooded).

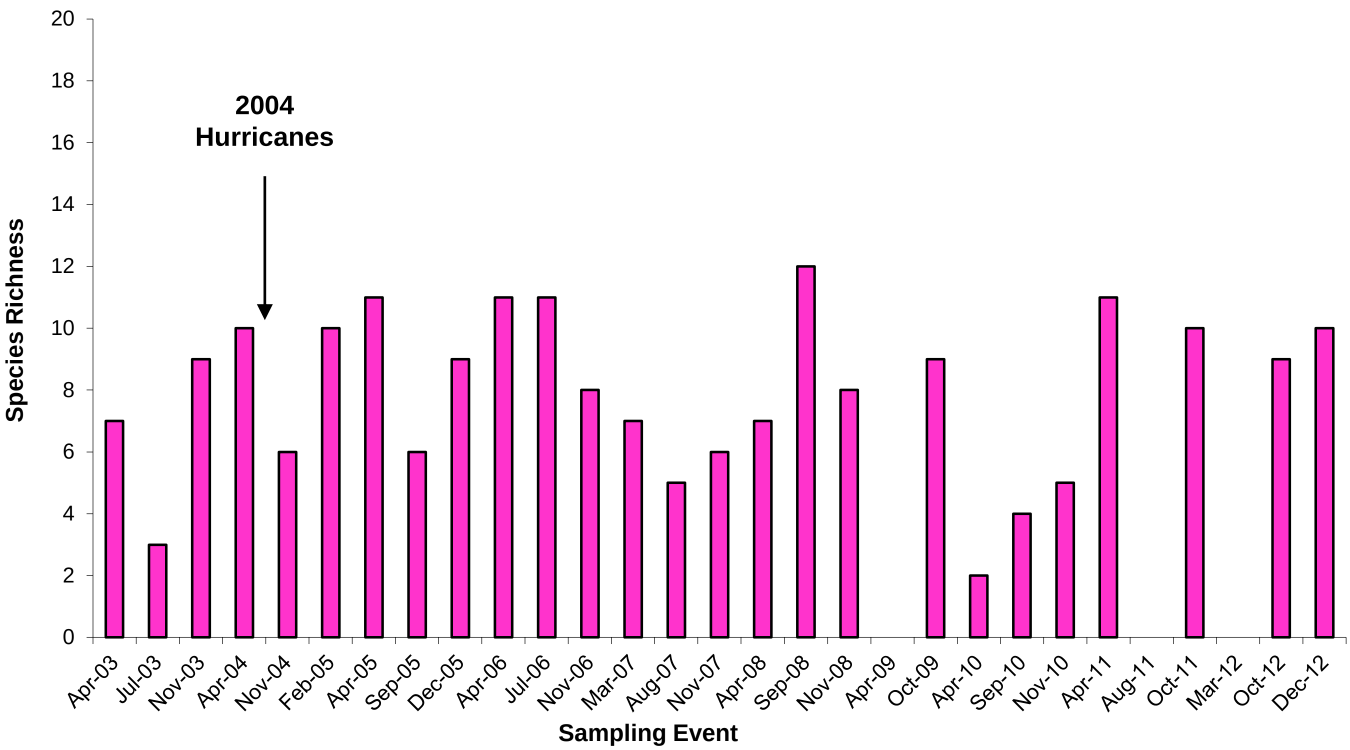

Figure 94. Species Richness for Fish from HCSW-1 on Horse Creek from 2003 - 2012.

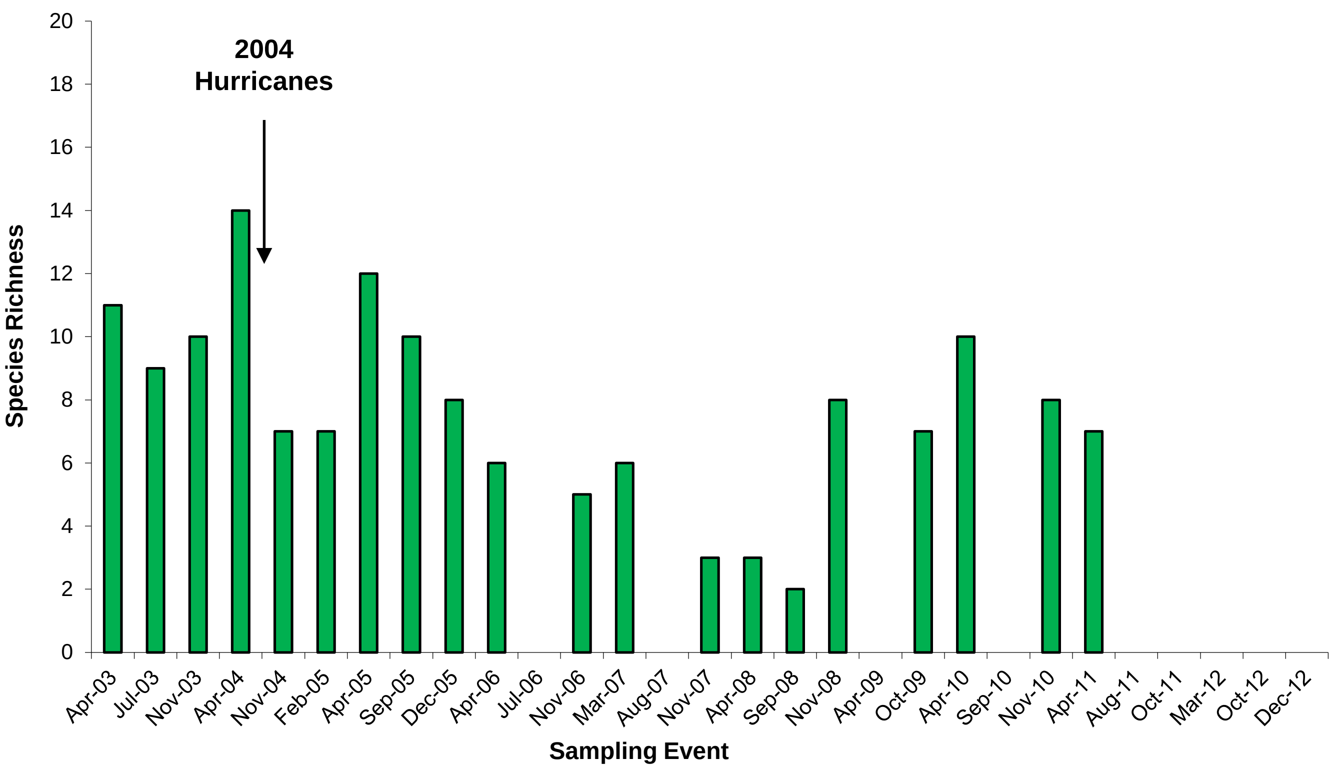

Figure 95. Species Richness for Fish from HCSW-2 on Horse Creek from 2003 - 2012.

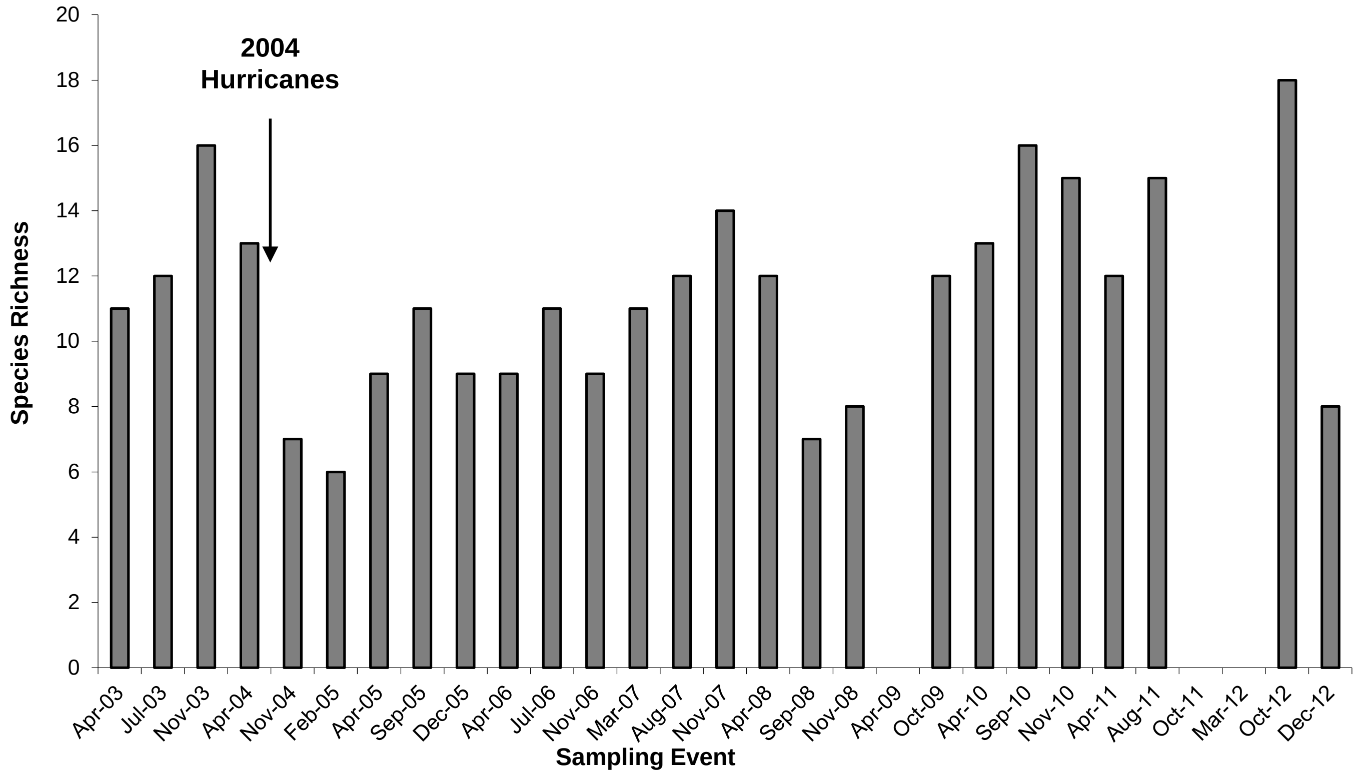

Figure 96. **Species Richness for Fish from HCSW-3 on Horse Creek from 2003 - 2012.**

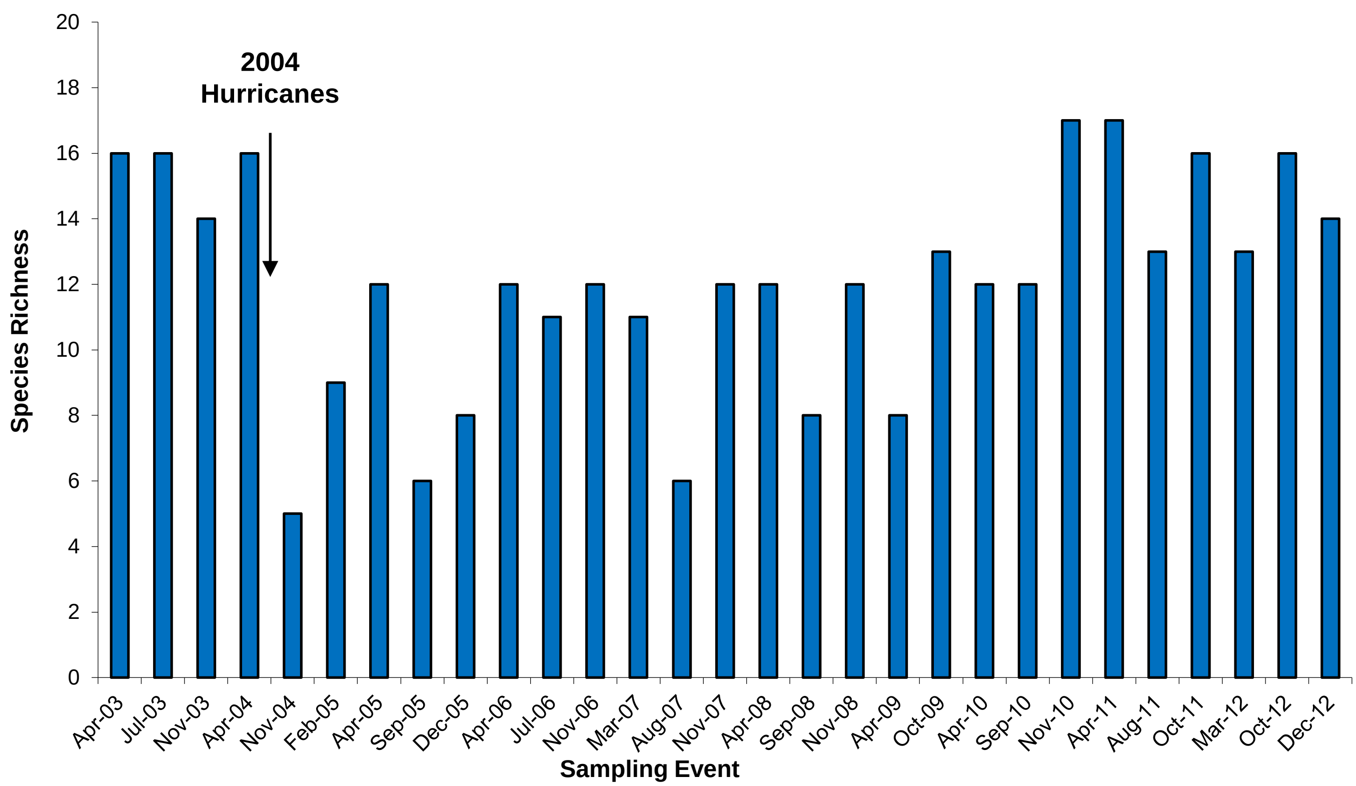

Figure 97. **Species Richness for Fish from HCSW-4 on Horse Creek from 2003 - 2012.**

5.4.2 Shannon-Wiener Diversity Index

Diversity of individual fish samples in 2012 ranged from 1.44 (HCSW-1, October and HCSW-3, December) to 2.24 (HCSW-4, March), similar to 2003 to 2011 ranges (Figures 98-101). When fish samples were combined across all sampling events within a year, HCSW-1 had the highest species diversity in 2004 – 2006 (after the hurricanes), but it had lower diversity in 2003 and 2010 than other stations (Figure 102). HCSW-4 had lower diversity in late 2004 and 2005 after the hurricanes and in 2010 and 2011 after abnormally cold winters. HCSW-3 followed the same pattern as HCSW-4 until 2008 and 2009; the lower diversity in late 2008 and 2009 may be related to difficulties in accessing fish habitats at this station when stream stage is high. The diversity for HCSW-3 was more average during 2010-2012. Fish diversity at HCSW-2 was lower in 2003-2009, because of changes in the amount of fish and fish habitat available for sampling, related to climate changes that affected flow and dissolved oxygen concentrations and physical changes to the stream segment where biological sampling occurs. Diversity increased during 2010 and 2011, but there were a limited number of sampling events per year (and no samples collected in 2012).

Diversity was significantly different between dates when stations were combined, with November 2010 having the lowest diversity score and August 2011 the highest (ANOVA $F = 1.75$, $p < 0.05$, Duncan's multiple range test: $p < 0.05$, Figure 103). Over all sampling dates combined (Figure 104), fish diversity was significantly lower at HCSW-2 than at the other stations (ANOVA $F = 7.33$, $p < 0.001$). Over all stations combined (Figure 105), fish diversity was significantly lower in 2010 than other years (ANOVA $F = 2.02$, $p < 0.05$) and has been slightly decreasing, but the differences were not significant (Kendall Tau of medians, $p > 0.05$).

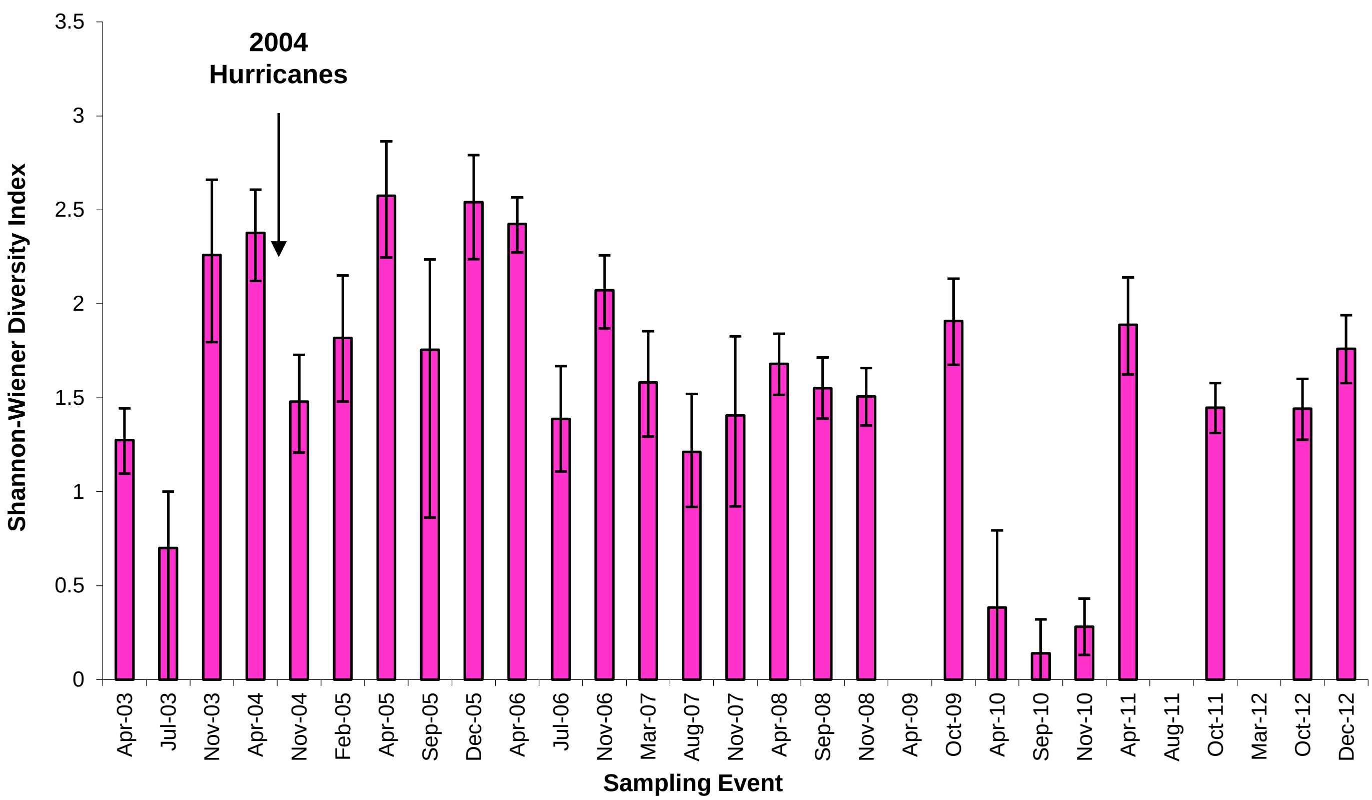

Figure 98. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from HCSW-1 on Horse Creek in 2003 – 2012.

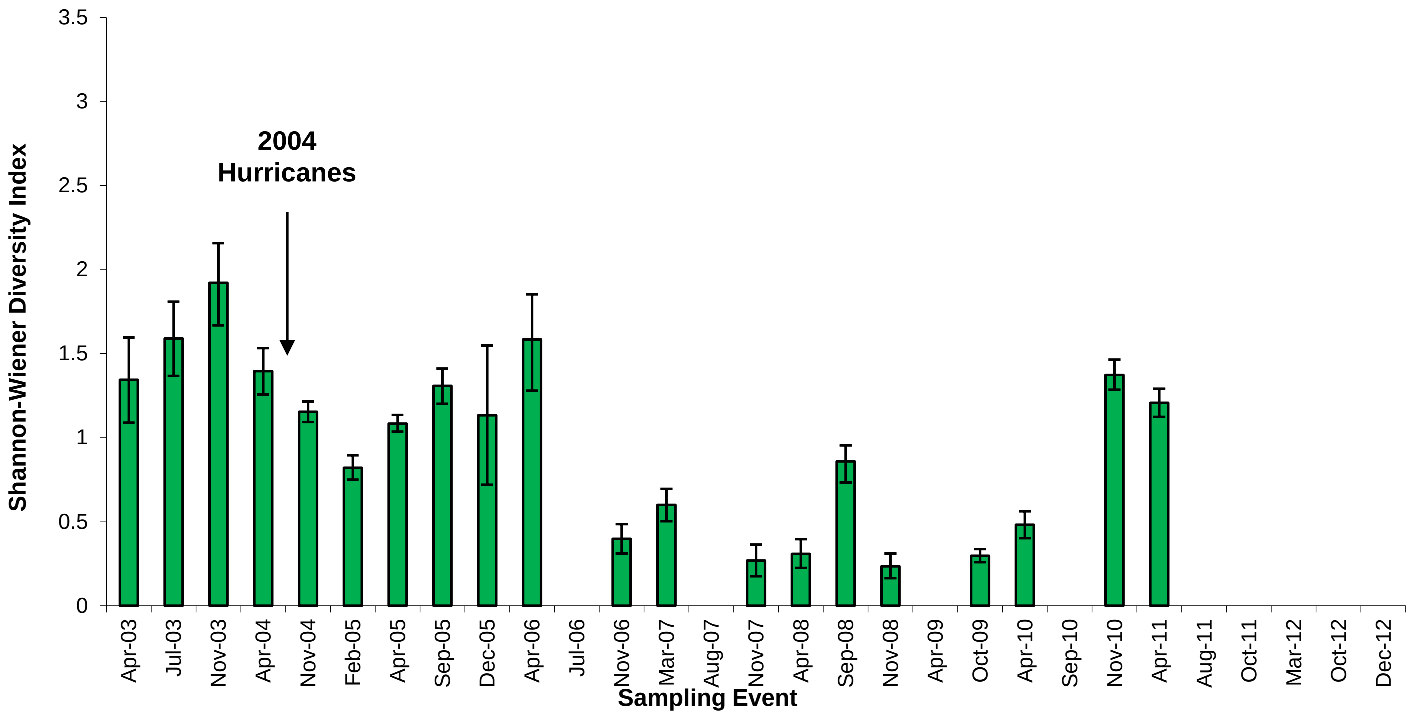

Figure 99. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from HCSW-2 on Horse Creek in 2003 – 2012.

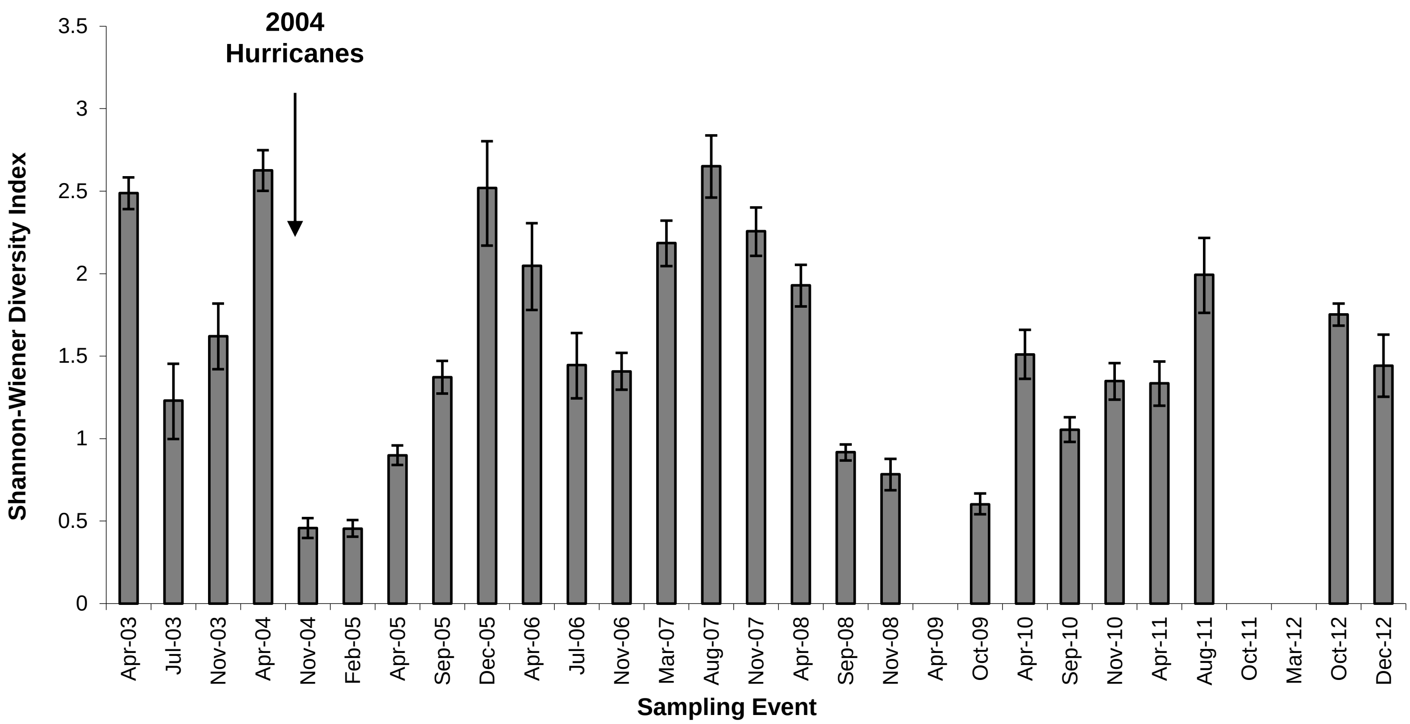

Figure 100. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from HCSW-3 on Horse Creek in 2003 – 2012.

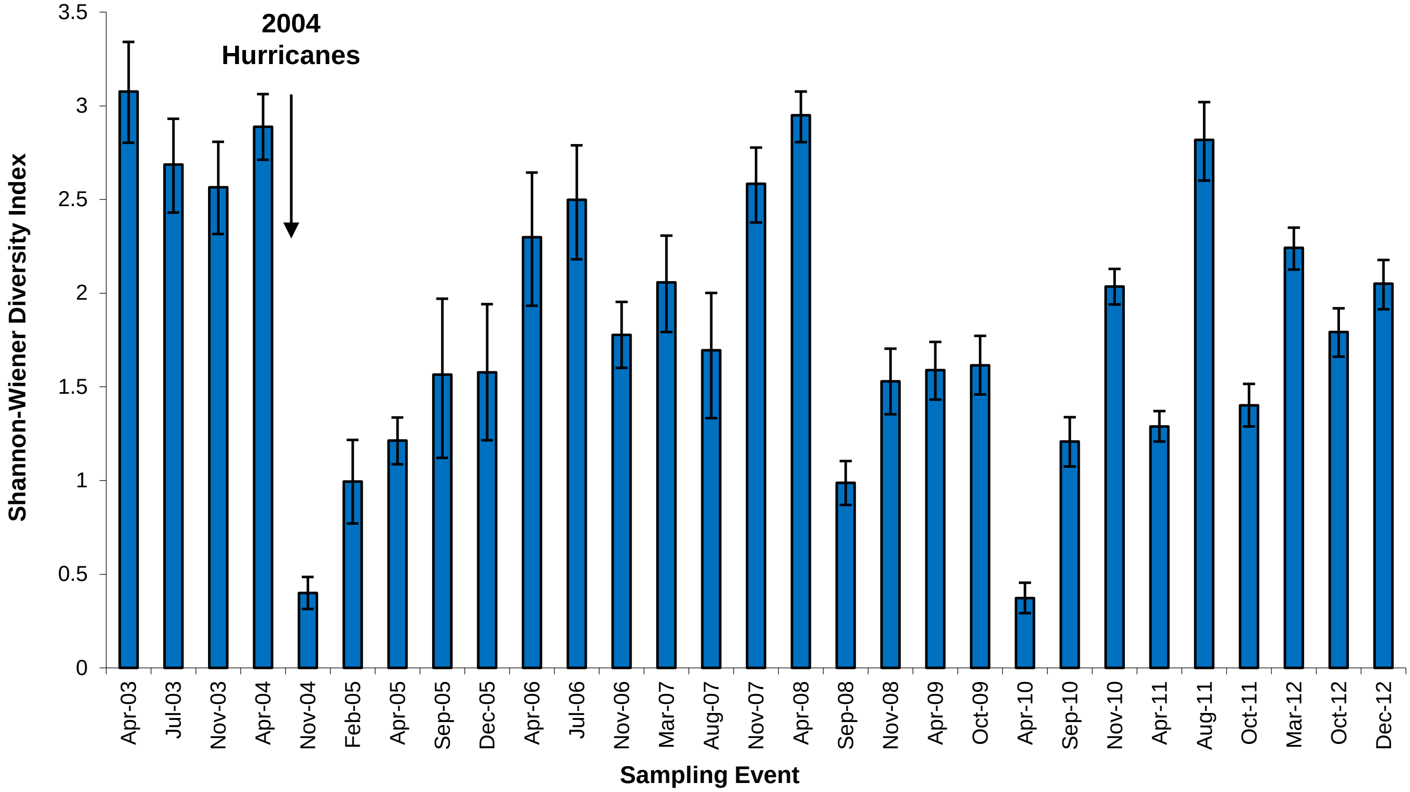

Figure 101. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from HCSW-4 on Horse Creek in 2003 – 2012.

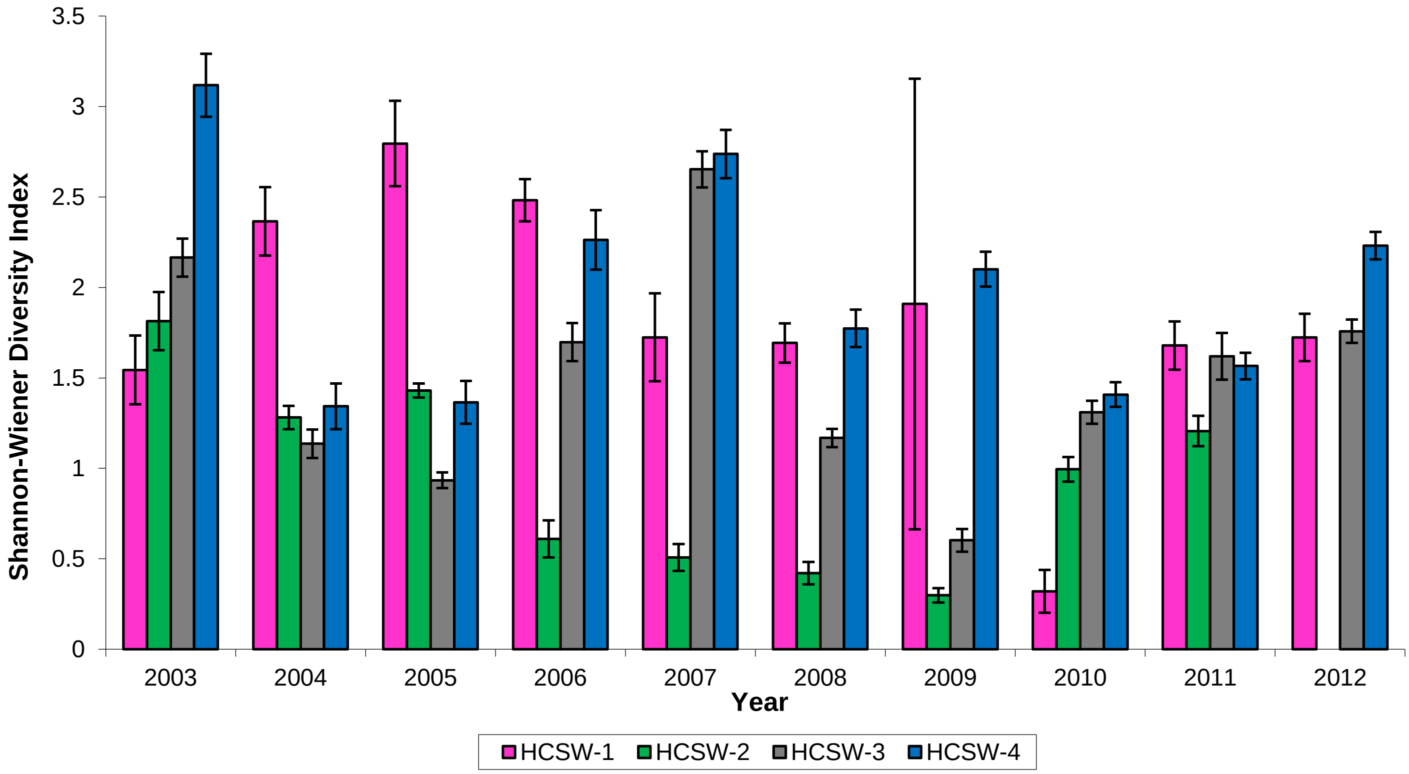

Figure 102. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from Four Stations on Horse Creek summarized over sampling events within each year.

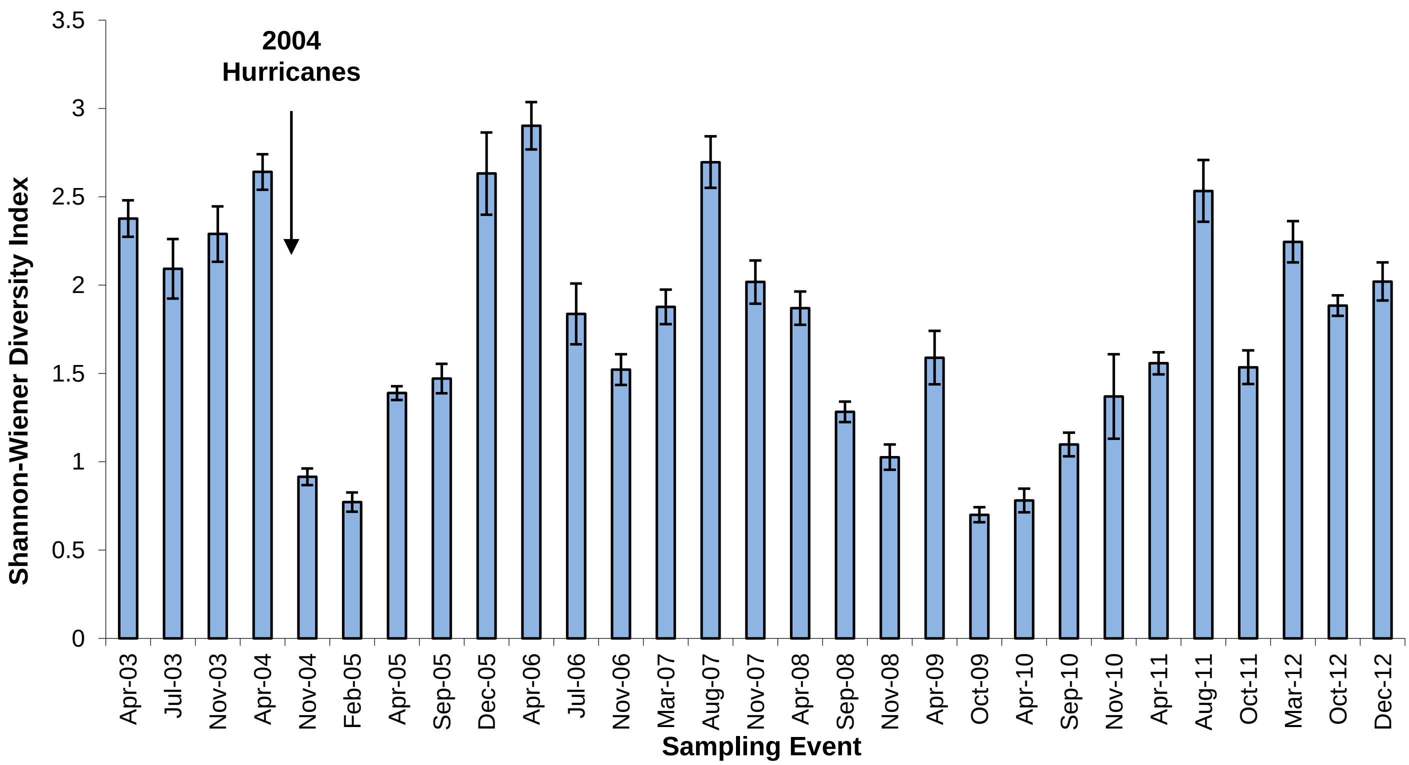

Figure 103. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from Horse Creek summarized over all stations per sampling event.

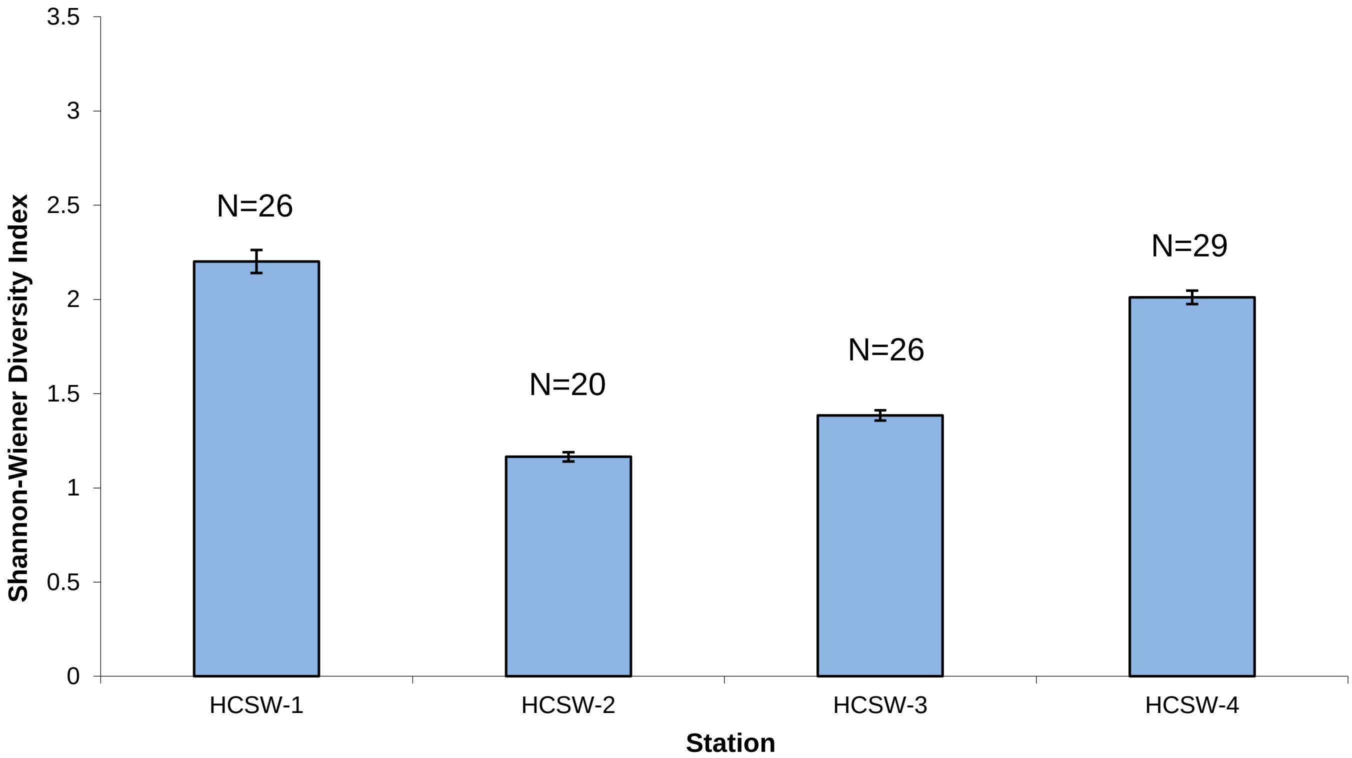

Figure 104. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from Four stations on Horse Creek summarized over all sampling dates.

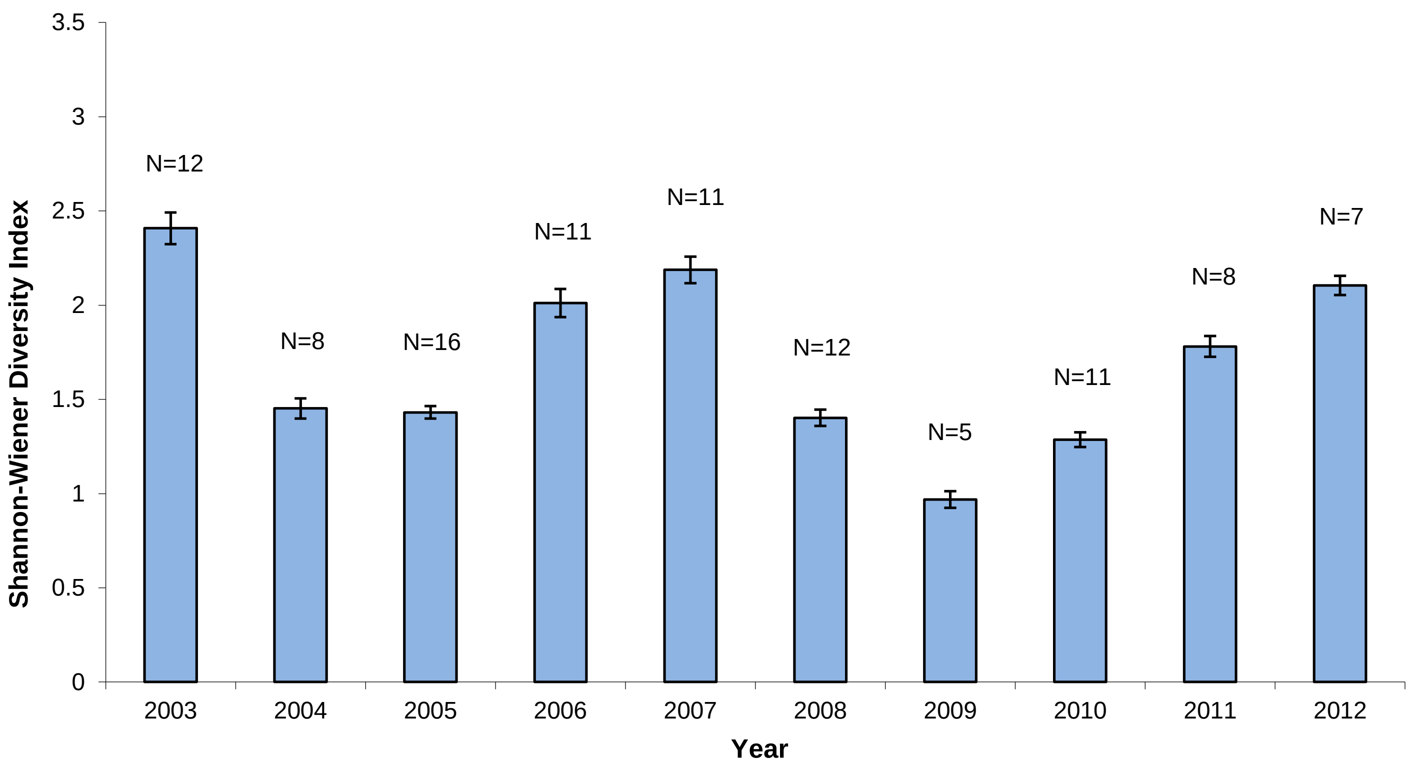

Figure 105. Shannon-Wiener Diversity Indices and 95% Confidence Limits for Fish Samples from Eight Years on Horse Creek summarized over all stations combined.

5.4.3 Morisita's Index of Similarity

Morisita's Index of Similarity measures the similarity of two communities by comparing the relative abundance of each species within and between communities. Of the similarity measures available, this index is preferred because it is nearly independent of sample size (Krebs 1998). Morisita's Index of Similarity is calculated as:

$$C_\lambda = \frac{2\sum X_{ij} X_{ik}}{(\lambda_1 + \lambda_2) N_j N_k}$$

Where C_λ = Morisita's index of similarity between sample *j* and *k*

X_{ij}, X_{ik} = Number of individuals of species *i* in sample *j* and sample *k*

N_j = Σ X_{ij} = Total number of individuals in sample *j*

N_k = Σ X_{ik} = Total number of individuals in sample *k*

Morisita's Index varies from 0 (no similarity – no species in common) to about 1 (complete similarity – all species in common) (Krebs 1998).

Table 22 includes Morisita's Index values combined by year or station. When all sampling locations for a given year or station are combined, fish communities were very similar (88%-100%, Table 23).

Table 23. Morisita's Similarity Index Matrix Comparing Sampling Dates within Stations or within Years for 2003 to 2012 Samples.

	HCSW-1	HCSW-2	HCSW-3	HCSW-4						
HCSW-1	1	0.88	0.90	0.95						
HCSW-2		1	0.98	0.97						
HCSW-3			1	0.99						
HCSW-4				1						
	2003	**2004**	**2005**	**2006**	**2007**	**2008**	**2009**	**2010**	**2011**	**2012**
2003	1	0.96	0.96	0.99	0.99	0.96	0.92	0.94	0.98	0.98
2004		1	0.98	0.99	0.95	1	0.99	1	0.99	0.95
2005			1	0.98	0.94	0.97	0.94	0.97	0.99	0.94
2006				1	0.99	0.99	0.97	0.98	0.99	0.98
2007					1	0.96	0.93	0.94	0.97	0.99
2008						1	0.99	1	0.99	0.95
2009							1	1	0.96	0.93
2010								1	0.98	0.95
2011									1	0.96
2012										1

5.4.4 Species Accumulation Curves

One way to determine when enough individuals in a community have been sampled to accurately estimate species diversity with some level of confidence is to plot the cumulative number of species collected through the sampling period. The result should be a curve that increases steeply at first when new species are continually being found, then gradually levels off when new species become very rare. The asymptote of the curve suggests the point at which additional sampling will provide no additional species. The total number of species in a community, as well as the number of rare species, strongly influences the sampling effort needed to offer some certainty that most species have been reported. As indicated by the curves plotted for each of the sampling locations, as well as that for all stations combined, we continue to collect very few new species with subsequent sampling events, and the curves shown have also leveled off for each station (Figure 106). This suggests that very few, if any additional species will be collected in the future (although one new species was observed in 2012, bringing the grand total to 41 species).

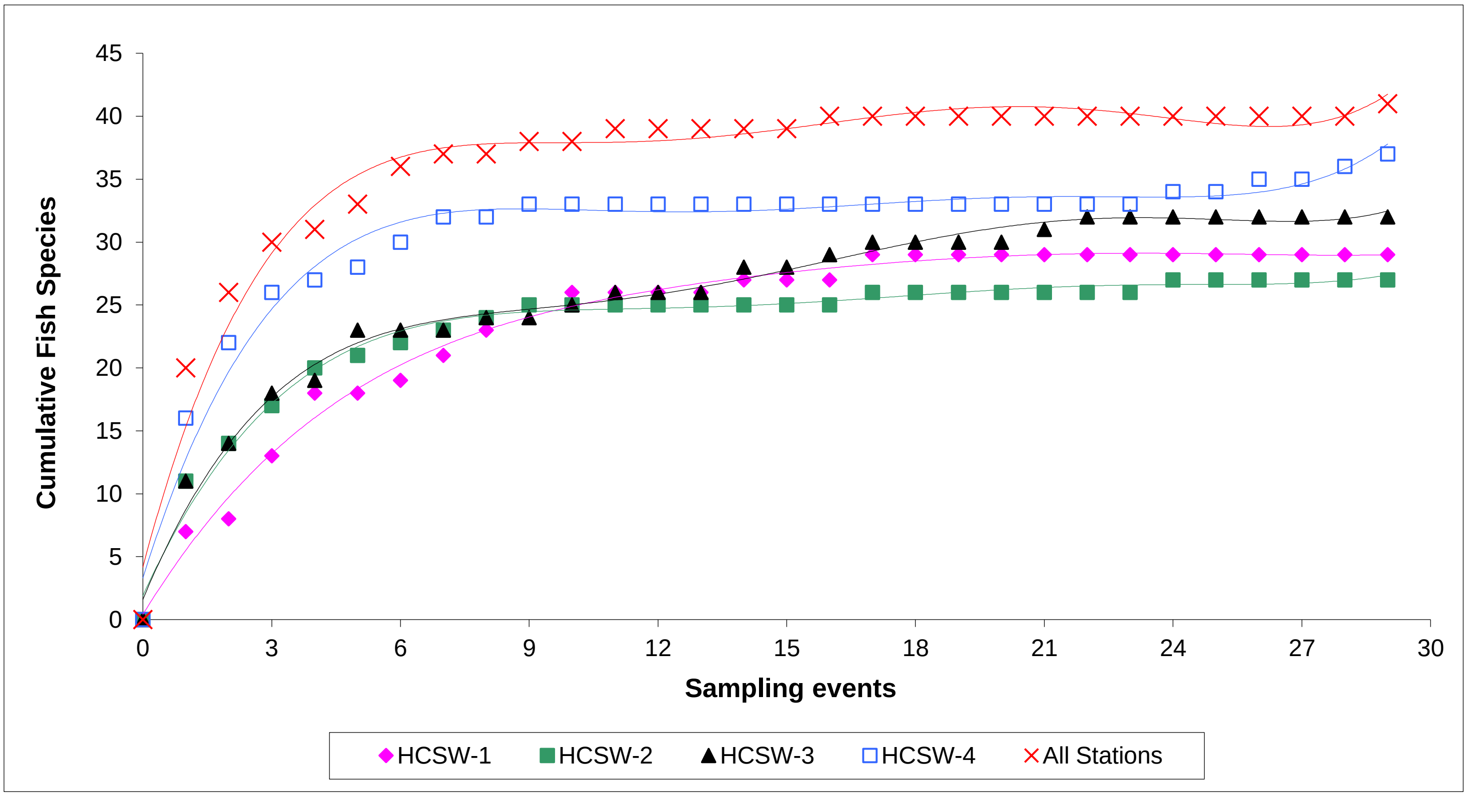

Figure 106. Cumulative Numbers of Fish Species Collected at Horse Creek Stations During 2003 -2012. (Species accumulation curves were fit for visual purposes only.)

5.4.5 Catch Per Unit Effort Analysis

Because of inconsistent sampling conditions during the biological sampling events, fish sampling effort may vary slightly between stations or sampling dates. To give a better representation in possible changes in sampling success over time, we have standardized the fish seining and electroshocking results to the number of individuals (Figures 107-110) and number of species (Figures 111-114) per 500 seconds of electroshocking or 5 seine hauls.

From Figures 107-110, it is clear that the standard number of individuals collected during seining is often greater than the number collected during electrofishing. This is expected because the seine technique is more likely to catch many small fish (such as mosquitofish or sailfin molly) in one haul; these small fish are often found in schools, which increase the likelihood of capturing a large group at one time. The figures also indicate that more fish are captured per unit effort at HCSW-2 and HCSW-3 than at other stations. At these two stations, the water levels are generally lower than at HCSW-4, leaving more habitats accessible for fish sampling; this is especially true for areas where seining small schooling species is most effective. Also, while the catch is higher at HCSW-2 it does not have the same diversity as the other locations. Eastern mosquitofish, which are common at this location and are usually caught in groups (hence the higher catch numbers), can survive in lower DO concentration waters. Other catfish or exotic species also tolerate low flow or low DO conditions. HCSW-1, which has the lowest catch per unit effort, is the narrowest station, which can lead to higher velocity and water levels during some sampling events that will interfere with fish sampling. In addition, HCSW-1 has less fish refuge habitat (areas isolated from main channel, snags, large roots, aquatic vegetation) than other stations, making it more difficult to sample fish without them being alerted. All stations show some variability over time, with a noticeable decrease in individuals per unit effort at HCSW-2 and HCSW-3 during the dry years of 2006-2007.

Fish richness per sampling effort is more similar between sampling methods (seine vs. shock), between stations, and over time (Figures 111-114). Richness per sampling effort is generally 12 species or less for each technique, station, and sampling event combination, with HCSW-1 and HCSW-2 on the lower end and HCSW-3 and HCSW-4 on the higher end of the spectrum. Given that the larger Peace River is a potential source of recruitment to the downstream stations of Horse Creek, this outcome is not surprising. In addition, HCSW-2 often has very low dissolved oxygen concentrations because of its proximity to a large wetland area; therefore, often only hypoxia-tolerant fish species are present at that station.

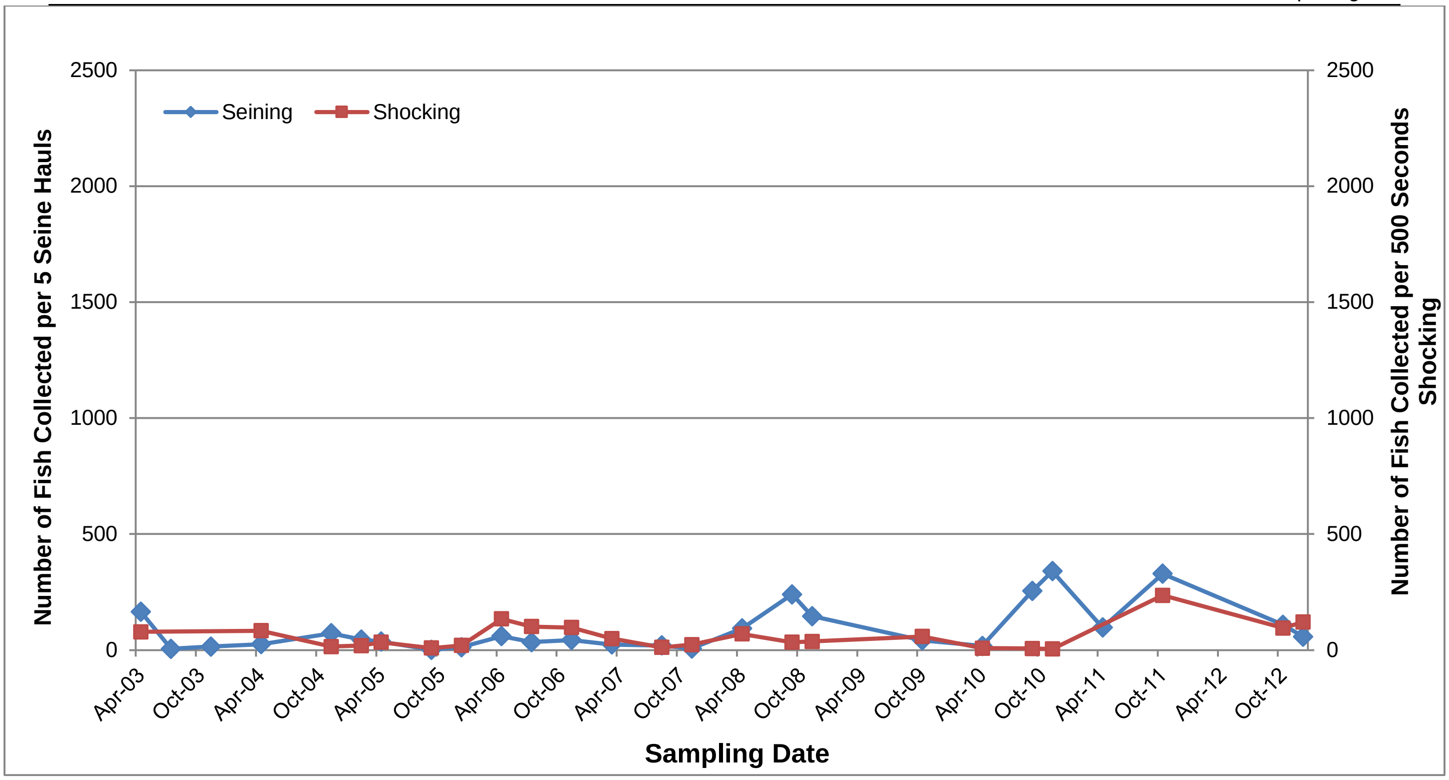

Figure 107. **Number of fish individuals per unit effort at HCSW-1 from 2003 – 2012 (total fish collected standardized to 5 seine hauls and 500 seconds shocking).**

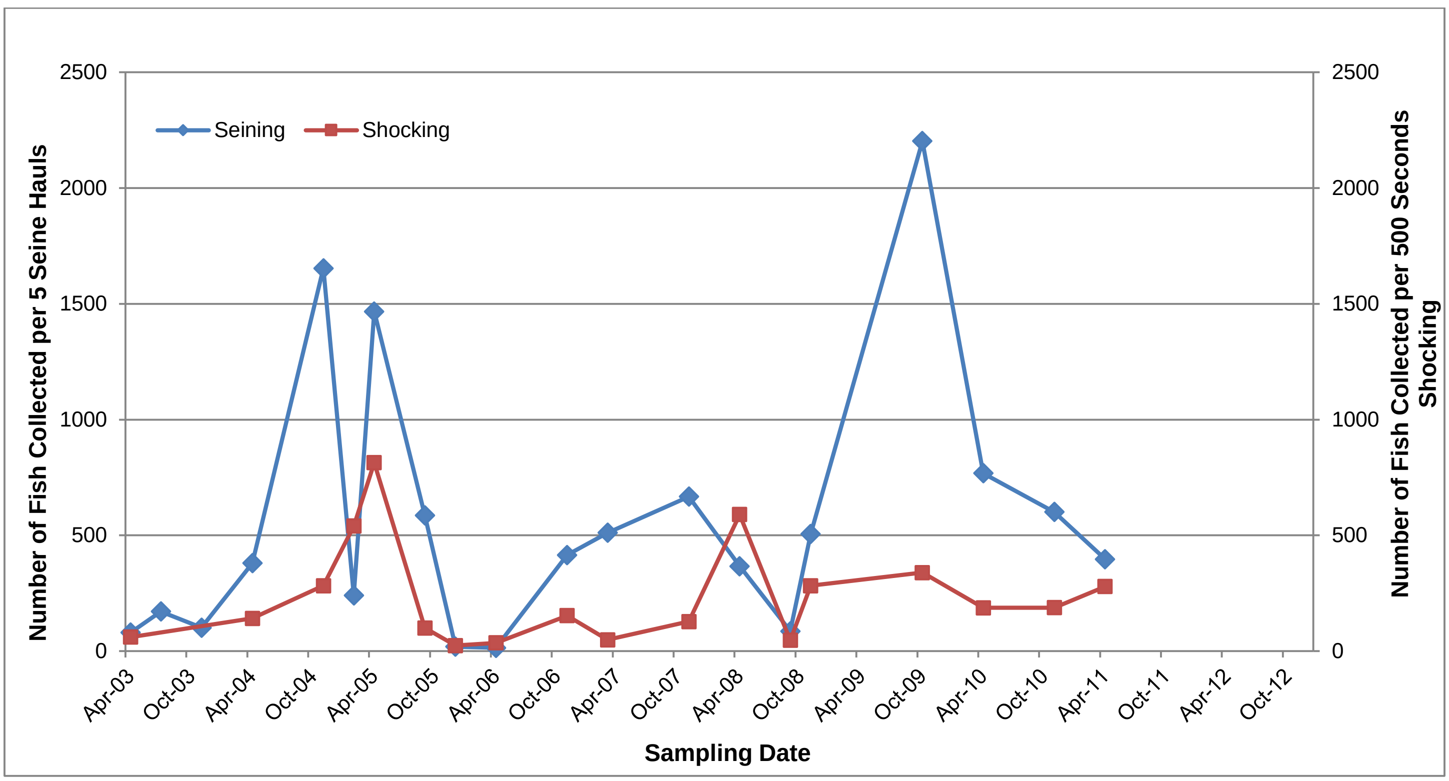

Figure 108. **Number of fish individuals per unit effort at HCSW-2 from 2003 – 2012 (total fish collected standardized to 5 seine hauls and 500 seconds shocking).**

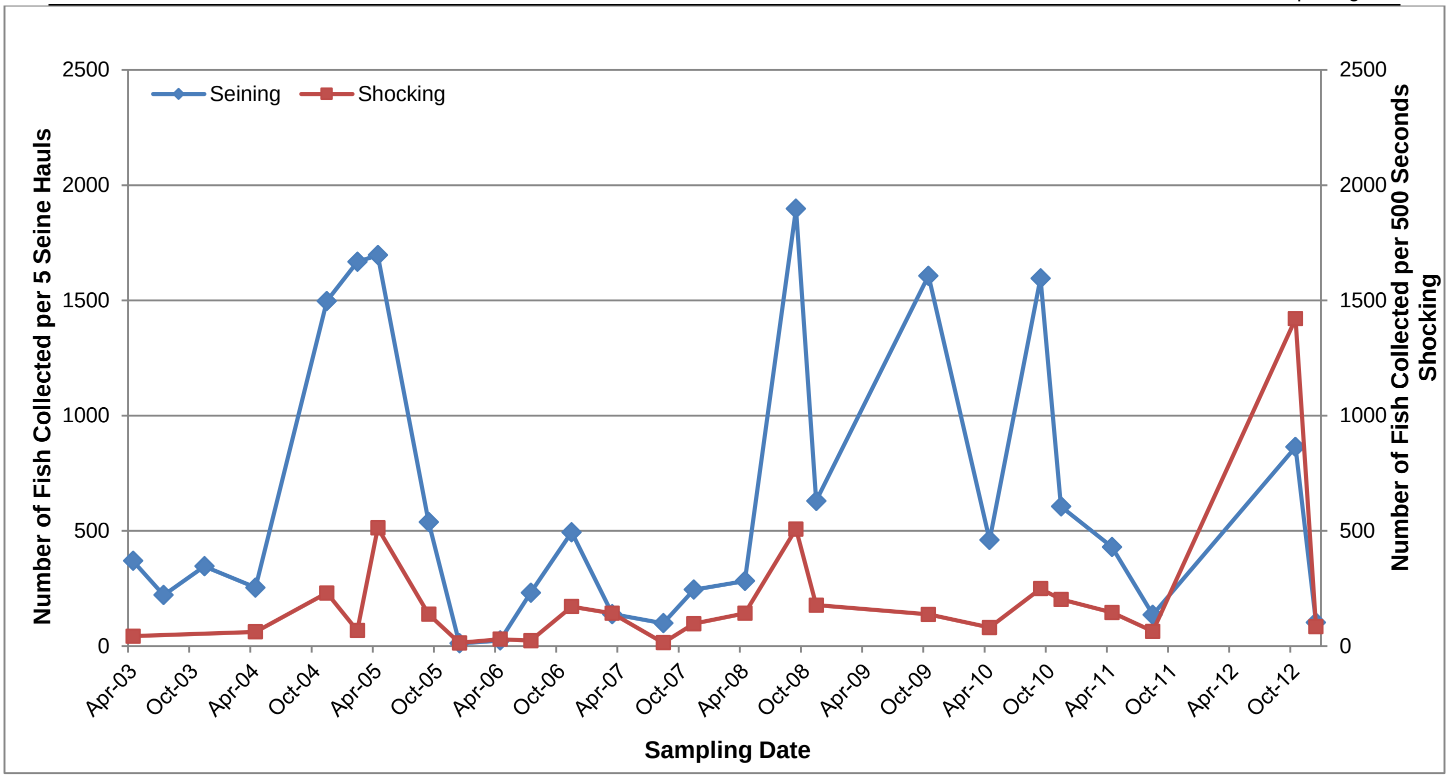

Figure 109. **Number of fish individuals per unit effort at HCSW-3 from 2003 – 2012 (total fish collected standardized to 5 seine hauls and 500 seconds shocking).**

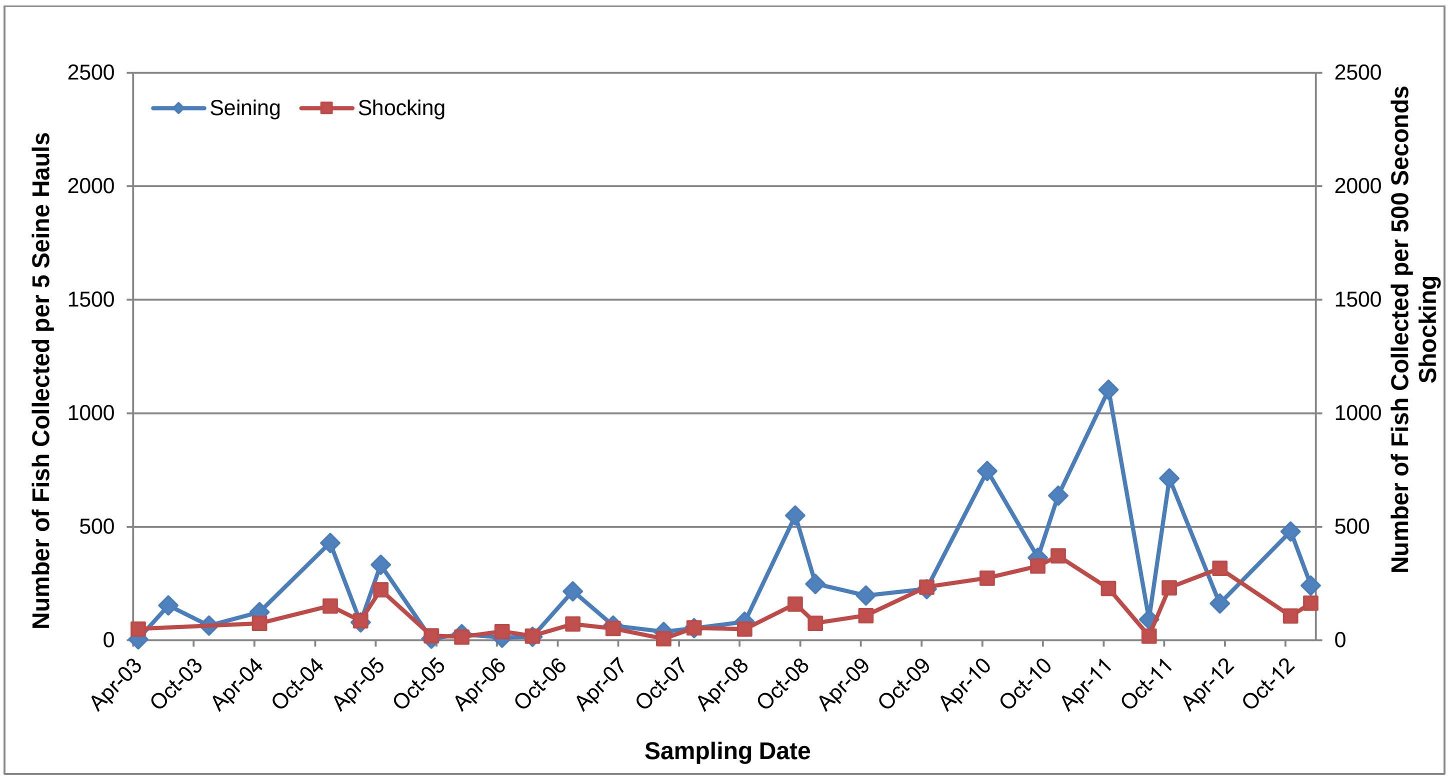

Figure 110. **Number of fish individuals per unit effort at HCSW-4 from 2003 – 2012 (total fish collected standardized to 5 seine hauls and 500 seconds shocking).**

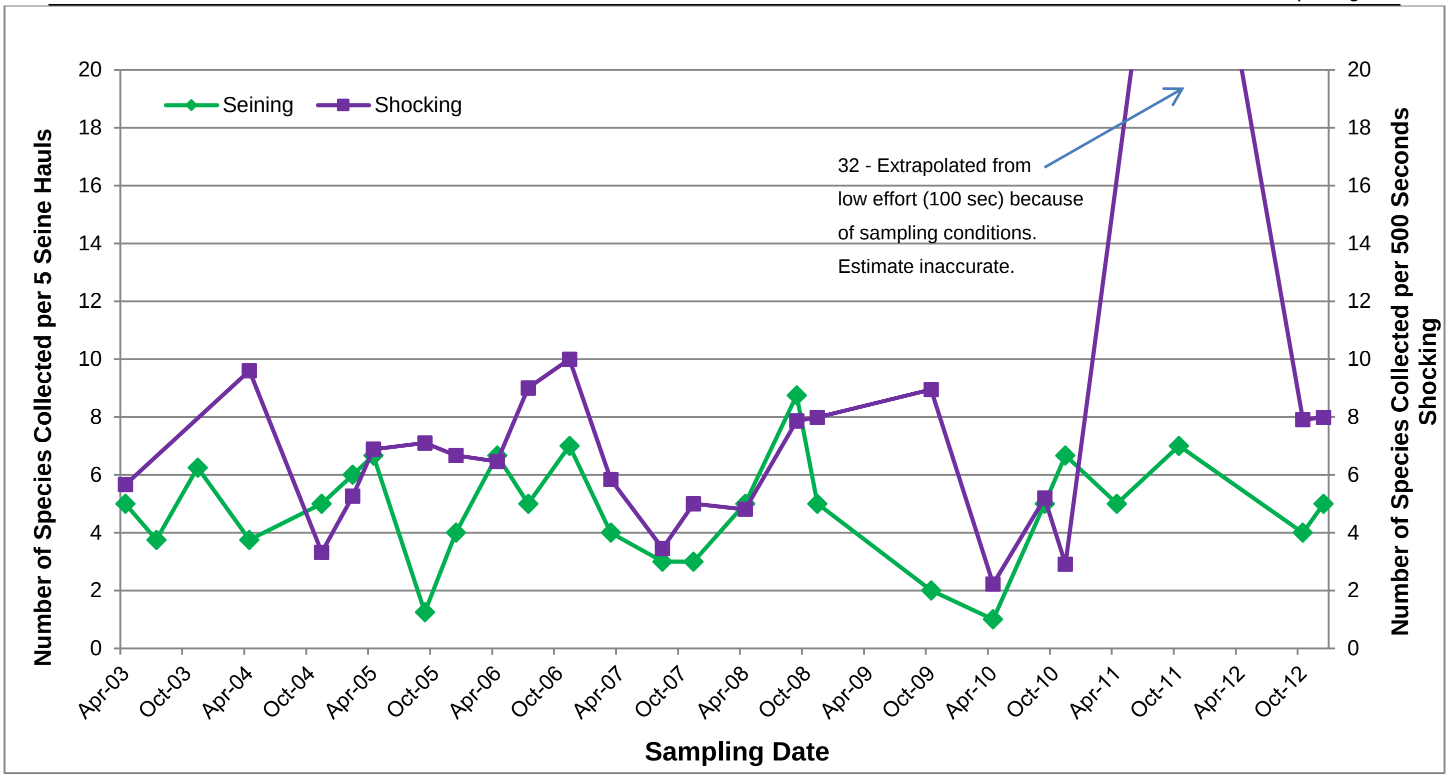

Figure 111. **Fishing richness per unit effort at HCSW-1 from 2003 – 2012 (fish species collected standardized to 5 seine hauls and 500 seconds shocking).**

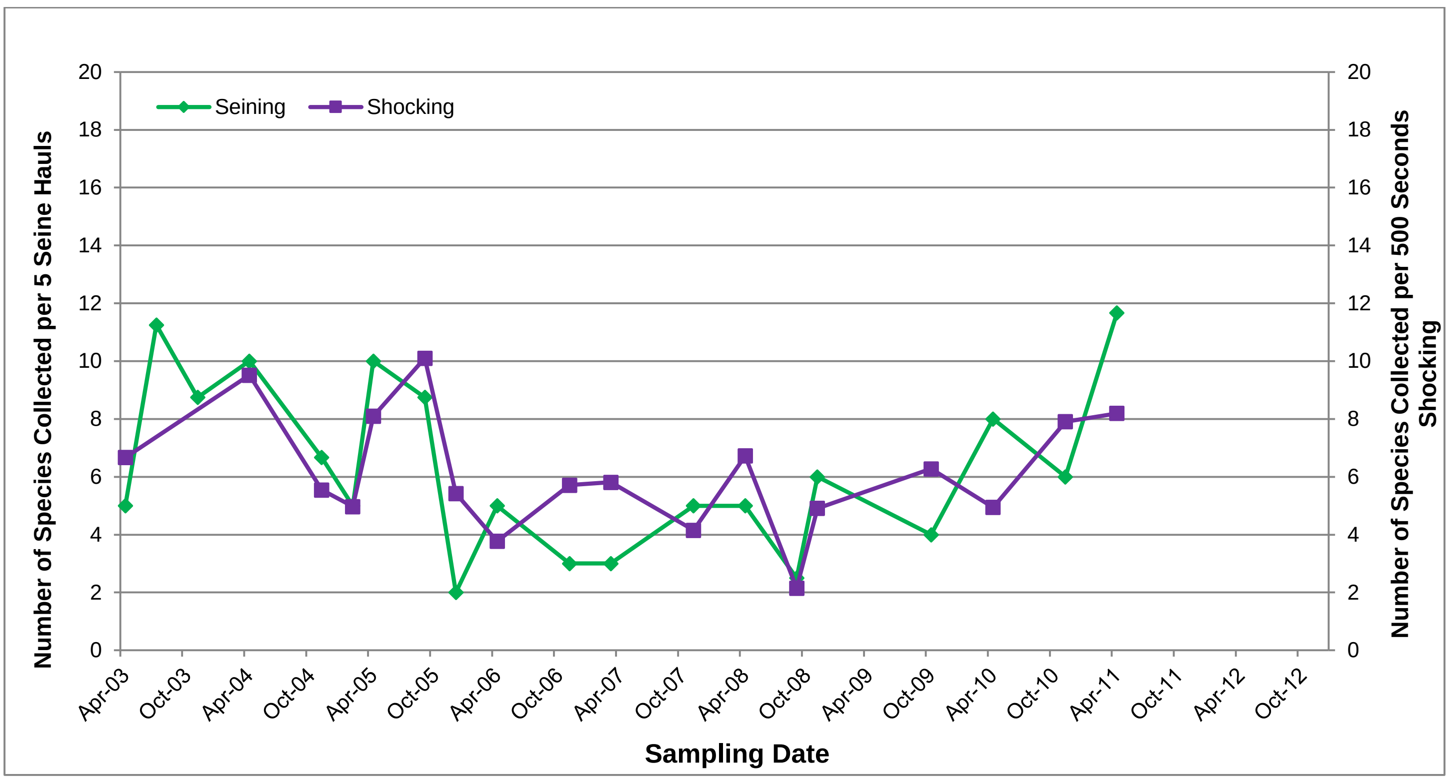

Figure 112. **Fishing richness per unit effort at HCSW-2 from 2003 – 2012 (fish species collected standardized to 5 seine hauls and 500 seconds shocking).**

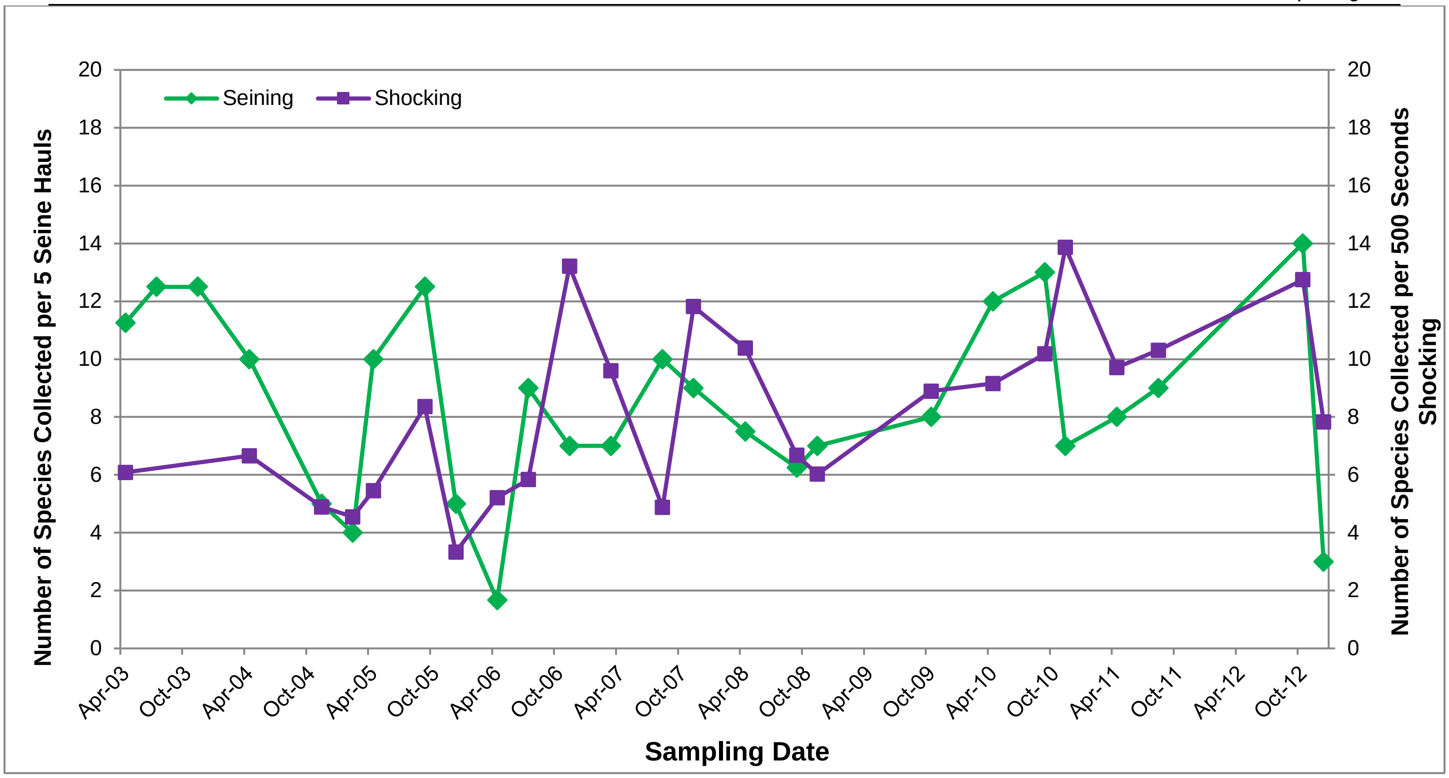

Figure 113. **Fishing richness per unit effort at HCSW-3 from 2003 – 2012 (fish species collected standardized to 5 seine hauls and 500 seconds shocking).**

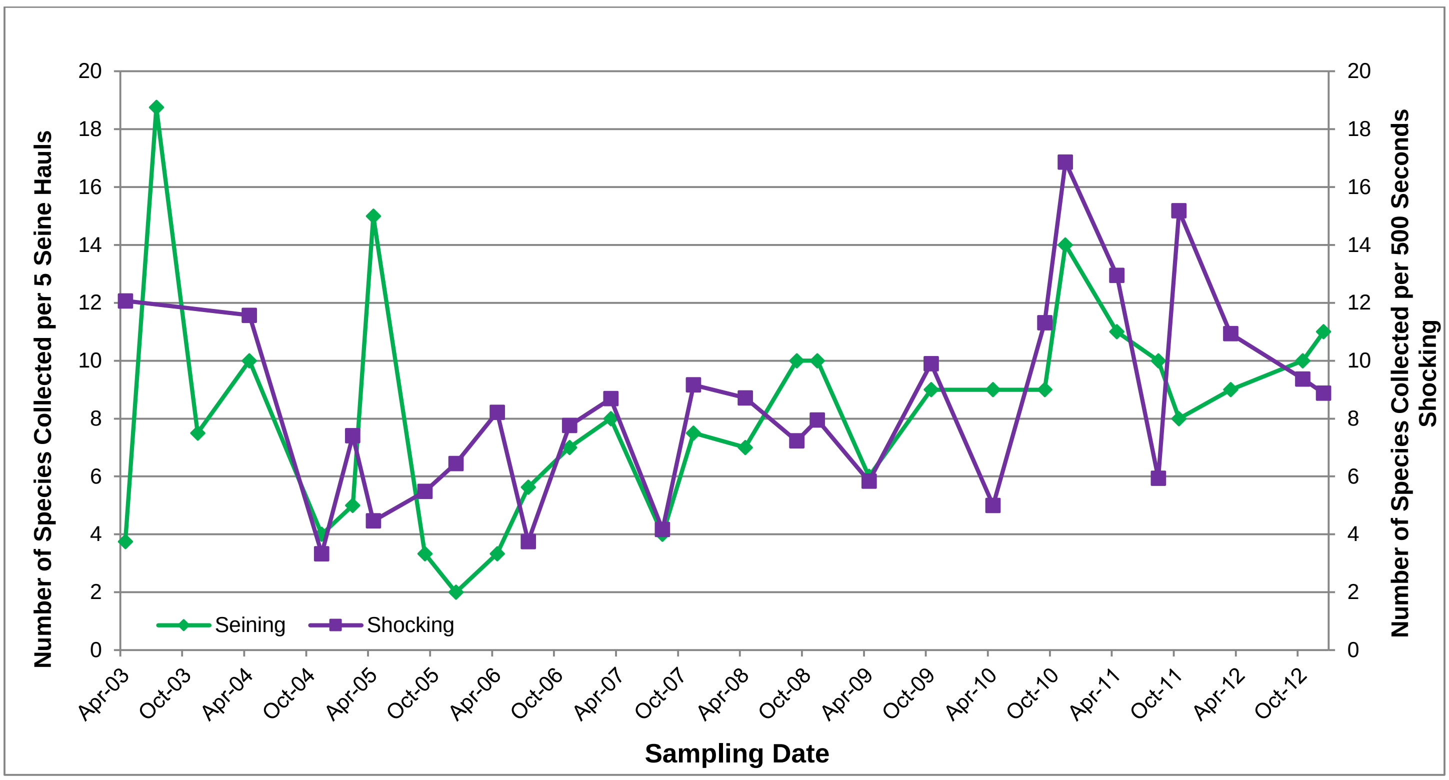

Figure 114. **Fishing richness per unit effort at HCSW-4 from 2003 – 2012 (fish species collected standardized to 5 seine hauls and 500 seconds shocking).**

5.4.6 Summary of Fish Results

Forty-one species of fish were collected in 2003 to 2012, with most captured individuals belonging to one of five families (Table 24). We expect to add very few additional species during future monitoring events, because the species accumulation curves based on the samples collected in 2003 to 2012 have leveled off. Several native species are almost certainly present in Horse Creek but were not collected in 2003 to 2012. These include the American eel (*Anguilla rostrata*) and black crappie (*Pomoxis nigromaculatus*). Samples collected included nine exotic species: walking catfish, African jewelfish, brown hoplo, oriental weatherfish, blue tilapia, vermiculated sailfin catfish, Orinoco sailfin catfish, sailfin catfish, and *P. gibbiceps*. Over 30 species of exotic fish have established reproducing populations in Florida (http://floridafisheries.com), and more will likely continue to be introduced in spite of laws restricting such introductions; thus, we expect to continue to collect additional exotic species in Horse Creek during future monitoring events as new introductions occur and as such species expand their ranges in Florida.

Table 24. Percentage of individual fish captured per year for most abundant fish families/groups in Horse Creek during 2003 – 2012 as part of the Horse Creek Stewardship Program.

Fish Family	HCSW-1	HCSW-2	HCSW-3	HCSW-4	Total
Poeciliidae	60%	97%	90%	79%	87%
Cyprinidae	25%	0.02%	3%	8%	5%
Centrarchidae	7%	1%	2%	5%	3%
Cyprinodontidae	1%	1%	1%	4%	2%
Athernininidae	4%	0%	1%	2%	1%
Exotics	1%	1%	1%	1%	1%

Table 25 presents a summary of the number of individual fish captured for several major fish groups at each station per year, including exotic fish species. At each station, the number of individuals per fish groups varies considerably over time and is heavily influenced by sampling conditions. The inherent variability of the data, as well as the change in sampling effort between years and stations, makes it difficult to look for trends in fish group abundance over time, but a visual examination shows no general trends of increase or decrease in the exotic group or the native fish groups.

During 2012, 25 species of fish were collected from the four Horse Creek sampling stations. In 2012, one new fish species was collected at HCSW-4, the Orinoco sailfin catfish. Fewer fish species were collected at HCSW-1 during two sampling events in 2012 than the other stations because of the unique characteristics of that sampling location (Table 22, Figures 94 - 97). In addition, water levels and streamflow were fairly high during biological sampling at all stations in October which led to HCSW-2 not being sampled; higher water levels did not allow for some habitats to be reached by our sampling equipment. Abnormally cold winters in 2009 – 2010 and 2010 – 2011 may have led to decreased fish diversity at some stations in 2010 and 2011, with evidence of recovery and recruitment in 2012. Over the period of record, fish richness and diversity was lowest at HCSW-2, with no significant annual trends. Fish communities were similar for all years when stations were combined and for all stations when years were combined. Catch per effort is variable over time and dependent on sampling technique, a station's physical characteristics, water levels, and available recruitment sources. No trends were evident in the abundance of fish from exotic and native fish groups.

Table 25. Number of individual fish captured per year for major native and exotic fish groups in Horse Creek during 2003 – 2012 as part of the Horse Creek Stewardship Program.

HCSW-1										
	2003	**2004**	**2005**	**2006**	**2007**	**2008**	**2009**	**2010**	**2011**	**2012**
Native Poecilids	181	78	75	341	25	275	47	328	308	213
Native Sunfish	46	26	33	20	23	24	14	7	14	9
Native Catfish	5	9	3	4	3	2	0	1	2	1
Native Other	25	69	57	140	87	268	33	4	164	155
Exotics	2	1	5	0	0	1	7	0	1	6
Total Fish	259	183	171	505	138	570	101	340	489	384
Sampling Events	3	2	4	3	3	3	1	3	2	2
HCSW-2										
	2003	**2004**	**2005**	**2006**	**2007**	**2008**	**2009**	**2010**	**2011**	**2012**
Native Poecilids	363	1,735	3,093	568	908	1,335	2,519	1,696	394	0
Native Sunfish	41	15	9	13	2	1	1	1	1	0
Native Catfish	1	2	0	0	0	0	0	0	0	0
Native Other	21	61	43	1	6	12	4	50	13	0
Exotics	4	2	22	1	4	40	3	2	0	0
Total	430	1,815	3,176	583	920	1,388	2,527	1,749	408	0
Sampling Events	3	2	4	2	2	3	1	2	1	0
HCSW-3										
	2003	**2004**	**2005**	**2006**	**2007**	**2008**	**2009**	**2010**	**2011**	**2012**
Native Poecilids	669	1,606	4,125	727	489	3,122	1,677	2,874	1,364	2,092
Native Sunfish	49	24	35	31	44	19	5	78	78	28
Native Catfish	1	0	0	0	4	1	0	1	1	2
Native Other	180	114	23	145	202	106	11	215	143	299
Exotics	1	14	37	12	17	23	53	7	3	80
Total	900	1,758	4,220	915	756	3,271	1,746	3,175	1,589	2,501
Sampling Events	3	2	4	3	3	3	1	3	2	2
HCSW-4										
	2003	**2004**	**2005**	**2006**	**2007**	**2008**	**2009**	**2010**	**2011**	**2012**
Native Poecilids	172	713	705	280	62	794	409	2,423	2,112	998
Native Sunfish	52	27	5	67	54	62	66	38	97	74
Native Catfish	6	2	2	0	0	1	0	0	1	1
Native Other	77	52	12	53	174	173	311	205	188	425
Exotics	15	6	31	20	4	12	5	19	3	20
Total	322	800	755	420	294	1,042	791	2,685	2,401	1,518
Sampling Events	3	2	4	3	3	3	2	3	3	3

6 Conclusions

6.1 Water Quantity Results

Although low and median Horse Creek discharge in 2012 was average for the region, rainfall in 2012 was below the long-term average annual rainfall of 52.72 inches (1908-2012) . For 2012, temporal patterns of average daily stream flow and stage were similar across all stations, with the majority of high flows and stages occurring during May to August, during the rainy season. The dry season (January to May 2012) was extremely dry, resulting in periods of little to no streamflow and no NPDES discharge. Summer rains began in late-May/early-June, with lag in streamflow response until mid-June. At the beginning of the wet season, the lag effect between rainfall and streamflow is more apparent as the surrounding wetlands or small creeks either need to fill or resume flow. However, later in the wet season when the wetlands and small creeks are full, the lag is much shorter. Higher streamflow continued through the end of October/first of November.

In September and October, NPDES contributed up to 75 percent of the streamflow at HCSW-1 compared to rainfall; in late October 2012, NPDES discharge accounted for almost all of the streamflow at HCSW-1. NPDES discharge from August to November was also a lagged response to rain that occurred from late-May to early October 2012; NPDES discharge from the Horse Creek outfalls usually does not occur until sufficient water storage accumulates in the circulation system, resulting in a lag. In general, the lower than average rainfall resulted in lower than average streamflow in Horse Creek, with some lags. There is no evidence that mining and reclamation activities in the basin caused any significant decrease in total streamflow in 2012. In a previous study (Robbins and Durbin 2011) that compared streamflow during dry years in reference and potentially impacted streams before and during phosphate mining, there was no evidence that phosphate mining practices caused lower monthly flows in the potentially impacted streams (including Horse Creek) than what would be expected given the conditions in a reference stream (Charlie Creek).

6.2 Water Quality Results

Water quality parameters in 2012 were almost always within the desirable range relative to trigger levels and water quality standards at the station closest to mining (HCSW-1, Table 18). Trigger levels were exceeded only once at HCSW-1 in 2012 with the alkalinity exceedance in November. At HCSW-2, trigger levels were exceeded for dissolved oxygen during half of the year (Table 18). Based upon historical conditions in Horse Creek (Durbin and Raymond 2006), the reported values for dissolved oxygen at HCSW-2 are the result of natural conditions (proximity to hypoxic segment of stream – Horse Creek Prairie) and are not related to mining activities. The chlorophyll *a* trigger level was exceeded during low-flow periods at HCSW-2 in February, April, May, and December, and the pH lower trigger level was exceeded in October. HCSW-3 exceeded trigger levels for dissolved oxygen (June-September), calcium (April), sulfate (February-April and June), and TDS (January-April and June). HCSW-4 exceeded trigger levels for specific conductivity (June), dissolved oxygen (July and September), calcium (April-June), iron (July-October), alkalinity (May), sulfate (March-April and June), and TDS (January-June). Dissolved oxygen triggers were exceeded during summer wet months of 2012, when high temperatures reduce the oxygen carrying capacity of the stream. Sulfate, calcium, TDS, and other ions were exceeded in the dry season, when low rainfall and streamflow likely led to increased groundwater inputs from baseflow and agricultural runoff. Dissolved iron concentrations consistently exceeded the trigger value set at HCSW-4, but Mosaic and the PRMRWSA agree that the trigger value at that station has been set too low given historical and upstream concentrations of dissolved iron. Based on impact assessments already completed, none of the observed exceedances pose a significant adverse ecological impact to Horse Creek that would be attributable to mining.

Several of the trends detected during the statistical analysis either have an estimated slope that 1) was not in the direction of an adverse trend (color, ammonia, and dissolved iron) or 2) was very small compared to limits in laboratory prediction or the observed differences between primary and duplicate field samples (pH, ammonia, orthophosphate). For the trends with higher estimated rates of change (specific conductivity and various dissolved ions), the potential trends are discussed in Appendix I. Orthophosphate was also discussed further in this impact assessment as a follow-up to the trend impact assessment of the 2010 Annual Report (Robbins et al. 2014). Appendix I shows that the apparent trend in orthophosphate from 2003-2012 is caused by a data bias, and that extending the period of record into pre-2003 data eliminates this trend. Specific conductivity and other ions may have been influenced by changes in mining practices, but the current concentrations are stable and not biologically harmful.

A significant increasing trend in specific conductivity was detected at HCSW-1 even when the period of record was expanded, but the inclusion of data through 2012 (Appendix I) makes it clear that the most visible increase is a single step-change rather than a monotonic, continuing increase. In addition, at least part of the increase may be caused by regional influences, as Charlie Creek also shows a slow increase in conductivity over time. In addition, specific conductivity at Horse Creek and West Fork Horse Creek stations upstream of the NPDES outfalls has also been increasing over the same time period (data found in Mosaic DRI reports to Manatee County). Specific conductivity at HCSW-1 began to rise during a very dry period in 2006-2008 when Mosaic had very little NPDES discharge into Horse Creek, and the stream was dominated by baseflow (surficial aquifer) contributions. After the Horse Creek outfalls began to receive water decanting from Wingate Mine clays, the specific conductivity remained at the higher levels, presumably because of the greater proportion of groundwater involved in the dredge mining. If groundwater influence is the main cause of increased specific conductivity at HCSW-1, then concentrations have reached a threshold and should not continue to increase. This theory has been validated considering that the range and median specific conductivity from 2009 to 2012 has been very consistent with no expectation that it will increase from the recently observed range in the future.

Dissolved ion concentrations at HCSW-1 in recent years still meet all applicable state drinking water and Class III surface water standards, except where otherwise noted in previous monthly impact assessments. In addition, the current specific conductivity concentrations at HCSW-1 are within the preferred range of Horse Creek freshwater fish species, and the diversity and composition of fish populations have remained steady over the HCSP study period with no correlation with specific conductivity. SCI Scores, an indicator of benthic Macroinvertebrate community health, have remained steady over the HCSP study period and show no relationship with specific conductivity.

Significant differences between stations were evident for several parameters. Overall, HCSW-2 was the most dissimilar from the other three stations, especially in pH, dissolved oxygen, and some dissolved ions. Some nutrients (nitrate + nitrite) and dissolved ions (specific conductivity, calcium, chloride, sulfate) had higher concentrations downstream in Horse Creek, probably because of increased groundwater seepage and agricultural runoff in the lower Horse Creek basin. Differences in topography, geology, and land use that could account for these trends in Horse Creek are examined in the Horse Creek Stewardship Program Historical Report (Durbin and Raymond 2006).

Water quality parameters were compared with water quantity variables recorded during the same month: average daily streamflow for the month, average daily NPDES discharge for the month, and total monthly rainfall (Figure 75). In general, pH, dissolved oxygen, and most dissolved ions are higher when the overall quantity of water in the Horse Creek system is low. (In this report, specific conductivity, calcium, alkalinity, sulfate, and TDS showed the opposite pattern with NPDES discharge at HCSW-1; possible reasons behind this are discussed as part of the trend analysis in Appendix I.) Conversely, turbidity, color, iron, and nitrogen are high when the water quantity is also high. When water quantity in Horse

Creek is low, the stream may be pooled or slow-moving, leading to algal blooms that may increase pH and chlorophyll a. In addition, the majority of water in the stream during low quantity periods may be from groundwater (seepage or agricultural runoff); groundwater has a higher concentration of dissolved ions than surface water. When water quantity is high, an increased amount of sediment and organic debris is washed into the stream, leading to increases in turbidity, color, iron, and nitrogen.

6.3 Benthic Invertebrate Results

Benthic invertebrate habitat scores were “Optimal” to “Sub-optimal” and SCI scores were “Healthy” or “Exceptional” at all stations in 2012; these scores are typical of southwestern Florida streams, including those used to develop the Habitat Assessment and SCI indices. Overall, taxa diversity indices and SCI metrics show few monotonic trends over time and are very similar between sampling events and stations. However, HCSW-2 has slightly lower diversity and significantly lower SCI metrics than other stations. Habitat conditions at HCSW-2 are consistently poor, with lower streamflow, dissolved oxygen, and pH than other Horse Creek stations. The source of the poor conditions are related to the lower than average streamflow and rainfall of the previous few years and the presence of Horse Creek Prairie, the large marsh located upstream of the biological sampling station.

6.4 Fish Results

During 2012, 25 species of fish were collected from the four Horse Creek sampling stations. In 2012, one new fish species was collected at HCSW-4, the Orinoco sailfin catfish. Fewer fish species were collected at HCSW-1 during two sampling events in 2012 than the other stations because of the unique characteristics of that sampling location (Table 22, Figures 94 - 97). In addition, water levels and streamflow were fairly high during biological sampling at all stations in October which led to HCSW-2 not being sampled; higher water levels did not allow for some habitats to be reached by our sampling equipment. Abnormally cold winters in 2009 – 2010 and 2010 – 2011 may have led to decreased fish diversity at some stations in 2010 and 2011, with evidence of recovery and recruitment in 2012. Over the period of record, fish richness and diversity was lowest at HCSW-2, with no significant annual trends. Fish communities were similar for all years when stations were combined and for all stations when years were combined. Catch per effort is variable over time and dependent on sampling technique, a station’s physical characteristics, water levels, and available recruitment sources. No trends were evident in the abundance of fish from exotic and native fish groups.

7 Recommendations

7.1 Previous Annual Report Recommendations

During the TAG meetings for the 2010 Annual Report draft (February 29, 2012) and the 2010 Impact Assessment (August 20, 2013), the following recommendations were made:

- In the 2010 HCSP Annual Report and subsequent reports, Cardno will add the report year's median water quality concentrations to the water quality trend summary table for context.
    - Included in final 2010 HCSP Annual Report and Appendix H.
- In the 2010 HCSP Annual Report and subsequent reports, Cardno will add a table, graphic, or text depicting major mine operation changes or alterations both during and prior to the HCSP.
    - Included in final 2010 HCSP Annual Report, Appendix I, and included in Appendix K of the 2011 HCSP Annual Report and in all subsequent reports.
- In the 2010 HCSP Annual Report and subsequent reports, Cardno will add a paragraph to explain the NPDES discharge make-up in the past and how it has changed in both water quantity and quality.
    - Included in final 2010 HCSP Annual Report, Appendix I and all subsequent reports.
- In the 2010 HCSP Annual Report, Cardno will include reference to the Streamflow Analysis Report that was prepared for SWFWMD (which was part of the EMP for Mosaic's WUP).
    - Included in final 2010 HCSP Annual Report.
- In the 2010 report and subsequent reports Cardno will add a sentence stating that all graphs within the report contain data collected only during the HCSP period-of-record unless otherwise noted.
    - Included in final 2010 HCSP Annual Report.
- In the 2010 report and subsequent reports Cardno will provide a discussion of the table showing native versus exotic fish species.
    - Included in final 2010 HCSP Annual Report.
- Cardno will provide a milestone section (as an appendix) in the 2011 report and all future reports.
    - Included as Appendix K in 2011 HCSP Annual Report and will be in all subsequent reports.
- Cardno will add the outfall locations to Figure 3 (Mining and Reclamation) in the 2011 and future reports.
    - Included in 2011 HCSP Annual Report and all subsequent reports.

- Cardno and Mosaic will research total pumpage (removal) and the number of SWFWMD Water Use Permits (WUPs) within the Peace River Basin. This may be incorporated into the 2012 Annual Report, if appropriate.
    - This information has not be received to date.
- Cardno will expand flow discussion in 2012 Annual Report.
    - Included in 2012 HCSP Annual Report and all subsequent reports.
- Cardno will add new 2012 FDEP SCI SOP and NNC language to 2012 or 2013 Annual Reports.
    - See Section 7.3.
- Cardno will add a discussion of current to previous stream conditions in Section 4 of the 2012 and future reports.
    - Will be included in the 2013 or 2014 report.

7.2 Current TAG Recommendations

Cardno will contact the Authority for rain gauge data collected near their facility on the Peace River.

7.3 Current Annual Report Recommendations

During the TAG meeting for the 2011 and 2012 draft Annual Reports (October 29, 2014), the following recommendations were made:

- Cardno will add a graphic to the 2011 and all subsequent reports showing more historical flow in Horse Creek extending beyond the HCSP time period (back to at least 1978 to match the double mass curve).
- Cardno will add more discussion of the double mass curve to the 2011 and all subsequent reports.
- Cardno will add a Trend Summary Table to the 2011 Annual report and all subsequent reports instead of only including in Appendix I.
- In the 2013 Annual Report, Cardno will add an appendix that lists any erroneous data and remove problematic data from Appendix C graphs. Any future erroneous data will be added to the appendix as the program continues.
- For the 2013 Annual Report TAG Meeting, Cardno will include a discussion on the DO Saturation standard and implications in the main body of the report.
- In the 2014 annual report, Cardno will add a discussion on NNC standards and implications in the main body of the report.
- In the 2014 annual report, Cardno will add a discussion on legacy CF Industries operations in the Horse Creek Basin.

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APPENDIX A

HORSE CREEK STEWARDSHIP PROGRAM

Appendix A
Horse Creek Stewardship Program

Intent

The purpose of this program is two-fold. First, it provides a protocol for the collection of information on physical, chemical and biological characteristics of Horse Creek during IMC Phosphates' (IMC) mining activities in the watershed in order to detect any adverse conditions or significant trends that may occur as a result of mining. Second, it provides mechanisms for corrective action with regard to detrimental changes or trends caused by IMC's' activities, if any are found.

The overall goals of the program are to ensure that IMC Phosphates' mining activities do not interfere with the ability of the Peace River/Manasota Regional Water Supply Authority (Authority) to withdraw water from the Peace River for potable use nor adversely affect Horse Creek, the Peace River or Charlotte Harbor.

There are three basic components to this stewardship program:

Monitoring and Reporting on Stream Quality,

Investigating Adverse Conditions or Significant Trends Identified Through Monitoring, and

Implementing Corrective Action for Adverse Stream Quality Changes Attributable to IMC Activities

An important aspect of this program is that it will not rely solely upon the exceedance of a standard or threshold to bring about further investigation and, where appropriate, corrective action. The presence of a significant temporal trend alone will be sufficient to initiate such steps. This protection mechanism is not present in the vast majority of regulatory scenarios.

The mission of the Authority is to provide a reliable and safe drinking water supply to the citizens of the four counties comprising the Authority, Charlotte, DeSoto, Manatee and Sarasota Counties. The Peace River Facility is a critical component of the Authority's water supply system. The Peace River Facility located in DeSoto County utilizes the Peace River as its supply source.

It is critical for the Authority to protect the Peace River from impacts that would be detrimental to the operation of the Peace River Facility. As a tributary to the Peace River, the Authority's goal for the Horse Creek Stewardship Program is to provide assurance that the quantity and quality of Horse Creek flow as it contributes to the Peace River does not adversely impact the operation of the Peace River Facility.

Program Implementation and Oversight

IMC will implement and fund the Horse Creek Stewardship Program with oversight by the Authority. The Authority will create and coordinate a Technical Advisory Group (TAG) to consist of a representative from each of its members to review and provide input on the program throughout the duration of the monitoring. IMC will create a project-specific quality assurance and quality control (QA/QC) plan for the program detailing all sampling, laboratory procedures, benthic and fish monitoring protocols and data analysis. The QA/QC plan will be consistent with the analogous protocols established in the HydroBiological Monitoring Program (HBMP) for the Lower Peace River/Upper Charlotte Harbor.

Historical, Background and Contemporaneous Data

IMC will compile available data collected by others on water quality, quantity and aquatic biology of Horse Creek. This is expected to include, but is not limited to, information collected by the U.S. Geological Survey (USGS), the Florida Department of Environmental Protection (DEP), the Southwest Florida Water Management District (SWFWMD), the Charlotte Harbor Environmental Center (CHEC). Horse Creek data contained in the U.S. Environmental Protection Agency's (EPA) STORET database will also be obtained. Historic data will be reviewed to provide background information on Horse Creek, and data from ongoing collection efforts will be obtained to supplement that collected by IMC.

Monitoring Period

Water quantity, water quality, macroinvertebrates and fish will be monitored as outlined below during the time that IMC Phosphates is conducting mining and reclamation in the Horse Creek watershed. Monitoring will begin no later than April 2003. In the event of temporary interruptions in mining activities (up to one year), this monitoring will continue during the period of inactivity. Monitoring will cease when mining and reclamation operations are completed in the Horse Creek watershed.

1. Surface Water Monitoring Stations

Four locations on Horse Creek will be monitored for physical, chemical and biological parameters:

HCSW-1 - Horse Creek at State Road 64 (USGS Station 02297155)
HCSW-2 - Horse Creek at County Road 663A (Goose Pond Road)
HCSW-3 - Horse Creek at State Road 70
HCSW-4 - Horse Creek at State Road 72 (USGS Station 02297310)

As indicated above by their station ID numbers, HCSW-1 and HCSW-4 are also long-term US Geological Survey (USGS) gauging stations, with essentially continuous stage and discharge records since 1977 and 1950, respectively.

2. Water Quantity Monitoring and Analysis

Discharge data will be obtained from the USGS for stations HCSW-1 and HCSW-4 for compilation with other data collected through this monitoring program. If not already present, staff gauges will be installed in the stream at HCSW-2 and HCSW-3 and surveyed to NGVD datum. If not already available, stream cross sections will be surveyed at those locations, extending to the approximate limits of the 25-year floodplain. Staff gauge readings will be recorded at the time of any sampling efforts at those stations. Data on rainfall will be obtained using IMC's rain gauge array (including any additional gauges installed in the Horse Creek basin in the future).

Data analysis will focus upon, but not necessarily be limited to, the ongoing relationship between rainfall and streamflow in the Horse Creek watershed. This relationship can be established from data collected early in the monitoring program and used to track the potential effects of mining on streamflow. Analytical approaches are outlined under Water Quality below and such methods will be more fully described in the QA/QC plan to be developed as part of this stewardship program.

3. Surface Water Quality Monitoring and Analysis

Water quality data will be obtained monthly at each station where flow is present. Field measurements will be made of temperature, pH, specific conductance, turbidity and dissolved oxygen. Grab samples will be collected and analyzed for:

Nitrate + Nitrite
Total Kjeldahl Nitrogen
Total Nitrogen
Total Ammonia Nitrogen
Ortho Phosphate
Chlorophyll a
Calcium
Iron

Color
Total Alkalinity
Chloride
Fluoride
Radium 226 + 228
Sulfate
Mining Reagents (petroleum-based organics, fatty acids, fatty amido amines).

At Station HCSW-1, a continuous monitoring unit will be installed to record temperature, pH, conductivity, dissolved oxygen and turbidity. Because this station is located at a bridge crossing for a highway, the unit will be located some distance (within 100 m) upstream or downstream from the bridge to minimize the likelihood of vandalism. The unit will be permanently installed and its location surveyed. Data will be recorded frequently (at least hourly) and will be downloaded at least monthly. This data will provide for the characterization of natural background fluctuations and may allow for the detection of general water quality changes not observed during the collection of monthly grab samples.

Table 1 presents the analytical schedules and procedures. All sampling will be conducted according to DEP's Standard Operating Procedures (SOP) for field sampling. Laboratory analyses will be performed by experienced personnel according to National Environmental Laboratory Accreditation Council (NELAC) protocols, including quality assurance/quality control considerations. Invertebrate sampling will be conducted by personnel with training and experience in the DEP's SOP for such sampling.

Results will be tabulated to allow for comparisons among stations and sampling events and through time. Results will be compared with available historic data for Horse Creek and its tributaries, and with applicable Florida surface water quality standards. Typical parametric and non-parametric statistics will be used to describe the results. In particular, regression analysis is expected to be employed to examine the relationship between each parameter and time. Both linear and non-linear regression will be considered, depending upon the patterns observed in the data. Since at least some of the parameters can be expected to vary seasonally, use of methods such as the Seasonal Kendall's Tau Test is anticipated. Other potential methods include Locally Weighted Scatterplot Smooth (LOWESS). In addition to trend analyses, annual reports will contain general statistics such as mean, median, standard deviation and coefficient of variance for each numerical parameter. Such general statistics will be calculated on both an annual and seasonal basis. Because the data will be maintained in a standard software format (i.e., MS Excel or MS Access), there will be virtually no logistical limitations on the types of analyses that can be conducted. The only limitations will result from the nature of the data itself (i.e., data quantity, distributions, etc.).

For each parameter, data analysis will focus upon, but not necessarily be limited to, (1) the relationship between measured values and the "trigger values" as presented in Table 1 and (2) temporal patterns in the data which may indicate a statistically significantly trend toward the trigger value. Statistical significance will be based upon $\alpha=0.05$, unless data patterns/trends or other related information indicate that use of another significance level is more appropriate. Since the purpose of this monitoring is to detect trends toward the trigger values, should they be present, trend analyses and other statistical tests will generally focus only upon changes toward the trigger values. This will increase the statistical power for detecting such changes.

At least initially, the term over which trends are analyzed will be dependent upon the data collected to date. As the period of record increases, data analysis can move from a comparison of months, to seasons, to years. As noted above, seasonal patterns will always be considered during data analysis and attention will be given to differentiation between natural seasonal/climatic variation and anthropogenic effects (including mining), where possible. Where historic data exist for a given parameter or station, such data can be evaluated relative to that collected through this effort, although sampling frequency and consistency may not be sufficient to conduct standard trend analysis methods. Analytical methods will be more fully described in the QA/QC plan to be developed as part of this stewardship program.

4. Aquatic Macroinvertebrate Sampling and Analysis

Macroinvertebrate sampling will be performed three times annually and, in general, will be conducted concurrently with a monthly water quality sampling event. The first event would occur in March or April, the second event in July or August, and the third event in October or November. Specific months when sampling occurs may change from year to year to avoid very low or very high flows, which would impede representative sampling.

In accordance with the DEP Standard Operating Procedures (DEP-SOP-001/01 FS 7000 General Biological Community Sampling), invertebrate sampling will not be conducted ". . . during flood stage or recently dry conditions." This is interpreted here to mean that a given sampling station will not be sampled for macroinvertebrates if (a) water is above the top of the stream bank, or is too deep or fast-moving to sample safely, or (b) if the stream has been dry during the preceding 30 days. In the event either of these situations occurs, the station will be revisited approximately one month later to determine whether sampling is appropriate at that time. If the stream is still in flood, or has again been dry during the preceding 30 days, invertebrate sampling will be postponed until the next season's sampling event. Note that the above situations are expected to be quite rare at the Horse Creek stations, and sampling efforts will generally be planned to avoid such conditions.

Sampling will be conducted at the same four stations on Horse Creek used for flow and water quality monitoring. The aquatic habitats at each station will be characterized, streamside vegetation surveyed, and photostations established. Qualitative macroinvertebrate sampling will be performed according to the Stream Condition Index (SCI) protocol developed by DEP (DEP-SOP-002/01 LT 7200) or subsequently DEP-approved sampling methodology. Consistent with DEP protocols, each invertebrate sample will be processed and taxonomically analyzed. Data from the samples will be used to determine the ecological index values presented in Table 1. Additional indices may also be calculated to further evaluate the invertebrate community. As noted in Table 1, the focus of the analysis will be to screen for statistically significant declining trends with respect to presence, abundance, and distribution of native species, as well as SCI values. Results may also be compared with available historic macroinvertebrate data for Horse Creek and its tributaries, or with data from other concurrent collecting efforts in the region, if appropriate. Analysis of invertebrate community characteristics will include consideration of flow conditions, habitat conditions and selected water quality constituents.

Analytical approaches are outlined under Water Quality Monitoring and Analysis section above and such methods will be more fully described in the QA/QC plan to be developed as part of this Horse Creek Stewardship Program.

5. Fish Sampling and Analysis

Fish sampling will be conducted three times annually, concurrent with aquatic macroinvertebrate sampling at the same four stations on Horse Creek. Based upon stream morphology, flow conditions and in-stream structure (logs, sand bars, riffles, pools, etc.); several methods of sampling may be used, including seining, dipnetting, and electrofishing. Sample collection will be timed to standardize the sampling efforts among stations and between events.

All fish collected will be identified in the field according to the taxonomic nomenclature in Common and Scientific Names of Fishes from the United States and Canada (American Fisheries Society 1991, or subsequent editions). Voucher specimens will be taken of uncommonly encountered species and of individuals that cannot be readily identified in the field; with such specimens being preserved and logged in a reference collection maintained for this monitoring program. All fish will be enumerated and recorded. Total length and weight will be determined and recorded for individuals, however, for seine hauls with very large numbers of fish of the same species (a common occurrence with species like *Gambusia holbrooki*, *Heterandria formosa* and *Poecilia latipinna*), individuals of the same species may be counted and weighed en masse, with only a randomly selected subset (approximately 10 to 20 individuals of each such species) being individually measured for length and weight. Any external anomalies observed on specimens will be recorded.

Taxa richness and abundance and mean catch per unit effort will be determined for each station and each event, and data can be compared among stations and across sampling events. The ecological indices presented in Table 1 will be calculated and additional indices may also be calculated to evaluate the fish community, including similarity indices, species accumulation/rarefaction curves, diversity indices and evenness indices. As noted in Table 1, the focus of the analysis will be to screen for statistically significant declining trends with respect to presence, abundance and distribution of native species. Results may also be compared with available historic fisheries data for Horse Creek and its tributaries, and with data from other concurrent regional collecting efforts, if applicable. Analysis of fish community characteristics will include consideration of flow conditions, habitat conditions and selected water quality constituents.

Analytical approaches are outlined under Water Quality above and such methods will be more fully described in the QA/QC plan to be developed as part of this stewardship program.

6. Reporting

All data collected through this monitoring program will be compiled annually (January - December records) and a report will be generated summarizing the results. This report will include narrative, tabular and graphical presentation of the discharge records, surface water quality data, macroinvertebrate and fish sampling results. Results of statistical analyses will also be provided. Discussion will be included comparing across the sampling stations, as well as among seasons and sampling years. Emphasis will be placed upon identifying spatial and/or temporal trends in water quality and/or biological conditions. Where available, data collected from the same stations prior to the initiation of this program will be reviewed and incorporated to allow for longer-term evaluation of Horse Creek. In addition, data available from sampling/monitoring efforts by agencies or other public entities will be reviewed and incorporated, where pertinent. Each report will also provide general information on the location and extent of IMC mining activities in the Horse Creek watershed, as they relate to this monitoring effort. Reports will be submitted to the Authority, as well as to the DEP Bureau of Mine Reclamation (BMR) and Southwest Florida Water Management District (SWFWMD).

In addition to the reporting outlined above, raw data compiled through sampling will be provided to the Authority monthly. This data will be submitted within six (6) weeks of each sampling event (pending the completion of laboratory/taxonomic analyses).

Monitoring Program Evaluation

To ensure this program is providing useful information throughout its tenure, it will be evaluated regularly. Each annual report will include a section devoted to a summary of the immediate and long-term utility of each information type being collected. Recommendations will also be provided in the report regarding possible revisions, additions or deletions to the monitoring program to ensure that it is appropriately focused. Based upon such recommendations, IMC Phosphates will coordinate with the Authority and TAG on a regular basis regarding amendments to the monitoring program. Coordination on this issue may be initiated at any time by either party and will occur at least once every five years, whether or not either party individually requests it.

Protocol for Addressing Potential Problems Identified Through Monitoring

An important element of the monitoring program will be the ongoing analyses of data to detect exceedances of specific trigger values (see Table 1) as well as statistically significant temporal trends toward, but not necessarily in excess of, those values. The analyses will evaluate the data collected through this Horse Creek Stewardship Program, as well as that reported by other entities where appropriate.

Impact Assessment/Characterization

In the event the annual data evaluation identifies trigger value exceedances or statistically significant trends in Horse Creek, IMC will conduct an impact assessment to identify the cause of the adverse trend. The impact assessment may include more intensive monitoring of water quality in terms of frequency of sampling, laboratory analyses conducted, or locations monitored. In all cases, however, the impact assessment will include supplemental quantitative and qualitative data evaluations and consultation with Authority scientists, as well as perhaps other investigations within the basin (e.g., examination of land use changes, discharge monitoring records reviews of others, water use permit reports of others, etc.).

If the "impact assessment" demonstrates to the satisfaction of IMC and Authority scientists that IMC's activities in the Horse Creek watershed did not cause the exceedance or trend, IMC would support the Authority's efforts to implement actions to reverse or abate the conditions. IMC's support will focus upon scientific solutions where IMC can assist in the abatement of others' problems.

If the impact assessment indicates or suggests that IMC is the cause of the exceedances or trend, then IMC shall take immediate corrective actions. The intensity of such actions would be based upon the potential for ecological harm to the ecology of Horse Creek or the integrity of the potable water supply to the Authority.

Corrective Action Alternatives Evaluation and Implementation

The first step in the corrective action process shall be to prepare quantitative projections of the short-term and long-term impacts of the trigger value exceedance or adverse trends. Quantitative models and other analytical tools will provide IMC and Authority scientists with the analyses necessary to determine: (1) whether the impacts will persist or subside over the long term; (2) the cause(s) of the adverse trend(s) in terms of specific IMC activities that are contributing to the trend(s); and (3) alternative steps that IMC could effectuate to reverse the adverse trend, if needed.

If impact modeling confirms that adverse trends in water quality or a trigger value exceedance is caused by IMC activities in the Horse Creek watershed, IMC shall meet with Authority within 30 days of detection of the adverse trend or trigger exceedance to evaluate alternative solutions developed by IMC. IMC shall begin implementation of its proposed alternative solution selected by the Authority within 30 days and report to Authority as implementation milestones are reached. Throughout the modeling, alternatives assessment, and preferred alternative implementation steps of the corrective action process, more intensive impact assessment monitoring will continue to track the continuation, or the abatement, of the trigger value exceedance or adverse trend. Only when the impact assessment monitoring demonstrated conclusively that the condition has been reversed, with respect to the particular parameter(s) of concern, would IMC reduce its efforts back to the general monitoring and reporting program.

Alternative solutions may include conventional strategies such as the implementation of additional best management practices, raw material substitutions, hydraulic augmentation of wetlands, etc. IMC shall consider "out of the box" solutions (such as discharges of water to result in lower downstream concentrations of a parameter of concern, where the pollutant does not originate from IMC's activities) and emerging principles and technologies for water quantity management, water quality treatment and watershed protection, as well as other innovative solutions recommended by Authority.

Table 1. Parameters, General Monitoring Protocols and Corrective Action Trigger Values for the Horse Creek Stewardship Plan

Pollutant Category	Analytical Parameters	Analytical Method	Reporting Units	Monitoring Frequency	Trigger Level	Basis for Initiating Corrective Action Process
General Physio-chemical Indicators	pH	Calibrated Meter	Std. Units	Monthly	<6.0- >8.5	Excursions beyond range or statistically significant trend line predicting excursions from trigger level minimum or
	Dissolved Oxygen	Calibrated Meter	mg/L[1]	Monthly	<5.0	Excursions below trigger level or statistically significant trend line predicting concentrations below trigger level.
	Turbidity	Calibrated Meter	NTU[2]	Monthly	>29	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Color	EPA 110-2	PCU	Monthly	<25	Excursions below trigger level or statistically significant trend line predicting concentrations below trigger level.
Nutrients	Total Nitrogen	EPA 351 + 353	mg/L	Monthly	>3.0	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Ammonia	EPA 350.1	mg/L	Monthly	>0.3	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Ortho Phosphate	EPA 365	mg/L	Monthly	>2.5	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Chlorophyll *a*	EPA 445	mg/L	Monthly	>15	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
Dissolved Minerals	Specific Conductance	Calibrated Meter	µs/cm[3]	Monthly	>1,275	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Alkalinity	EPA 310.1	mg/L	Monthly	>100	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Calcium	EPA 200.7	mg/L	Monthly	>100	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Iron	EPA 200.7	mg/L	Monthly	>0.3 [6]; >1.0[7]	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Chloride	EPA 325	mg/L	Monthly	>250	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Fluoride	EPA 300	mg/L	Monthly	>1.5[6]; >4[7]	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Radium 226+228	EPA 903	pCi/L[4]	Quarterly	>5	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Sulfate	EPA 375	Mg/L	Monthly	>250	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
	Total Dissolved Solids	EPA 160	Mg/L	Monthly	>500	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.
Mining	Petroleum Range Organics	EPA 8015 (FL-PRO)	mg/L	Monthly[5]	>5.0	Exceedance of or statistically significant trend line predicting concentrations in excess of, trigger level.

Pollutant Category	Analytical Parameters	Analytical Method	Reporting Units	Monitoring Frequency	Trigger Level	Basis for Initiating Corrective Action Process
Reagents	Total fatty acids, including Oleic, Linoleic, and Linolenic acid.	EPA/600/4-91/002	mg/L	Monthly(5)	>NOEL	Statistically significant trend line predicting concentrations in excess of the No Observed Effects Level (NOEL to be determined through standard toxicity testing with IMC reagents early in monitoring program, NOEL to be
	Fatty amido-amines	EPA/600/4-91-002	mg/L	Monthly(5)	>NOEL	Statistically significant upward trend line predicting concentrations in excess of No Observed Effects Level (NOEL to be determined through standard toxicity testing with IMC reagents early in monitoring program, NOEL to be expressed as a concentration – e.g., mg/L)
Biological Indices: Macroinverte brates	Total Number of Taxa Abundance Percent Diptera Number of Chironomid Shannon Weaver Florida Index EPT Index Percent Contribution of Dominant Taxon Percent Suspension Feeders/Filterers	Stream Condition Index (SCI) sampling protocol, taxonomic analysis, calculation of indices according to SOP-002/01 LT 7200 Stream Condition Index (SCI) Determination	Units vary based upon metric or index	3 times per year	N/A	Statistically significant declining trend with respect to SCI values, as well as presence, abundance or distribution of native species
Biological Indices: Fish	Total Number of Taxa Abundance Shannon-Weaver Species Turnover (Morisita Similarity Index(a) Rarefaction/Species Accumulation Curves(b)	Various appropriate standard sampling methods, taxonomic analysis, calculation of indices using published formulas	Units vary based upon metric or index	3 times per year	N/A	Statistically significant declining trend with respect to presence, abundance or distribution of native species

Notes:

(1) Milligrams per liter
(2) Nephelometric turbidity units
(3) Microsiemens per centimeter.
(4) PicoCuries per liter.
(5) f reagents are not detected after two years, sampling frequency will be reduced to quarterly - if subsequent data indicate the presence of reagents, monthly sampling will be resumed.
(6) At Station HC SW-4 only, recognizing that existing levels during low-flow conditions exceed the trigger level.
(7) At Stations HC SW-1, HC SW-2, and HC SW-3.

References:

(a) Brower, J. E., Zar, J. H., von Ende, C. N. Field and Laboratory Methods for General Ecology. 3rd Edition. Wm. C. Brown Co., Dubuque, IA. pp. 237; 1990
(b) Gotelli, N.J., and G.R. Graves. 1996. Null Models in Ecology. Smithsonian Institution Press, Washington, DC.

APPENDIX

B

CUMULATIVE CHRONOLOGICAL LIST OF PROCEDURAL CHANGES TO THE HCSP

Appendix B
Cumulative Chronological List of Procedural Changes to the HCSP

Change 1: Summer Biological sampling from July – Aug to July – Sep.
Year Implemented: 2004
Comments: Allows flexibility with sampling during high flows.
Provisional Acceptance: 2004
Final Acceptance: April 4, 2007

Change 2: Fall Biological sampling from Oct – Nov to Oct – Dec.
Year Implemented: 2004
Comments: Allows flexibility with sampling during high flows.
Provisional Acceptance: 2004
Final Acceptance: April 4, 2007

Change 3: Biological sampling should be separated by at least 6 weeks in time.
Year Implemented: 2004
Comments: Ensures that sample results capture seasonal variation.
Provisional Acceptance: 2004
Final Acceptance: April 4, 2007

Change 4: Accept that historical background levels of dissolved iron at HCSW-4 exceeds the trigger level of 0.3 mg/l.
Year Implemented: 2004
Comments: Station HCSW-4 trigger levels reflect the more stringent Class I levels. Historically Station HCSW-4 background levels for dissolved iron are similar to the rest of the basin but also higher than 0.3.
Provisional Acceptance: 2004
Final Acceptance: April 4, 2007

Change 5: Accept that historical background levels of dissolved oxygen and chlorophyll at HCSW-2 exceeds the trigger level.
Year Implemented: 2004
Comments: Station HCSW-2 is directly downstream of Horse Creek Prairie which routinely delivers slow moving water low in dissolved oxygen and high in chlorophyll to station HCSW-2
Provisional Acceptance: 2004
Final Acceptance: April 4, 2007

Change 6: Continue to compile, compare, present and discuss ongoing Horse Creek Data from WMD, DEP and USGS with HCSP data.
Year implemented: 2005
Comments: Enhances program
Provisional Acceptance: July 2006
Final Acceptance: April 4, 2007

Change 7: Biological Sampling stage level criteria from > 10 ft at HCSW-1 & > 5 ft at HCSW-4 to > 10 ft at HCSW-1 & > 4 ft at HCSW-4
Year implemented: 2007
Comments: Biological samples will be collected when stage levels are below these stated levels to ensure safety and quality samples.
Provisional Acceptance: July 2006
Final Acceptance: April 4, 2007

Change 8: The data range used in the historical water quality comparison should be static historical data beginning somewhere around 1990-1993.
Year Implemented: Beginning with the 2007 Annual Report.
Comments: Historical water quality comparison should be static instead of a moving window allowing consistent and continuous comparison with historical data.
Provisional Acceptance: June 2008
Final Acceptance: November 4, 2009

Change 9: Add clay settling area (CSA) FM-1 to existing monitoring program.
Year Implemented: Prior to 2009 wet season.
Comments: Recently constructed SCA FM-1 will be added to existing CSA’s providing real time monitoring to Authority.
Provisional Acceptance: March 2009
Final Acceptance: November 4, 2009

Change 10: Deletion of three water quality parameters: FL-PRO, Fatty Acids, and Total Amines.
Year Implemented: 2009
Comments: These parameters have rarely been above the detection limit and chemical processing plants are not found in the Horse Creek watershed.
Provisional Acceptance: March 2009
Final Acceptance: November 4, 2009

Change 11: Addition of new water quality sample location for Brushy Creek @ Hwy 64.
Year Implemented: 2009
Comments: In lieu of deleted three parameters Mosaic will collect samples and provide data monthly from this location minus trigger levels and impact assessments.
Provisional Acceptance: March 2009

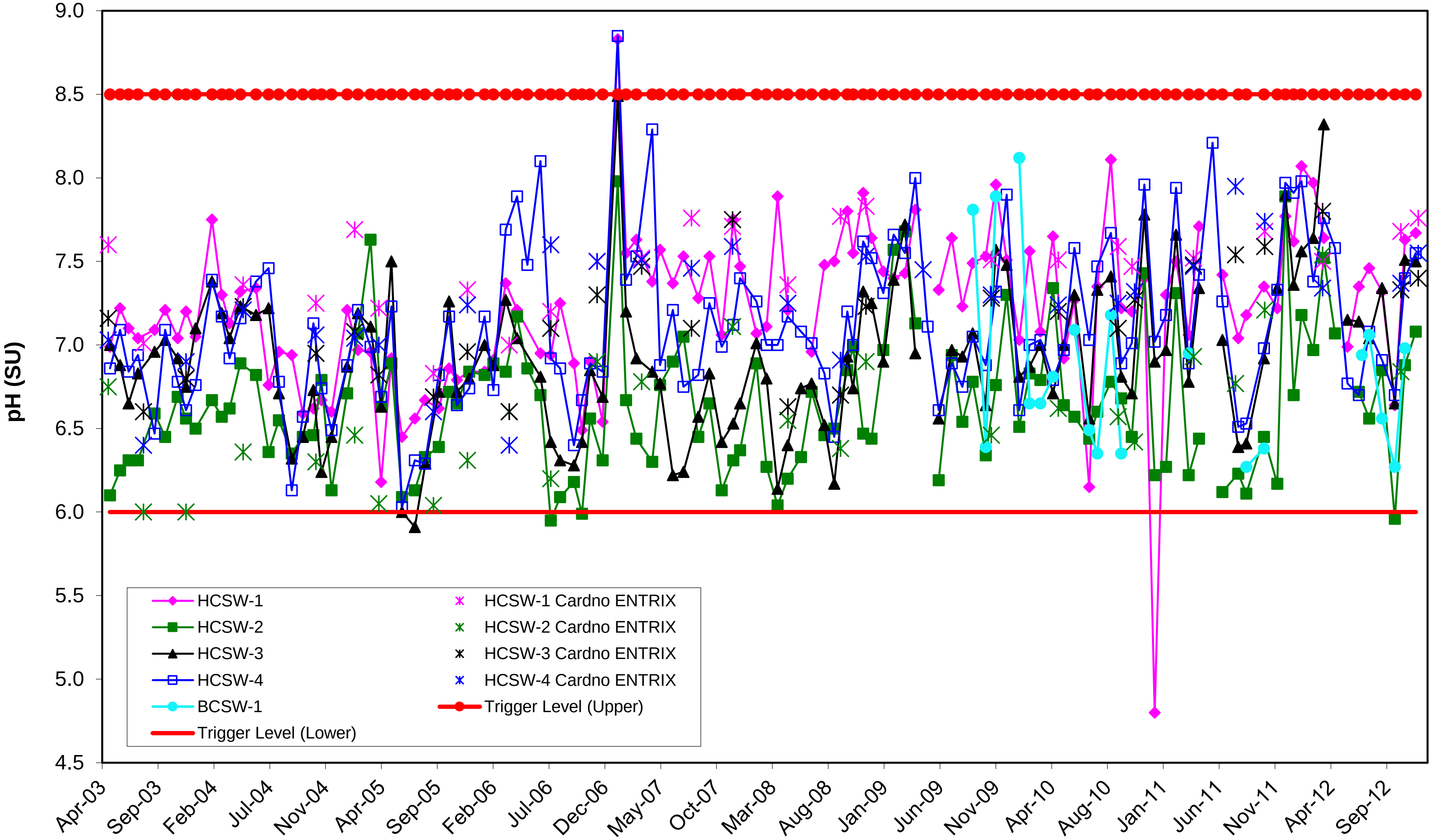
Final Acceptance: November 4, 2009

APPENDIX

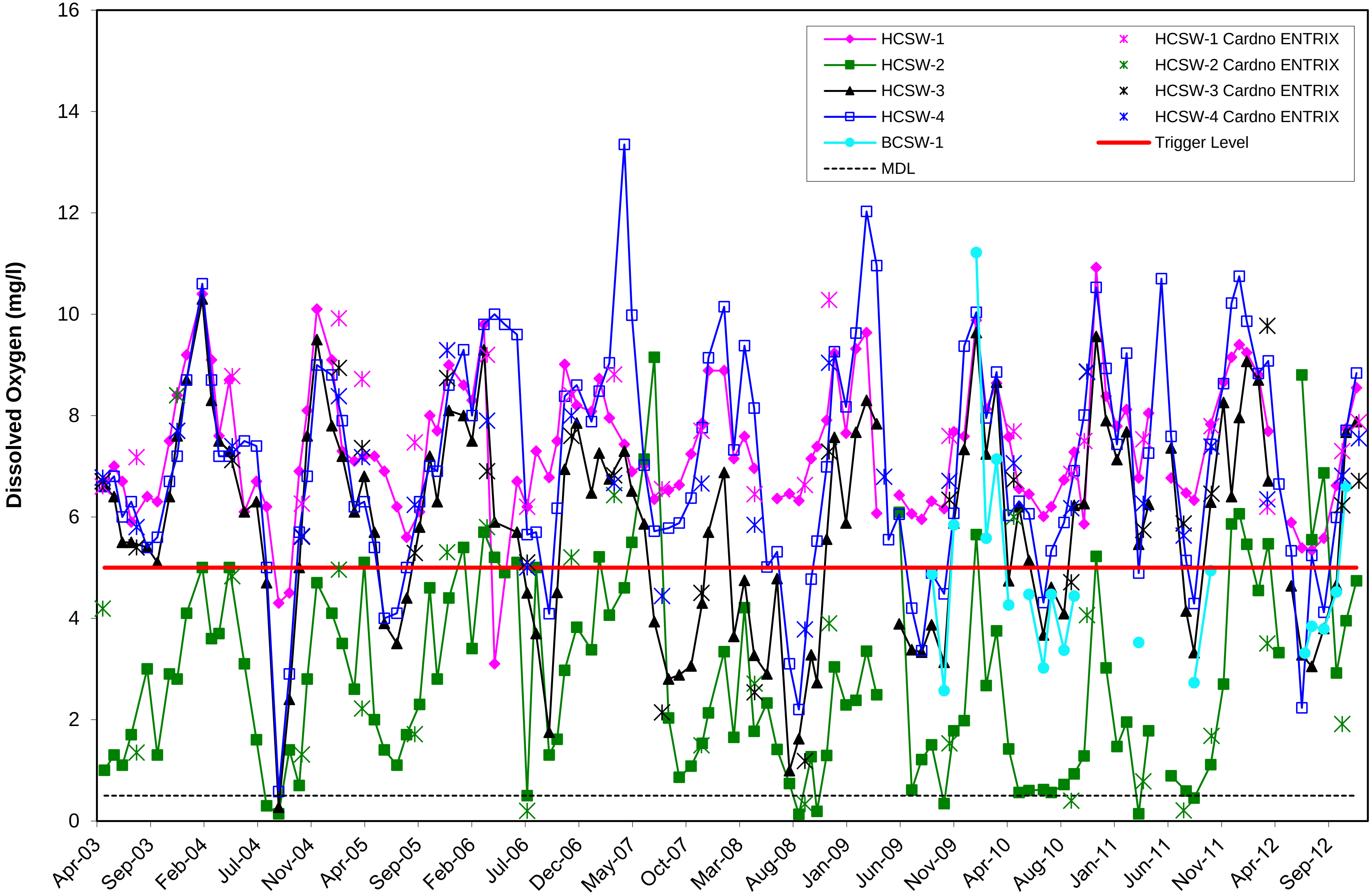
C

HORSE CREEK STEWARDSHIP PROGRAM WATER QUALITY FROM 2003-2012

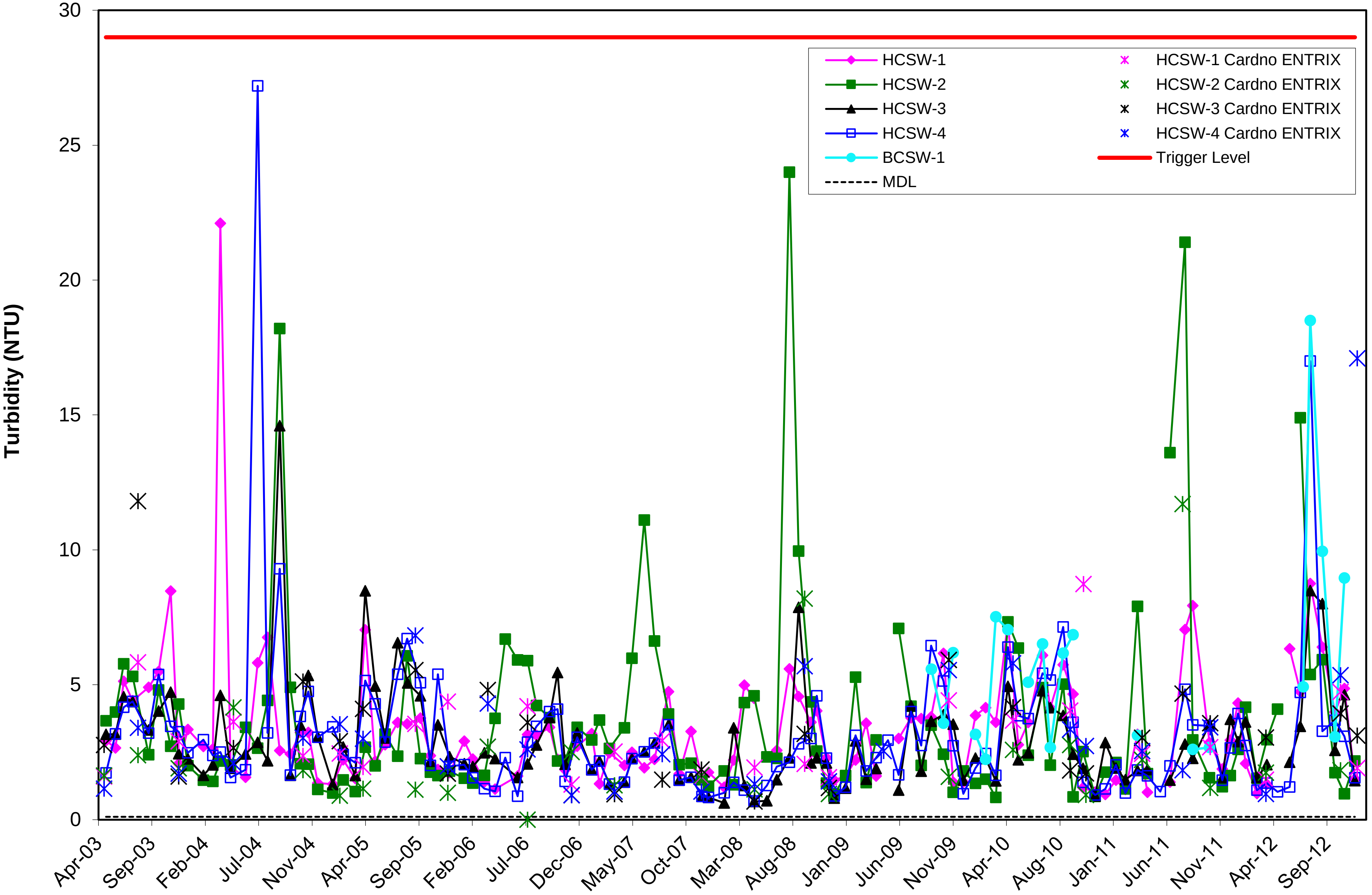
Appendix C
Water Quality from 2003-2012



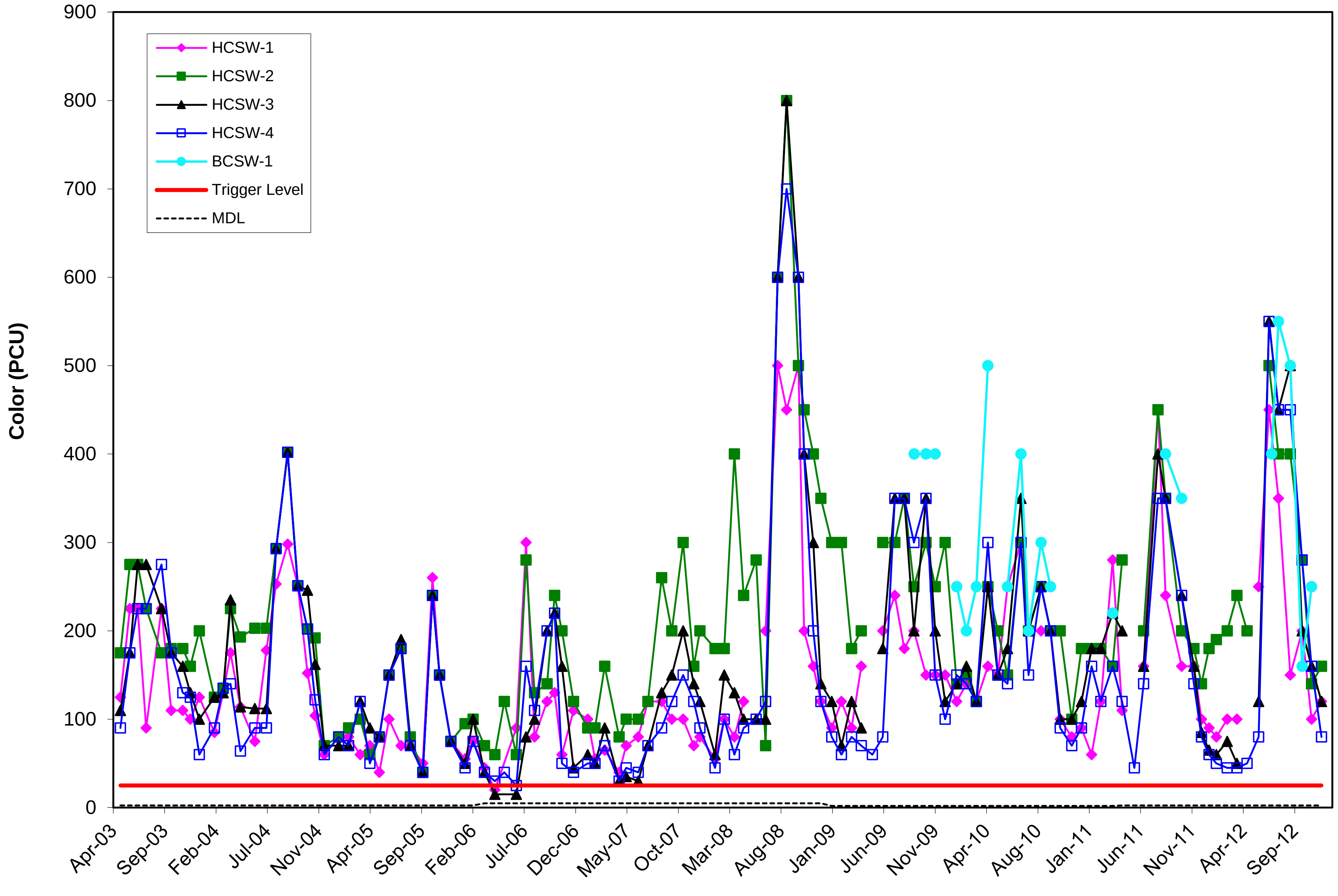
C-1. **Values of pH Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events from 2003-2012. Minimum Detection Limit – 1 su.**



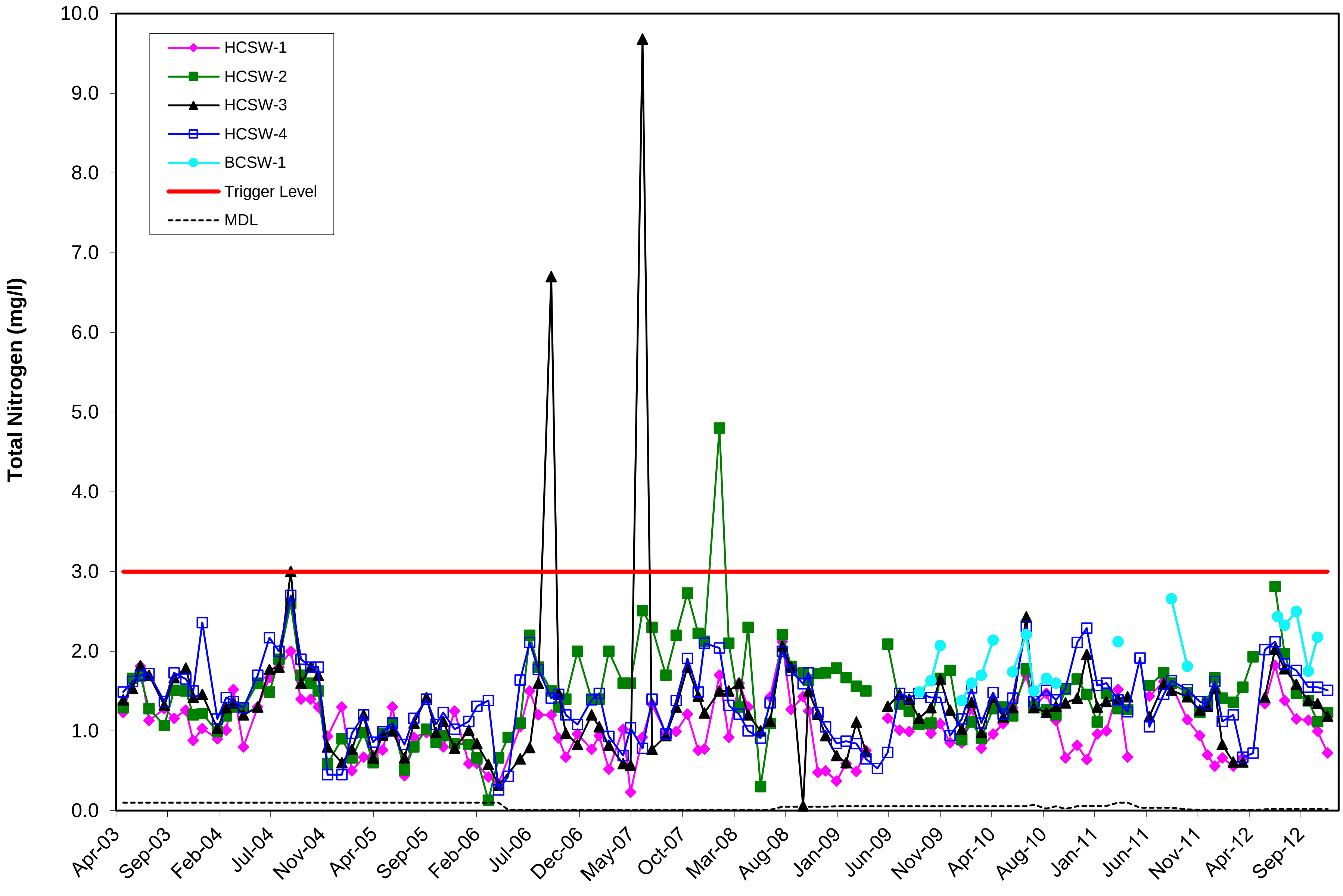
C-2. Dissolved Oxygen Levels Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events from 2003-2012.



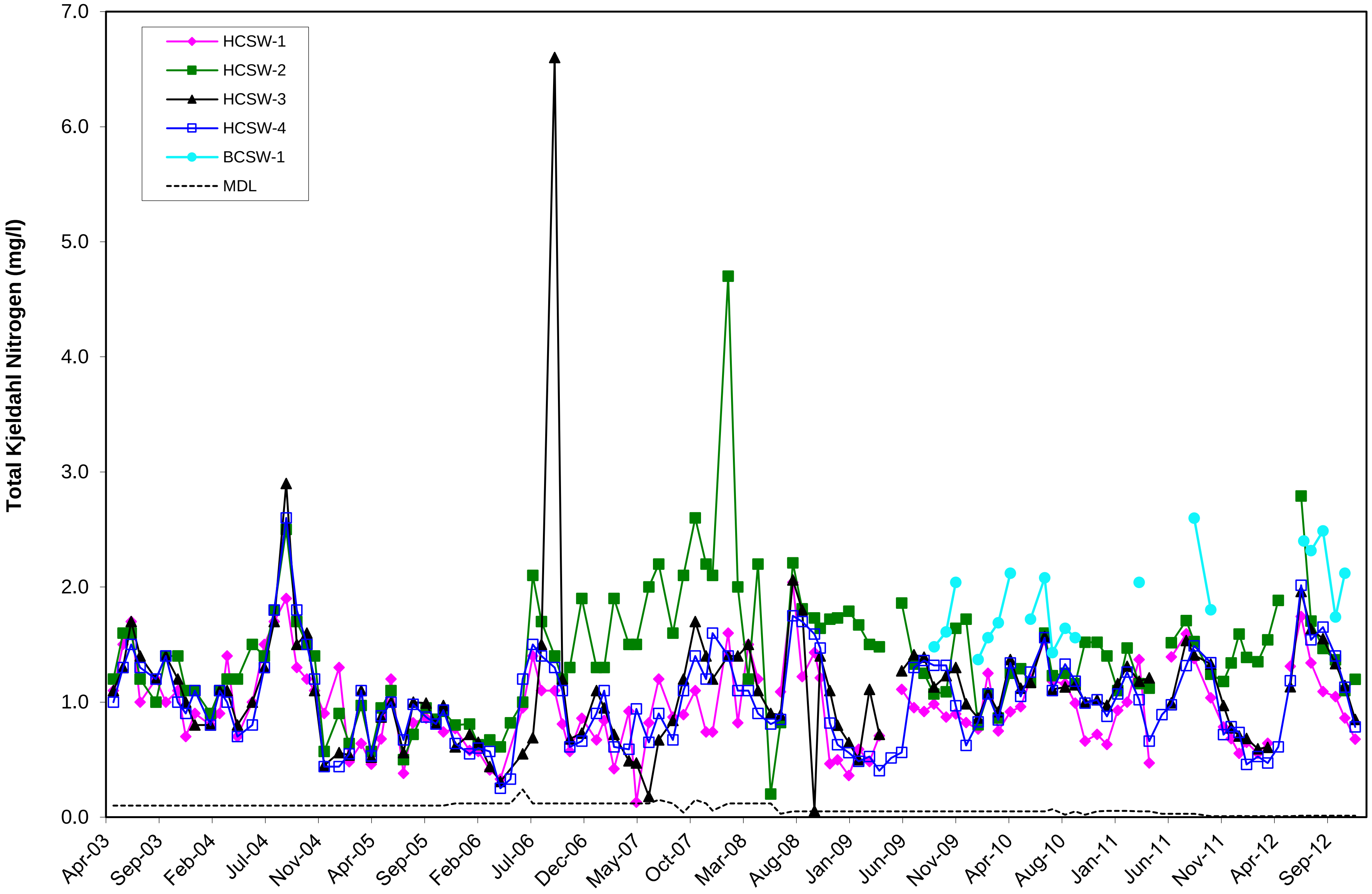
C-3. Turbidity Levels Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events from 2003-2012.



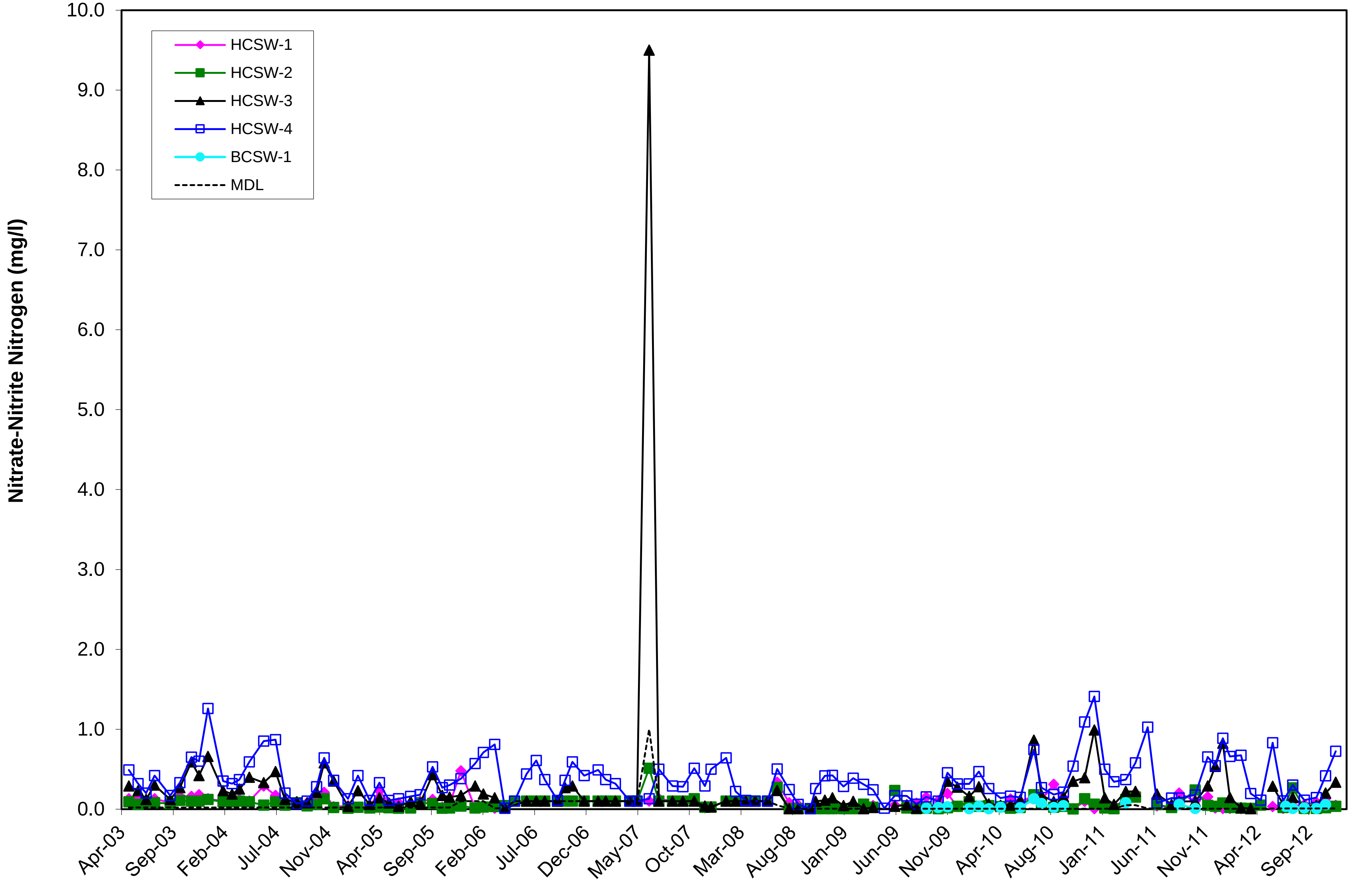
C-4. Color Levels Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



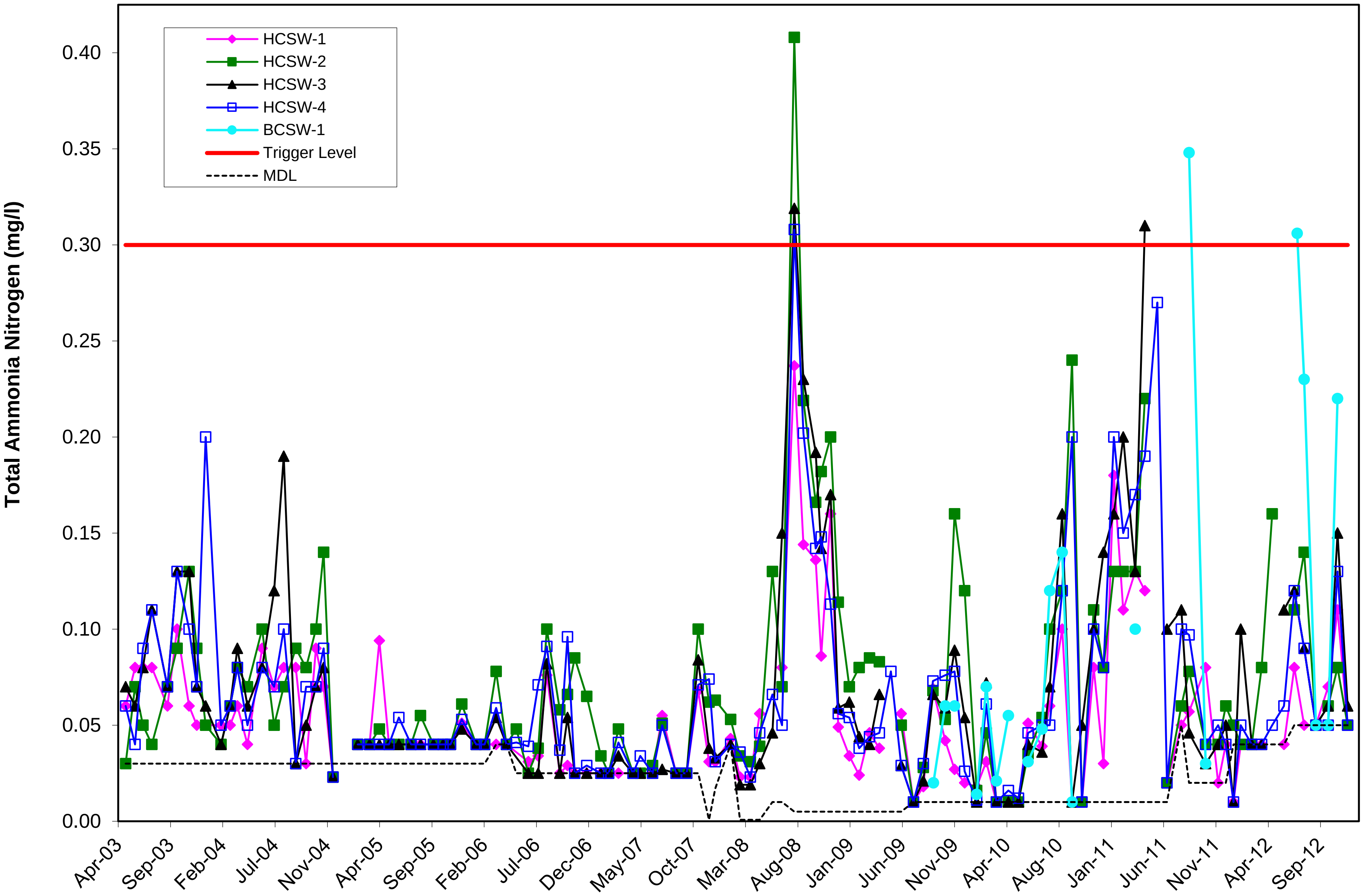
C-5. Total Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



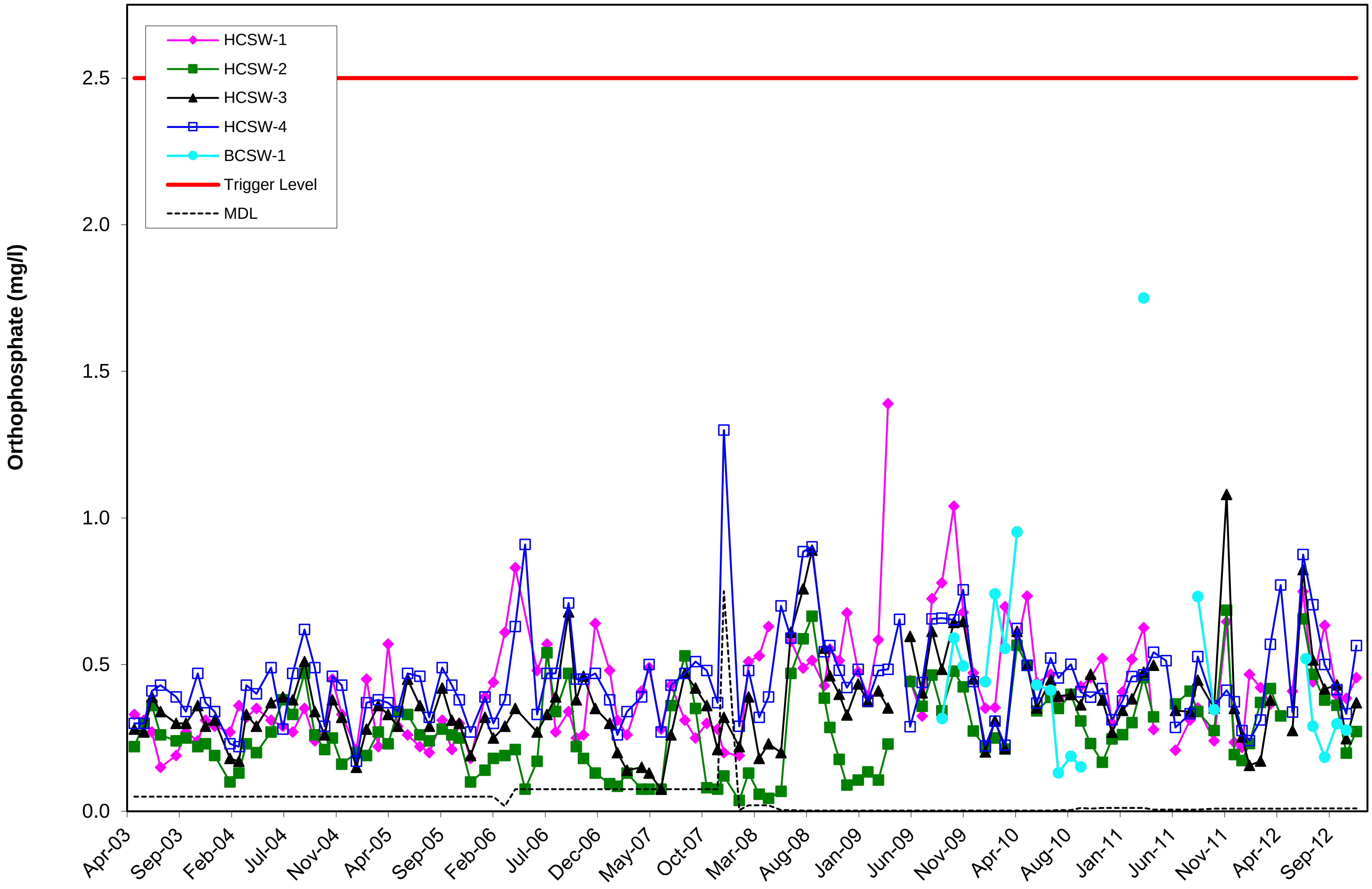
C-6. **Total Kjeldahl Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.**



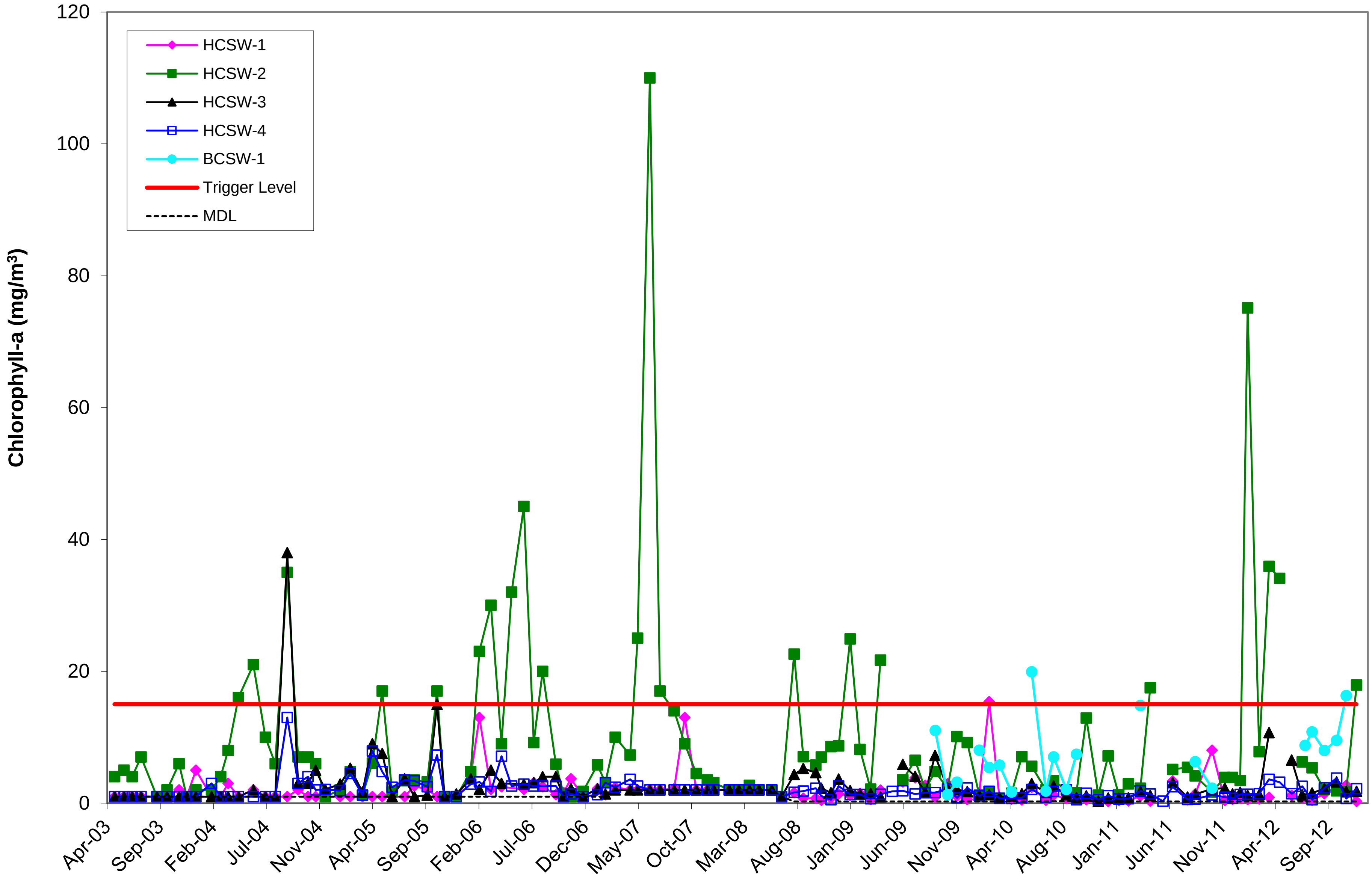
C-7. Nitrate-Nitrite Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



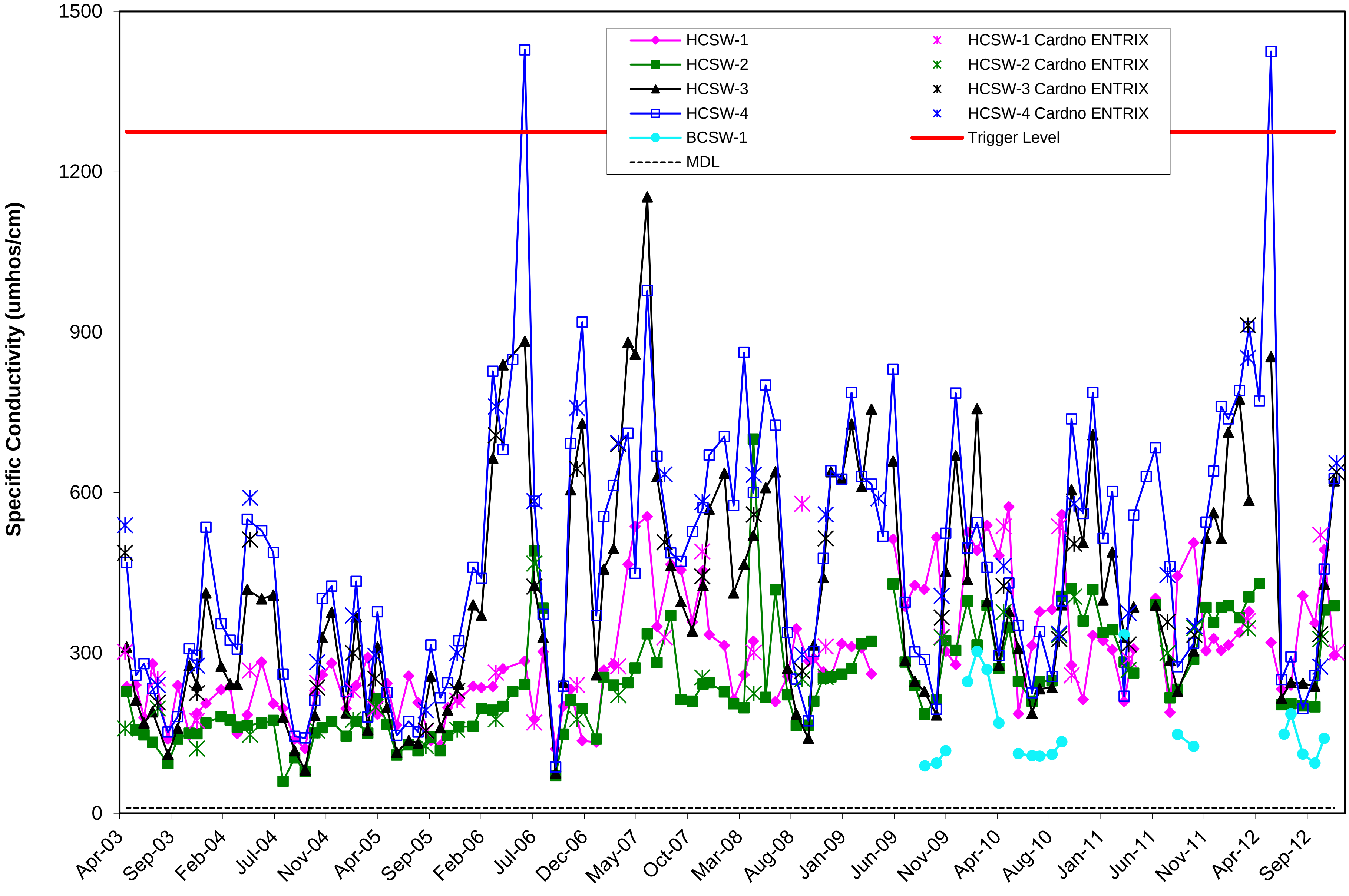
C-8. Total Ammonia Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



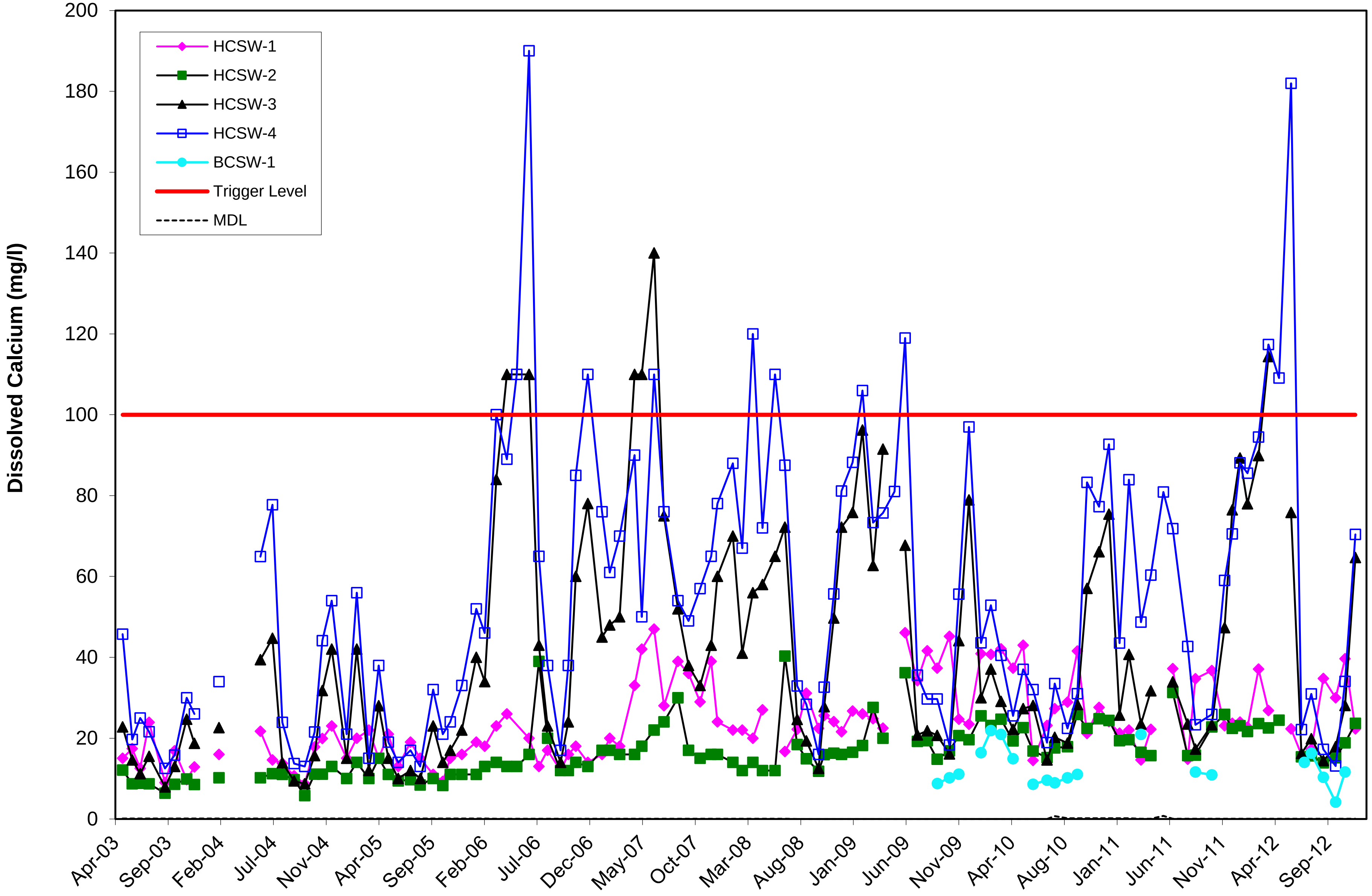
C-9. Orthophosphate Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



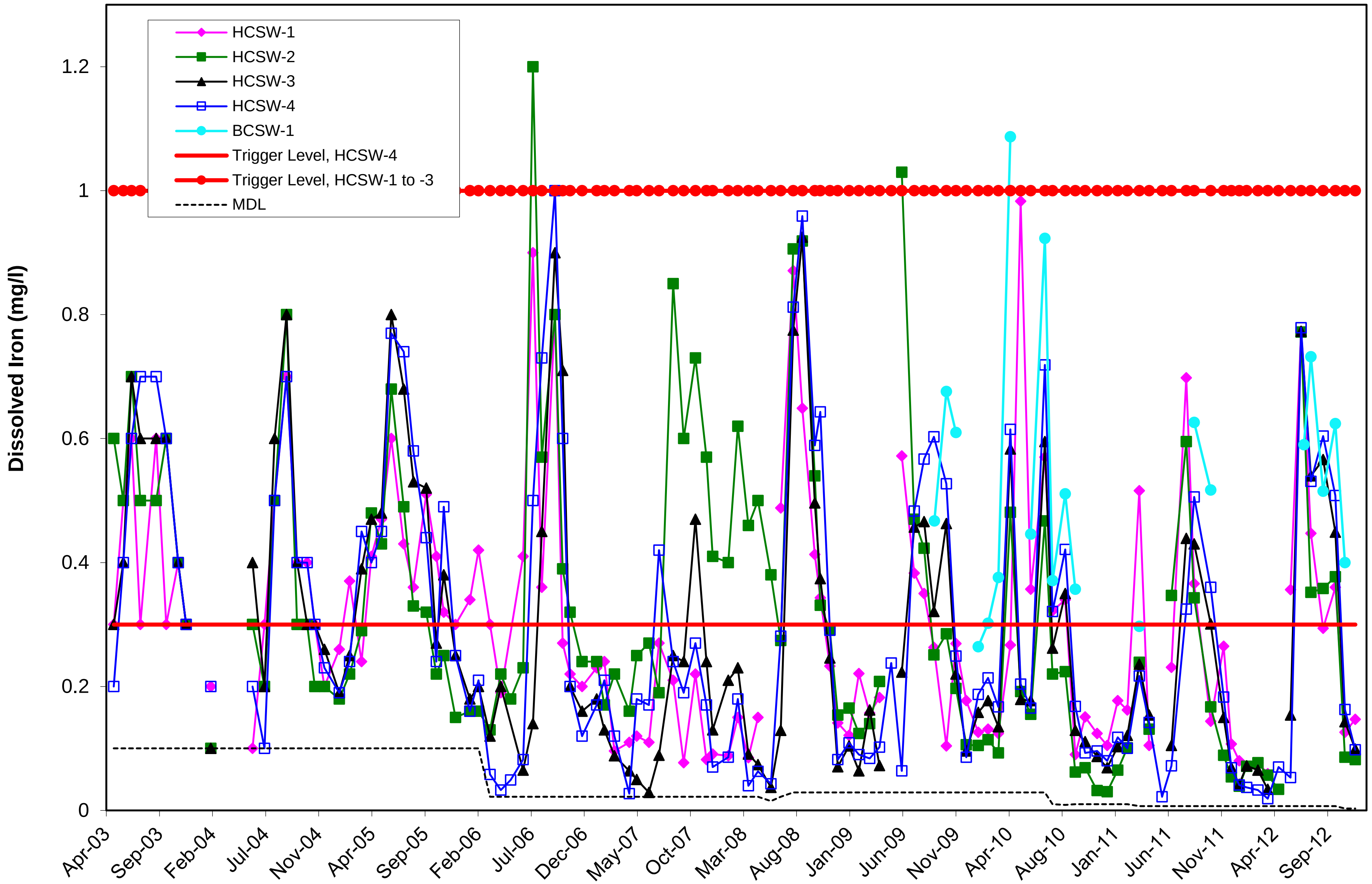
C-10. **Chlorophyll-a Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.**



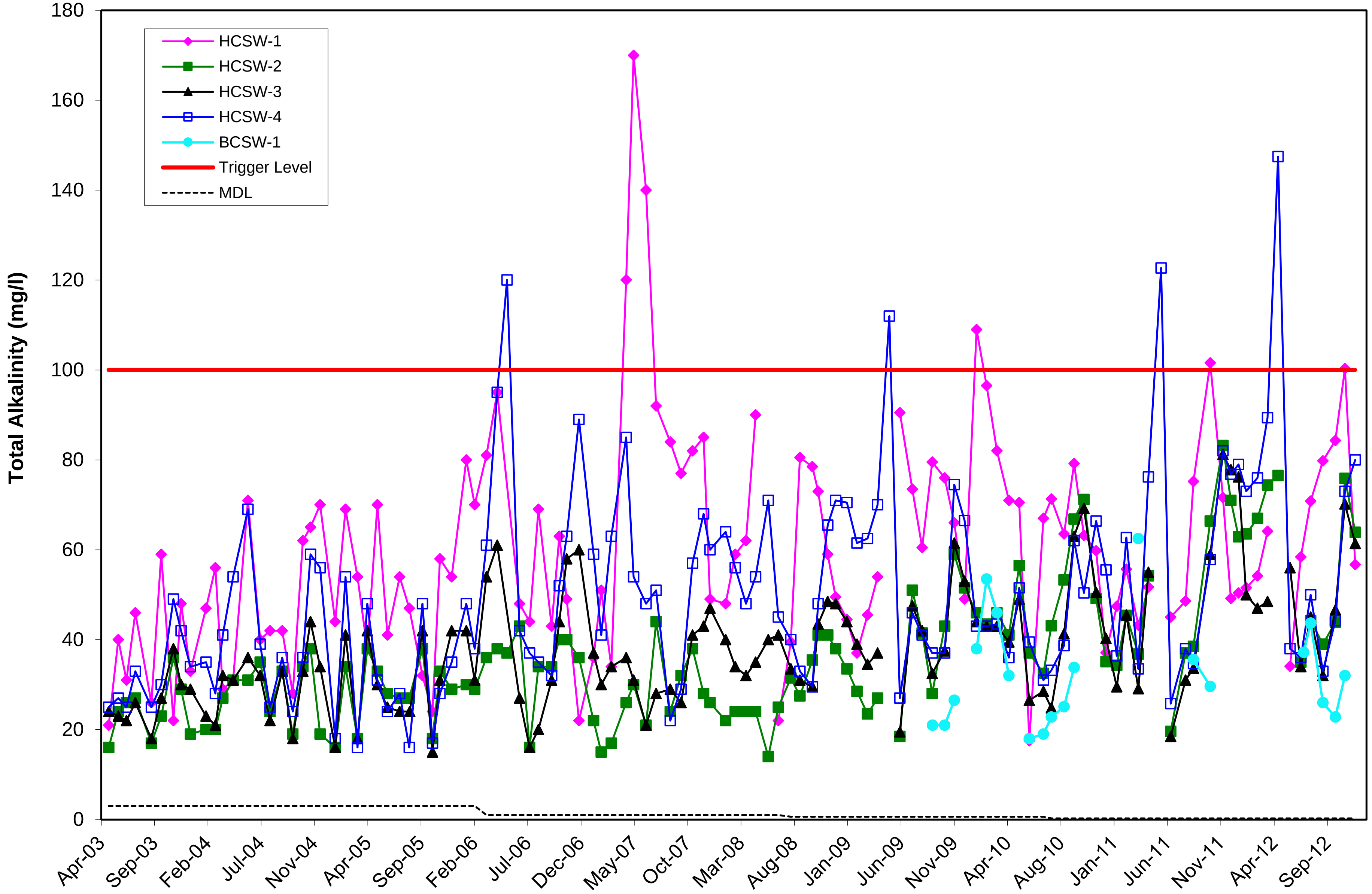
C-11. Levels of Specific Conductivity Obtained During Monthly HCSP Water Quality Sampling and Biological Sampling Events from 2003-2012.



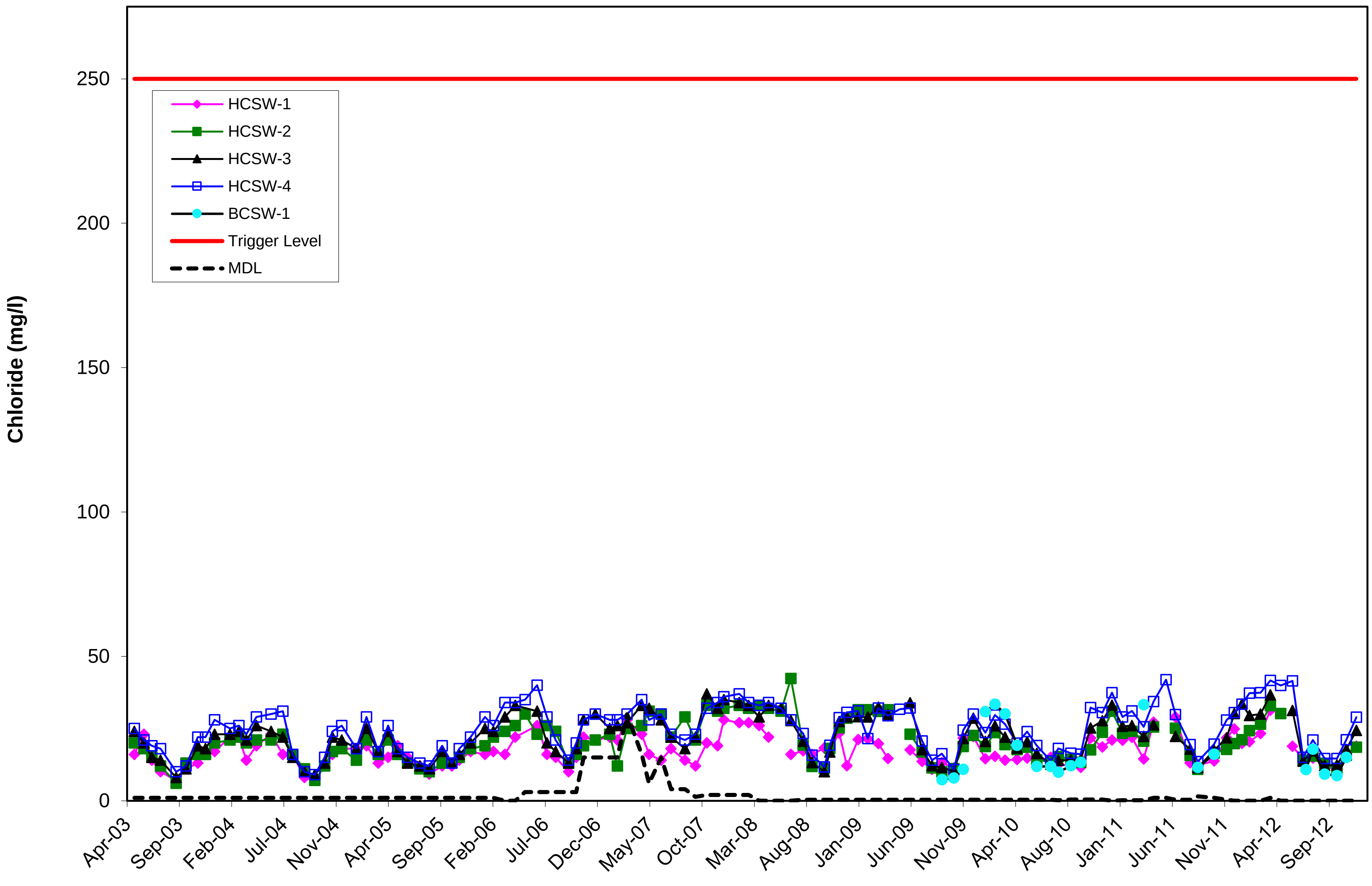
C-12. Dissolved Calcium Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



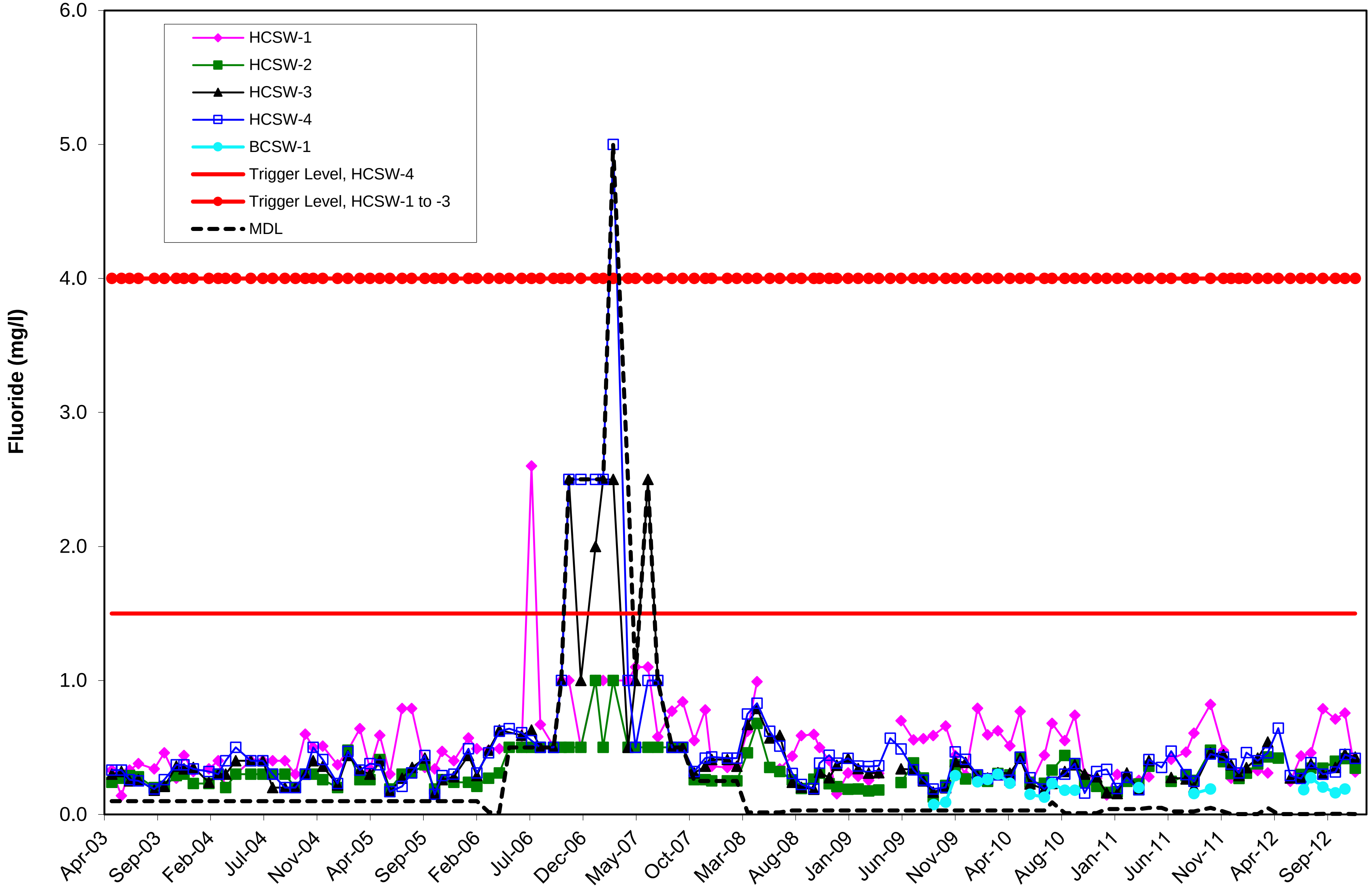
C-13. Dissolved Iron Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



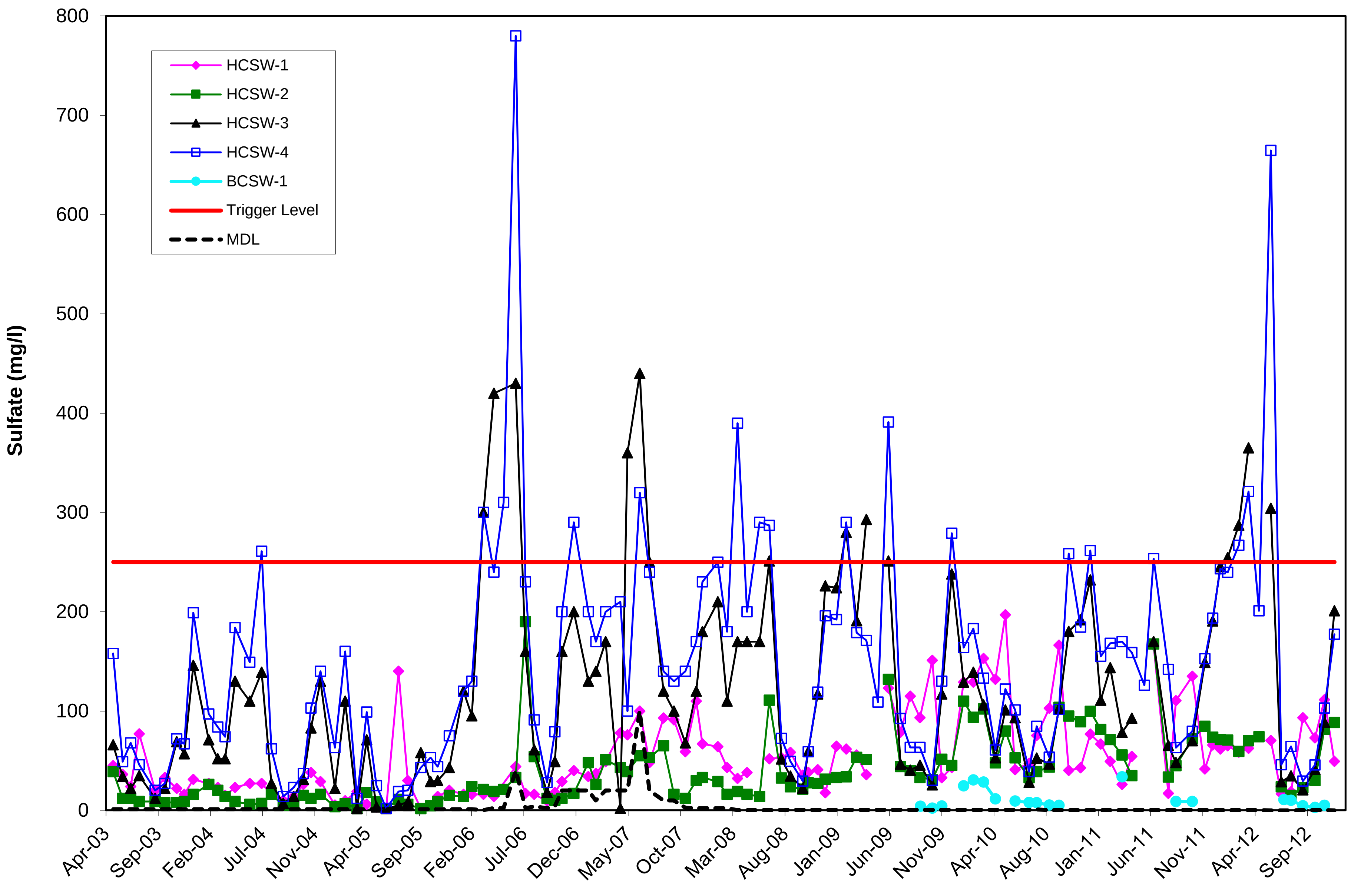
C-14. Total Alkalinity Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



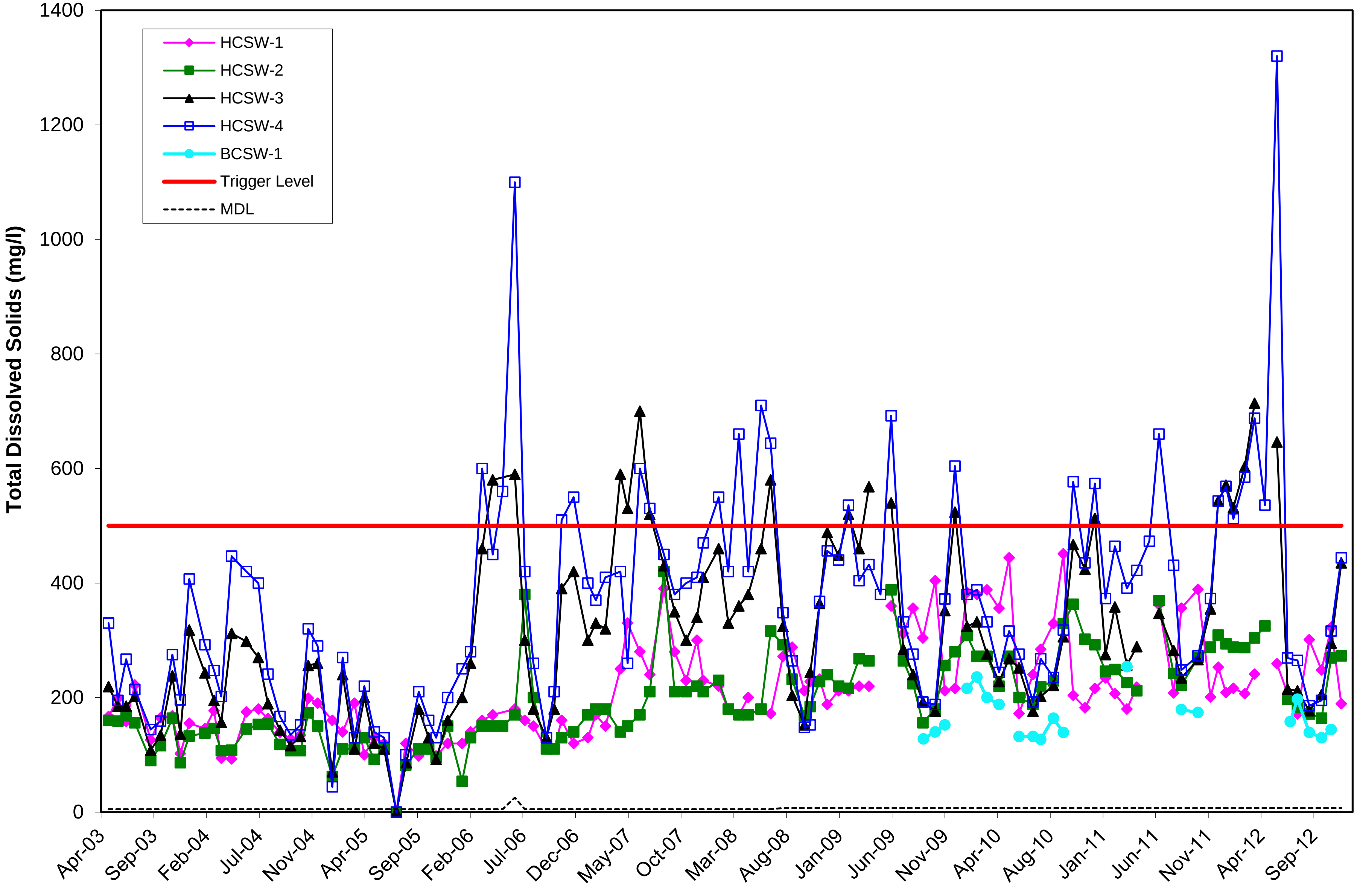
C-15. Chloride Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



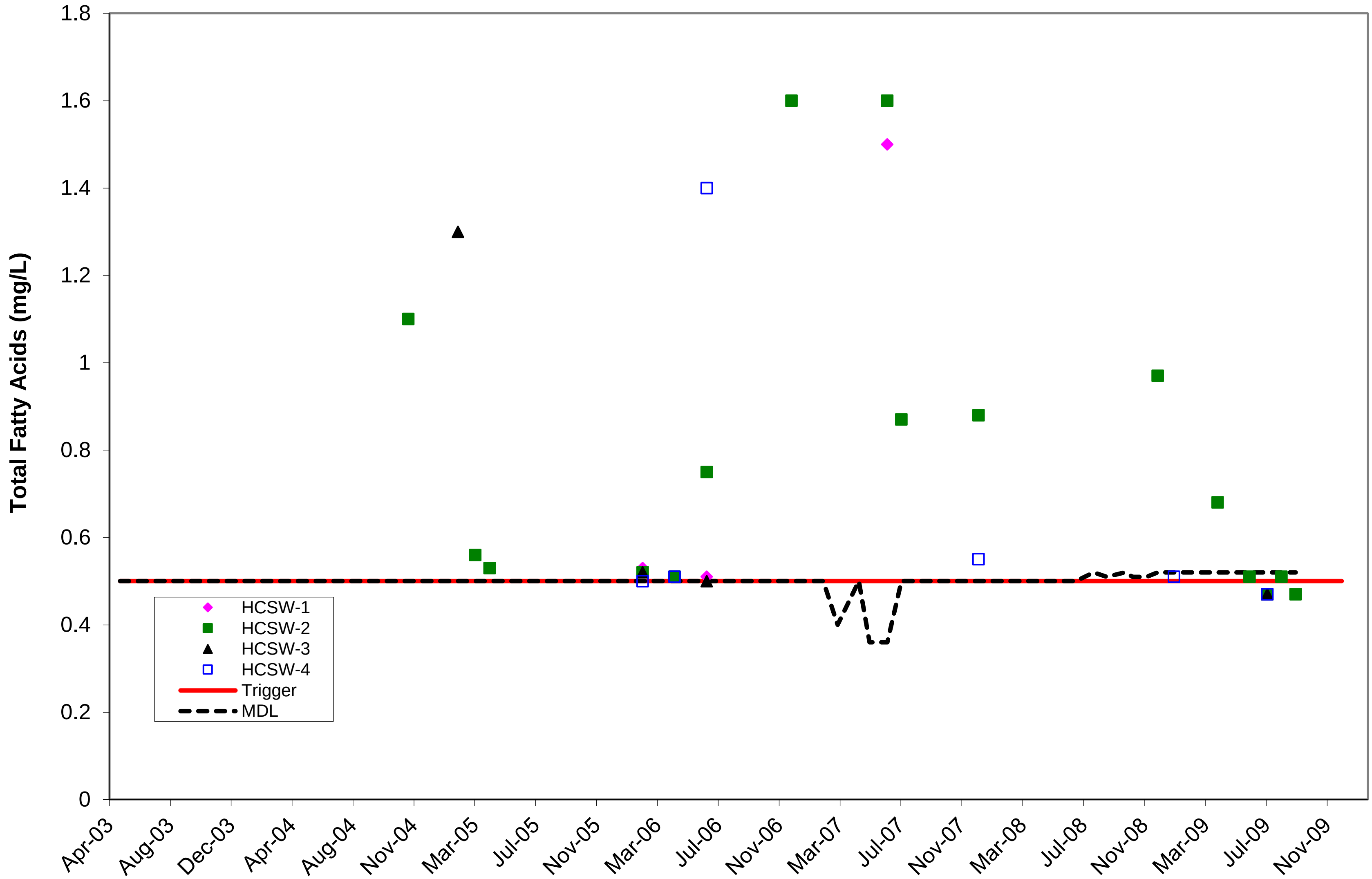
C-16. Fluoride Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



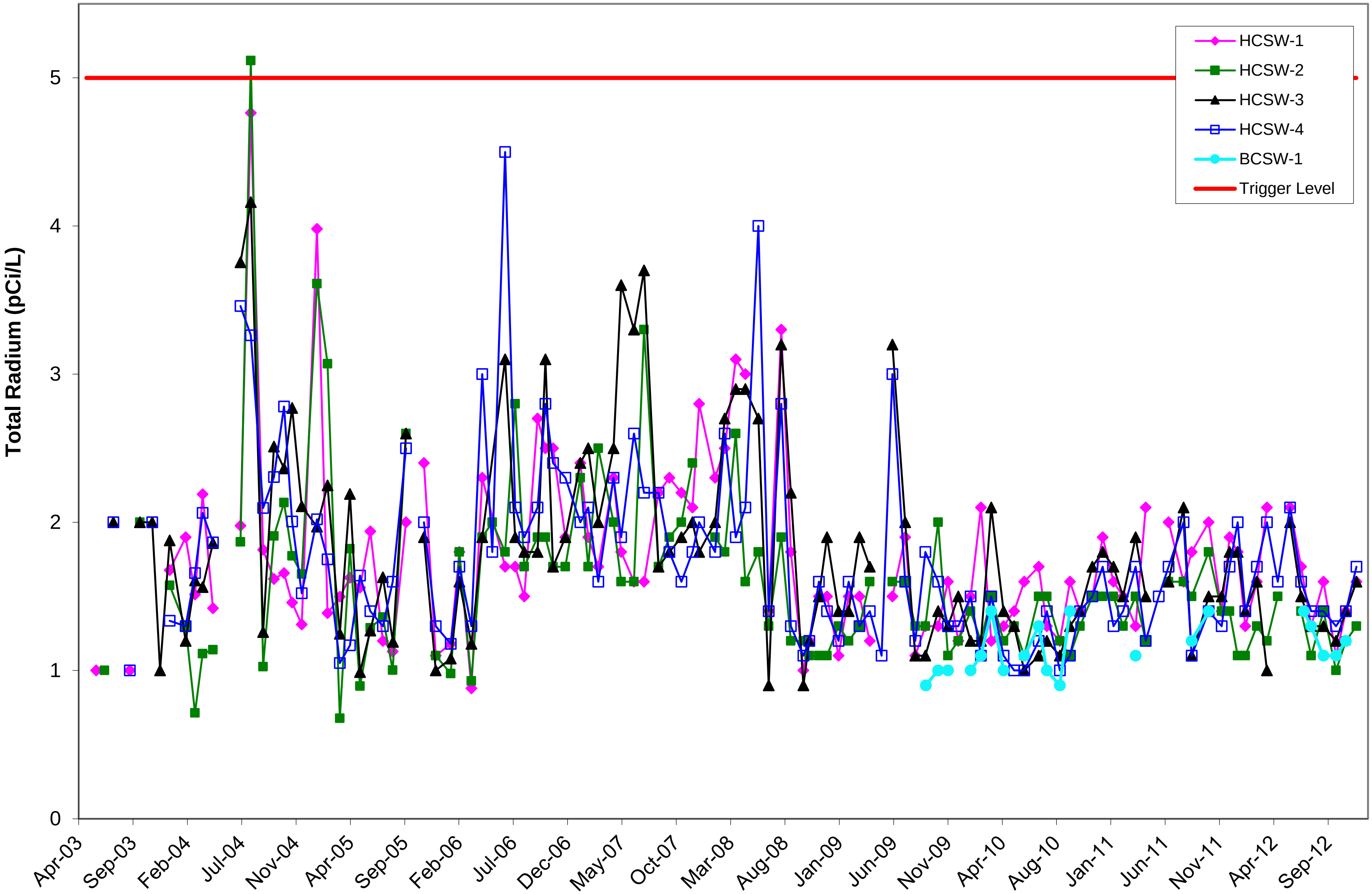
C-17. Sulfate Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.



C-18. **Levels of Total Dissolved Solids Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.**



C-19. Levels of Total Fatty Acids (Above MDL only) Obtained During Monthly HCSP Water Quality Sampling from 2003-2009.



C-20. Levels of Total Radium (Combination of Radium 226 and Radium 228) Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

APPENDIX

D

LITERATURE REVIEW OF STATISTICAL TREND ANALYSIS METHODS

Appendix D
Literature Review of Statistical Trend Analysis Method

The following is a literature review of water quality data trend detection tests, intended to identify the best monotonic trend detection method for use in the Horse Creek Stewardship Program (HCSP). Based on information gleaned from a variety of source, including the USGS (United States Geological Survey), the Seasonal Kendall test was determined to be the best method for use in the HCSP. Because the method needs a minimum of five years of data collection, the 2008 HCSP Annual Report will be the first report to include this analysis. In this (2007) and previous annual reports (2003 – 2006), a variation of this test, the annual median Mann-Kendall, was used on the combined data from several data sources (HCSP, SWFWMD, USGS) for the period 1990 through 2007 to detect possible changes over time. Any changes over time detected using either method may result from a variety of causes, including changes in analytical methods, climatic variation, or anthropogenic causes; an impact assessment may be conducted to determine if the trend is caused by Mosaic mining activities. The following review describes both trend methods that have been or will be used in the HCSP.

Water quality monitoring data exhibits several characteristics that make trend analysis with traditional parametric statistics methods difficult. Water quality datasets often violate the assumptions of parametric statistics, such as the need for independent observations, normal distributions, and constant variance (Berryman et al. 1988, Lettenmaier 1988). In addition, water quality data may be seasonally cyclical or flow-dependent, and datasets may contain missing, censored, or truncated data. Although many methods have been proposed for trend detection, nonparametric methods are the most recommended for detecting trends in water quality data (Berryman et al. 1988, Lettenmaier 1988, Hirsch et al. 1982).

Trend detection methods include graphical methods, time series analysis, parametric statistical tests, and nonparametric statistical tests (Berryman et al. 1988). Graphical methods of trend analysis involve visual interpretation of the data, with no explicit test for trends. This method is often used for exploratory data analysis before other trend detection methods are applied. In time series analysis, a time series is broken down into components (base level, trend, cycle, etc.) using equations. These equations can be combined into a predictive model that can be used to estimate future water quality. Although trends can be modeled using time series analysis, the method does not determine the trend significance, or the chance that the trend is not-random. Statistical tests may be used on the results of the time series analysis to detect trends, but it is considered more appropriate and efficient to use other statistical methods directly. In addition, time series analysis is not appropriate for datasets with irregularly spaced observations or truncated data (observations below method detection limit) (Berryman et al. 1988).

Statistical tests detect trends by applying a rule that the magnitude of the trend is large compared to the variance. Statistical tests may be parametric (based on a normal distribution) or nonparametric. Parametric methods assume that the data is normally distributed, independent, and of constant variance. Although parametric methods are robust against data that violates these assumptions, the power of the test to detect trends is reduced. When these assumptions are violated, as with most water quality data, nonparametric methods are preferred. Because nonparametric methods are based on ranks of observations rather than magnitudes, they can be used on datasets with non-normal distributions or truncated data (Berryman et al. 1988, Lettenmaier 1988, Hirsch et al. 1982). Nonparametric methods can also be adapted for data that is not independent with corrections for seasonality or serial autocorrelation (Berryman et al. 1988, Hirsch et al. 1982, Harcum et al. 1992).

Nonparametric tests may be used to detect monotonic trends, step trends, or multi-step trends (Berryman et al. 1988). Monotonic trends are gradual and unidirectional, but step trends may occur suddenly, be restricted to a limited time period, and may reverse direction over time. Nonparametric methods used to detect step trends include the Mann-Whitney (single step), Kolmogorov-Smirov (single step), and Kruskal-Wallis (multi-step) tests. For each of these tests, the mean ranks before and after a designated time-step are compared, similar to parametric t-tests or ANOVA.

In the absence of *a priori* knowledge of a time-specific potential impact that could affect water quality, monotonic trends are typically the most common trends examined. Nonparametric methods that detect monotonic trends include the Mann-Kendall, Spearman, Cox-Stuart, and Friedman's tests. The Spearman and Kendall are considered the most powerful; these methods detect trends by a significant correlation between the parameter values and time (Berryman et al. 1988). Several of these methods have been adapted for use with seasonally cyclic data; the most commonly used seasonally-adjusted method for water quality trend analysis is the Seasonal Kendall method (Lettenmaier 1988, Hirsch et al. 1982, Harcum et al. 1992, Helsel et al. 2005).

The Seasonal Kendall method is a frequently recommended method for detecting trends in water quality data (Lettenmaier 1988, Hirsch et al. 1982, Harcum et al. 1992, Helsel et al. 2005). The Seasonal Kendall was developed by and is now the method of choice for the United States Geological Survey (USGS) (Hirsch et al. 1982, Helsel et al. 2005). Other agencies that have used the Seasonal Kendall include the United States Environmental Protection Agency, South Florida Water Management District, Departments of Environmental Protection in Virginia and Oregon, Charlotte Harbor National Estuary Program, National Institute of Water and Atmospheric Research, and many universities.

The Seasonal Kendall test determines water quality trends after correcting for seasonality by only comparing values between similar seasons over time (Schertz et al. 1991). The Seasonal Kendall test selects one value for each season (average, median, or subsample) and makes all pair-wise comparisons between time-ordered seasonal values (Harcum et al. 1992). A test statistic (Tau) is computed by comparing the number of times a later value is larger than an earlier value in the data set, and vice-versa (Schertz et al. 1991). Results for the Seasonal Kendall include the magnitude of the statistic Tau, its significance (p), its slope (Sen slope estimator), and the direction of significant trends. The trend (Sen) slope is the median slope of all pairwise comparisons. The direction of this slope (positive or negative) is more resistant to the effects of observations below minimum detection limits and missing data than the magnitude of the slope. The slope is a measure of the monotonic trend; the actual temporal variation may include step trends or trend reversals. If alternate seasons exhibit trends in opposite directions, the Seasonal Kendall test will not detect an overall trend (Lettenmaier 1988).

The power of the Seasonal Kendall test to detect trends in water quality data depends on sample size, season size, significance level, and the magnitude of trend to be detected (Harcum et al. 1992, Hirsch et al. 1982). Collapsing monthly data into quarters or years will reduce the power of the test to detect trends. If monthly data exhibits serial autocorrelation (dependence between adjacent months), however, collapsing is necessary to preserve an accurate significance level (p). Serial autocorrelation may make the actual p value much higher than expected (i.e. $p = 0.15$ instead of $p = 0.05$), leading to a very liberal interpretation of the significance level of potential trends. The loss of power caused by collapsing the data into quarters ceases to matter as sample size increases or the desired trend magnitude increases (Table 1, Harcum et al. 1992). The power difference between monthly and quarterly data disappears in 10-year datasets when the desired trend magnitude is 0.02 units/year, and in 5-year datasets when the trend magnitude is between 0.05 and 0.20 units/year.

Table 1. Power Comparison for Monthly and Quarterly (Median) Data for Five and Ten Years of Data (adapted from figures in Harcum et al. 1992).

Years of Data	Trend slope (units/yr)	Power (Monthly Data)	Power (Quarterly Data)
5	0.002	0.05	0.05
5	0.005	0.09	0.06
5	0.020	0.60	0.31
5	0.050	0.97	0.83
5	0.200	1.00	1.00
5	0.500	1.00	1.00
10	0.002	0.12	0.10
10	0.005	0.45	0.32
10	0.020	0.98	0.95
10	0.050	0.99	0.99
10	0.200	1.00	1.00
10	0.500	1.00	1.00

The USGS recommends at least five years of data with less than five percent truncated observations for the Seasonal Kendall test. Trends detected in datasets with more than five percent of the observations below the method detection limit will have an accurate direction, but the slope magnitude will be a poor estimate (Schertz et al. 1991).

Based on this literature review on tests for water quality data trend detection, the best monotonic trend detection method for use in the Horse Creek Stewardship Program (HCSP) is the Seasonal Kendall. Because the USGS recommends a minimum of five years of data collection before applying the Seasonal Kendall test (Schertz et al. 1991), the 2008 Annual Report will be the first report to include this analysis.

When the Seasonal Kendall test is applied to data in the 2008 HCSP Annual Report, trend detection will be limited by several factors. With only five years of data, the power of the test to detect trends of small magnitude will be limited (Table 1, Harcum et al. 1992, Hirsch et al. 1982). In addition, the monthly data collected as part of the HCSP exhibits serial autocorrelation, meaning that adjacent monthly observations are not independent. Because the dependence in data for some parameters extends to observations made two months apart, collapsing the data into quarterly values is recommended (Harcum et al. 1992). This will reduce the power of the test by an additional margin (Table 1). Finally, data for parameters whose method detection limits have changed several times over the HCSP (fluoride, nitrate+nitrite) will have to be truncated to the highest detection limit, thereby reducing the available data for the test. As a result, trends will be harder to detect, and only the direction of the trend, not the magnitude of the trend, will be valid (Schertz et al. 1991). Despite these limitations, the Seasonal Kendall test is still the most appropriate to detect monotonic trends in HCSP water quality data, once five years of measurements have been collected.

In this (2007) and previous annual reports (2003 – 2006), the dataset collected by the HCSP was not of sufficient length for the Seasonal Kendall analysis. Instead, those reports included a variation of this test, the Mann-Kendall, where the data from several data sources (HCSP, SWFWMD, USGS) were collapsed into annual median values to detect possible changes over time for years 1990 to the present. Although collapsing the data into annual medians results in a loss of power to detect changes, it is a valid method for water quality trend detection (Harcum et al. 1992). The combined data set used in the HCSP reports includes data collected from 1990 to 2007 by FDEP, USGS, SWFWMD, and HCSP with various analytical methods, sampling frequencies, and method detection limits that may bias the results. The annual median

Mann-Kendall was chosen over the Seasonal Kendall as a more conservative approach. All trend analysis methods are heavily influenced by the observations at the beginning and end of a dataset, so the effects of the recent drought years should also be considered when examining potential changes.

Because of all of the potential sources of bias in the combined dataset (changes in methods, different agency sources, different sampling frequencies, climatic variation, etc), a statistically significant Mann-Kendall test may be caused by factors other than anthropogenic sources. Any changes over time detected using either the annual median Mann-Kendall or Seasonal Kendall test should be further examined to determine if the perceived change is caused by a data bias, if its magnitude is ecologically significant, and if the cause of the trend is related to Mosaic mining activities. If warranted, an impact analysis can be performed on statistically significant trends to determine if trends are caused by Mosaic mining activities and if a corrective action by Mosaic is necessary.

Literature Cited

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Schertz, T. L., R. B. Alexander, and D. J. Ohe. 1991. The computer program EStimate TREND (ESTREND), a system for the detection of trends in water-quality data: U.S. Geological Survey Water-Resources Investigations Report 91-4040. 63 p.

APPENDIX E

TAG MEETING SUMMARY

Appendix E
Tag Meeting Summary

Horse Creek Stewardship Program
Technical Advisory Group
Meeting Summary for October 29, 2014

Draft 2011 and 2012 Annual Reports

TAG Panel

Bill Byle	Charlotte County
John Ryan	Sarasota County

Presenters

Kris Robbins	Cardno

Attendees

Jeff Clark	EarthBalance
Santino Provenzano	Mosaic
Subrata Bandyopadhyay	Mosaic
Adam Platt	Mosaic
Ryan Tickles	Mosaic
Sheri Huelster	Cardno

1. Report Overview

Kris Robbins of Cardno provided a technical summary and overview of Program data presented in the 2011 and 2012 HCSP Annual Reports, including a summary of the impact assessment for orthophosphate and specific conductivity.

Similar to the 2010 Annual Report, potential trends were identified for various water quality parameters in both the 2011 and 2012 reports. For four parameters (pH, color, iron, and ammonia), the direction or the magnitude of the potential trends are not adverse. Five of the parameters (calcium, alkalinity, fluoride, sulfate, and TDS) are related to specific conductivity, so the discussion focused on specific conductivity as the best surrogate for all dissolved ions showing potential trends. The majority of the trend discussion focused on orthophosphate and specific conductivity changes over time.

A Seasonal and Annual Kendall Tau analysis of the orthophosphate data indicates that there is no significant trend in concentrations over the longer time period. Specific conductivity still showed an increasing trend when additional years were included. Background information from Mosaic on mining operations that have discharged through Horse Creek NPDES outfalls in the past offered some possible insight into visible changes in these two parameters over the last 10-15 years.

2. Action Items:

- Cardno will contact the Authority for rain gauge data collected near their facility on the Peace River.
- Cardno will add a graphic to the 2011 and all subsequent reports showing more historical flow in Horse Creek extending beyond the HCSP time period (back to at least 1978 to match the double mass curve).
- Cardno will add more discussion of the double mass curve to the 2011 and all subsequent reports.
- Cardno will add a Trend Summary Table to the 2011 Annual report and all subsequent reports instead of only including in Appendix I.
- In the 2013 Annual Report, Cardno will add an appendix that lists any erroneous data and remove problematic data from Appendix C graphs. Any future erroneous data will be added to the appendix as the program continues.
- For the 2013 Annual Report TAG Meeting, Cardno will include a discussion on the DO Saturation standard and implications in the main body of the report.
- In the 2014 annual report, Cardno will add a discussion on NNC standards and implications in the main body of the report.
- In the 2014 annual report, Cardno will add a discussion on legacy CF Industries operations in the Horse Creek Basin.
- TAG members will get any additional comments on the 2011 and 2012 reports to the Authority by November 14, 2013.
- Cardno will provide a Word document of all reviewers' questions/comments and responses to the Authority for transmittal to TAG members for the 2011 and 2012 reports.

3. Timeline for 2011 and 2012 Annual Reports

Mosaic and Cardno believe that the 2011 and 2012 final reports could be sent to the Authority by December 2014 or January 2015.

4. Timeline for 2013 Annual Report

Mosaic and Cardno believe that the 2013 draft report could be sent to the Authority by February or March 2015.

APPENDIX

F

SUMMARY OF TRIGGER EXCEEDANCES FROM 2003-2012

Appendix F
Summary of Trigger Exceedances from 2003-2012

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at State Road 64	HCSW-1	1/23/2007	pH (SU)	8.83	8.5
Horse Creek at State Road 64	HCSW-1	1/4/2011	pH (SU)	4.8	6
Horse Creek at Goose Pond Road	HCSW-2	7/27/2006	pH (SU)	5.95	6
Horse Creek at Goose Pond Road	HCSW-2	10/19/2006	pH (SU)	5.99	6
Horse Creek at Goose Pond Road	HCSW-2	10/10/2012	pH (SU)	5.96	6
Horse Creek at State Road 70	HCSW-3	7/27/2005	pH (SU)	5.9	6
Horse Creek at State Road 72	HCSW-4	1/23/2007	pH (SU)	8.85	8.5
Horse Creek at State Road 72	HCSW-4	6/5/2012	Specific Conductance (µmhos/cm)	1,425	1,275
Horse Creek at State Road 64	HCSW-1	8/30/2004	Dissolved Oxygen (mg/l)	4.3	5
Horse Creek at State Road 64	HCSW-1	9/29/2004	Dissolved Oxygen (mg/l)	4.5	5
Horse Creek at State Road 64	HCSW-1	4/27/2006	Dissolved Oxygen (mg/l)	3.1	5
Horse Creek at Goose Pond Road	HCSW-2	4/30/2003	Dissolved Oxygen (mg/l)	1	5
Horse Creek at Goose Pond Road	HCSW-2	5/27/2003	Dissolved Oxygen (mg/l)	1.3	5
Horse Creek at Goose Pond Road	HCSW-2	6/19/2003	Dissolved Oxygen (mg/l)	1.1	5
Horse Creek at Goose Pond Road	HCSW-2	7/14/2003	Dissolved Oxygen (mg/l)	1.7	5
Horse Creek at Goose Pond Road	HCSW-2	8/28/2003	Dissolved Oxygen (mg/l)	3	5
Horse Creek at Goose Pond Road	HCSW-2	9/25/2003	Dissolved Oxygen (mg/l)	1.3	5
Horse Creek at Goose Pond Road	HCSW-2	10/29/2003	Dissolved Oxygen (mg/l)	2.9	5
Horse Creek at Goose Pond Road	HCSW-2	11/20/2003	Dissolved Oxygen (mg/l)	2.8	5
Horse Creek at Goose Pond Road	HCSW-2	12/16/2003	Dissolved Oxygen (mg/l)	4.1	5
Horse Creek at Goose Pond Road	HCSW-2	1/29/2004	Dissolved Oxygen (mg/l)	5	5
Horse Creek at Goose Pond Road	HCSW-2	2/24/2004	Dissolved Oxygen (mg/l)	3.6	5
Horse Creek at Goose Pond Road	HCSW-2	3/16/2004	Dissolved Oxygen (mg/l)	3.7	5
Horse Creek at Goose Pond Road	HCSW-2	5/26/2004	Dissolved Oxygen (mg/l)	3.1	5
Horse Creek at Goose Pond Road	HCSW-2	6/29/2004	Dissolved Oxygen (mg/l)	1.6	5
Horse Creek at Goose Pond Road	HCSW-2	7/27/2004	Dissolved Oxygen (mg/l)	0.3	5
Horse Creek at Goose Pond Road	HCSW-2	8/30/2004	Dissolved Oxygen (mg/l)	0.14	5
Horse Creek at Goose Pond Road	HCSW-2	9/29/2004	Dissolved Oxygen (mg/l)	1.4	5
Horse Creek at Goose Pond Road	HCSW-2	10/27/2004	Dissolved Oxygen (mg/l)	0.7	5
Horse Creek at Goose Pond Road	HCSW-2	11/18/2004	Dissolved Oxygen (mg/l)	2.8	5
Horse Creek at Goose Pond Road	HCSW-2	12/15/2004	Dissolved Oxygen (mg/l)	4.7	5
Horse Creek at Goose Pond Road	HCSW-2	1/26/2005	Dissolved Oxygen (mg/l)	4.1	5
Horse Creek at Goose Pond Road	HCSW-2	2/24/2005	Dissolved Oxygen (mg/l)	3.5	5
Horse Creek at Goose Pond Road	HCSW-2	3/30/2005	Dissolved Oxygen (mg/l)	2.6	5
Horse Creek at Goose Pond Road	HCSW-2	5/27/2005	Dissolved Oxygen (mg/l)	2	5
Horse Creek at Goose Pond Road	HCSW-2	6/22/2005	Dissolved Oxygen (mg/l)	1.4	5
Horse Creek at Goose Pond Road	HCSW-2	7/27/2005	Dissolved Oxygen (mg/l)	1.1	5
Horse Creek at Goose Pond Road	HCSW-2	8/23/2005	Dissolved Oxygen (mg/l)	1.7	5
Horse Creek at Goose Pond Road	HCSW-2	9/29/2005	Dissolved Oxygen (mg/l)	2.3	5
Horse Creek at Goose Pond Road	HCSW-2	10/27/2005	Dissolved Oxygen (mg/l)	4.6	5

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at Goose Pond Road	HCSW-2	11/17/2005	Dissolved Oxygen (mg/l)	2.8	5
Horse Creek at Goose Pond Road	HCSW-2	12/20/2005	Dissolved Oxygen (mg/l)	4.4	5
Horse Creek at Goose Pond Road	HCSW-2	2/23/2006	Dissolved Oxygen (mg/l)	3.4	5
Horse Creek at Goose Pond Road	HCSW-2	5/25/2006	Dissolved Oxygen (mg/l)	4.9	5
Horse Creek at Goose Pond Road	HCSW-2	7/27/2006	Dissolved Oxygen (mg/l)	0.5	5
Horse Creek at Goose Pond Road	HCSW-2	9/27/2006	Dissolved Oxygen (mg/l)	1.3	5
Horse Creek at Goose Pond Road	HCSW-2	10/19/2006	Dissolved Oxygen (mg/l)	1.6	5
Horse Creek at Goose Pond Road	HCSW-2	11/9/2006	Dissolved Oxygen (mg/l)	2.9	5
Horse Creek at Goose Pond Road	HCSW-2	12/13/2006	Dissolved Oxygen (mg/l)	3.8	5
Horse Creek at Goose Pond Road	HCSW-2	1/23/2007	Dissolved Oxygen (mg/l)	3.38	5
Horse Creek at Goose Pond Road	HCSW-2	3/14/2007	Dissolved Oxygen (mg/l)	4.06	5
Horse Creek at Goose Pond Road	HCSW-2	4/25/2007	Dissolved Oxygen (mg/l)	4.6	5
Horse Creek at Goose Pond Road	HCSW-2	8/27/2007	Dissolved Oxygen (mg/l)	2.03	5
Horse Creek at Goose Pond Road	HCSW-2	9/26/2007	Dissolved Oxygen (mg/l)	0.86	5
Horse Creek at Goose Pond Road	HCSW-2	10/29/2007	Dissolved Oxygen (mg/l)	1.08	5
Horse Creek at Goose Pond Road	HCSW-2	11/29/2007	Dissolved Oxygen (mg/l)	1.53	5
Horse Creek at Goose Pond Road	HCSW-2	12/17/2007	Dissolved Oxygen (mg/l)	2.13	5
Horse Creek at Goose Pond Road	HCSW-2	1/30/2008	Dissolved Oxygen (mg/l)	3.34	5
Horse Creek at Goose Pond Road	HCSW-2	2/26/2008	Dissolved Oxygen (mg/l)	1.65	5
Horse Creek at Goose Pond Road	HCSW-2	3/27/2008	Dissolved Oxygen (mg/l)	4.21	5
Horse Creek at Goose Pond Road	HCSW-2	4/23/2008	Dissolved Oxygen (mg/l)	1.77	5
Horse Creek at Goose Pond Road	HCSW-2	5/29/2008	Dissolved Oxygen (mg/l)	2.33	5
Horse Creek at Goose Pond Road	HCSW-2	6/26/2008	Dissolved Oxygen (mg/l)	1.41	5
Horse Creek at Goose Pond Road	HCSW-2	7/31/2008	Dissolved Oxygen (mg/l)	0.74	5
Horse Creek at Goose Pond Road	HCSW-2	8/26/2008	Dissolved Oxygen (mg/l)	0.13	5
Horse Creek at Goose Pond Road	HCSW-2	9/30/2008	Dissolved Oxygen (mg/l)	1.27	5
Horse Creek at Goose Pond Road	HCSW-2	10/16/2008	Dissolved Oxygen (mg/l)	0.19	5
Horse Creek at Goose Pond Road	HCSW-2	11/12/2008	Dissolved Oxygen (mg/l)	1.29	5
Horse Creek at Goose Pond Road	HCSW-2	12/4/2008	Dissolved Oxygen (mg/l)	3.04	5
Horse Creek at Goose Pond Road	HCSW-2	1/5/2009	Dissolved Oxygen (mg/l)	2.29	5
Horse Creek at Goose Pond Road	HCSW-2	2/2/2009	Dissolved Oxygen (mg/l)	2.38	5
Horse Creek at Goose Pond Road	HCSW-2	3/4/2009	Dissolved Oxygen (mg/l)	3.35	5
Horse Creek at Goose Pond Road	HCSW-2	4/1/2009	Dissolved Oxygen (mg/l)	2.49	5
Horse Creek at Goose Pond Road	HCSW-2	7/8/2009	Dissolved Oxygen (mg/l)	0.61	5
Horse Creek at Goose Pond Road	HCSW-2	8/5/2009	Dissolved Oxygen (mg/l)	1.21	5
Horse Creek at Goose Pond Road	HCSW-2	9/2/2009	Dissolved Oxygen (mg/l)	1.5	5
Horse Creek at Goose Pond Road	HCSW-2	10/7/2009	Dissolved Oxygen (mg/l)	0.34	5
Horse Creek at Goose Pond Road	HCSW-2	11/3/2009	Dissolved Oxygen (mg/l)	1.78	5
Horse Creek at Goose Pond Road	HCSW-2	12/2/2009	Dissolved Oxygen (mg/l)	1.98	5
Horse Creek at Goose Pond Road	HCSW-2	2/2/2010	Dissolved Oxygen (mg/l)	2.67	5
Horse Creek at Goose Pond Road	HCSW-2	3/3/2010	Dissolved Oxygen (mg/l)	3.75	5
Horse Creek at Goose Pond Road	HCSW-2	4/6/2010	Dissolved Oxygen (mg/l)	1.42	5
Horse Creek at Goose Pond Road	HCSW-2	5/5/2010	Dissolved Oxygen (mg/l)	0.56	5
Horse Creek at Goose Pond Road	HCSW-2	6/2/2010	Dissolved Oxygen (mg/l)	0.6	5
Horse Creek at Goose Pond Road	HCSW-2	7/12/2010	Dissolved Oxygen (mg/l)	0.62	5
Horse Creek at Goose Pond Road	HCSW-2	8/3/2010	Dissolved Oxygen (mg/l)	0.56	5
Horse Creek at Goose Pond Road	HCSW-2	9/8/2010	Dissolved Oxygen (mg/l)	0.72	5
Horse Creek at Goose Pond Road	HCSW-2	10/6/2010	Dissolved Oxygen (mg/l)	0.93	5
Horse Creek at Goose Pond Road	HCSW-2	11/3/2010	Dissolved Oxygen (mg/l)	1.28	5

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at Goose Pond Road	HCSW-2	1/4/2011	Dissolved Oxygen (mg/l)	3.02	5
Horse Creek at Goose Pond Road	HCSW-2	2/3/2011	Dissolved Oxygen (mg/l)	1.47	5
Horse Creek at Goose Pond Road	HCSW-2	3/2/2001	Dissolved Oxygen (mg/l)	1.95	5
Horse Creek at Goose Pond Road	HCSW-2	4/5/2011	Dissolved Oxygen (mg/l)	0.14	5
Horse Creek at Goose Pond Road	HCSW-2	5/3/2011	Dissolved Oxygen (mg/l)	1.78	5
Horse Creek at Goose Pond Road	HCSW-2	7/5/2011	Dissolved Oxygen (mg/l)	0.89	5
Horse Creek at Goose Pond Road	HCSW-2	8/16/2011	Dissolved Oxygen (mg/l)	0.59	5
Horse Creek at Goose Pond Road	HCSW-2	9/7/2011	Dissolved Oxygen (mg/l)	0.45	5
Horse Creek at Goose Pond Road	HCSW-2	10/24/2011	Dissolved Oxygen (mg/l)	1.11	5
Horse Creek at Goose Pond Road	HCSW-2	11/29/2011	Dissolved Oxygen (mg/l)	2.7	5
Horse Creek at Goose Pond Road	HCSW-2	3/5/2012	Dissolved Oxygen (mg/l)	4.55	5
Horse Creek at Goose Pond Road	HCSW-2	5/2/2012	Dissolved Oxygen (mg/l)	3.32	5
Horse Creek at Goose Pond Road	HCSW-2	10/10/2012	Dissolved Oxygen (mg/l)	2.92	5
Horse Creek at Goose Pond Road	HCSW-2	11/6/2012	Dissolved Oxygen (mg/l)	3.95	5
Horse Creek at Goose Pond Road	HCSW-2	12/5/2012	Dissolved Oxygen (mg/l)	4.74	5
Horse Creek at State Road 70	HCSW-3	7/27/2004	Dissolved Oxygen (mg/l)	4.7	5
Horse Creek at State Road 70	HCSW-3	8/30/2004	Dissolved Oxygen (mg/l)	0.27	5
Horse Creek at State Road 70	HCSW-3	9/29/2004	Dissolved Oxygen (mg/l)	2.4	5
Horse Creek at State Road 70	HCSW-3	6/22/2005	Dissolved Oxygen (mg/l)	3.9	5
Horse Creek at State Road 70	HCSW-3	7/27/2005	Dissolved Oxygen (mg/l)	3.5	5
Horse Creek at State Road 70	HCSW-3	8/23/2005	Dissolved Oxygen (mg/l)	4.4	5
Horse Creek at State Road 70	HCSW-3	7/27/2006	Dissolved Oxygen (mg/l)	4.5	5
Horse Creek at State Road 70	HCSW-3	8/21/2006	Dissolved Oxygen (mg/l)	3.7	5
Horse Creek at State Road 70	HCSW-3	9/27/2006	Dissolved Oxygen (mg/l)	1.8	5
Horse Creek at State Road 70	HCSW-3	10/19/2006	Dissolved Oxygen (mg/l)	4.5	5
Horse Creek at State Road 70	HCSW-3	7/18/2007	Dissolved Oxygen (mg/l)	3.93	5
Horse Creek at State Road 70	HCSW-3	8/27/2007	Dissolved Oxygen (mg/l)	2.8	5
Horse Creek at State Road 70	HCSW-3	9/26/2007	Dissolved Oxygen (mg/l)	2.88	5
Horse Creek at State Road 70	HCSW-3	10/29/2007	Dissolved Oxygen (mg/l)	3.06	5
Horse Creek at State Road 70	HCSW-3	11/29/2007	Dissolved Oxygen (mg/l)	4.3	5
Horse Creek at State Road 70	HCSW-3	2/26/2008	Dissolved Oxygen (mg/l)	3.64	5
Horse Creek at State Road 70	HCSW-3	3/27/2008	Dissolved Oxygen (mg/l)	4.75	5
Horse Creek at State Road 70	HCSW-3	4/23/2008	Dissolved Oxygen (mg/l)	3.27	5
Horse Creek at State Road 70	HCSW-3	5/29/2008	Dissolved Oxygen (mg/l)	2.9	5
Horse Creek at State Road 70	HCSW-3	6/26/2008	Dissolved Oxygen (mg/l)	4.78	5
Horse Creek at State Road 70	HCSW-3	7/31/2008	Dissolved Oxygen (mg/l)	0.99	5
Horse Creek at State Road 70	HCSW-3	8/26/2008	Dissolved Oxygen (mg/l)	1.62	5
Horse Creek at State Road 70	HCSW-3	9/30/2008	Dissolved Oxygen (mg/l)	3.28	5
Horse Creek at State Road 70	HCSW-3	10/16/2008	Dissolved Oxygen (mg/l)	2.73	5
Horse Creek at State Road 70	HCSW-3	6/3/2009	Dissolved Oxygen (mg/l)	3.89	5
Horse Creek at State Road 70	HCSW-3	7/8/2009	Dissolved Oxygen (mg/l)	3.38	5
Horse Creek at State Road 70	HCSW-3	8/5/2009	Dissolved Oxygen (mg/l)	3.33	5
Horse Creek at State Road 70	HCSW-3	9/2/2009	Dissolved Oxygen (mg/l)	3.87	5
Horse Creek at State Road 70	HCSW-3	10/7/2009	Dissolved Oxygen (mg/l)	3.13	5
Horse Creek at State Road 70	HCSW-3	4/6/2010	Dissolved Oxygen (mg/l)	4.74	5
Horse Creek at State Road 70	HCSW-3	7/12/2010	Dissolved Oxygen (mg/l)	3.67	5
Horse Creek at State Road 70	HCSW-3	8/3/2010	Dissolved Oxygen (mg/l)	4.61	5
Horse Creek at State Road 70	HCSW-3	9/8/2010	Dissolved Oxygen (mg/l)	4.09	5
Horse Creek at State Road 70	HCSW-3	8/16/2011	Dissolved Oxygen (mg/l)	4.14	5

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at State Road 70	HCSW-3	9/7/2011	Dissolved Oxygen (mg/l)	3.32	5
Horse Creek at State Road 70	HCSW-3	6/5/2012	Dissolved Oxygen (mg/l)	4.64	5
Horse Creek at State Road 70	HCSW-3	7/5/2012	Dissolved Oxygen (mg/l)	3.28	5
Horse Creek at State Road 70	HCSW-3	8/2/2012	Dissolved Oxygen (mg/l)	3.05	5
Horse Creek at State Road 70	HCSW-3	9/5/2012	Dissolved Oxygen (mg/l)	3.8	5
Horse Creek at State Road 70	HCSW-3	10/10/2012	Dissolved Oxygen (mg/l)	4.66	5
Horse Creek at State Road 72	HCSW-4	8/30/2004	Dissolved Oxygen (mg/l)	0.58	5
Horse Creek at State Road 72	HCSW-4	9/29/2004	Dissolved Oxygen (mg/l)	2.9	5
Horse Creek at State Road 72	HCSW-4	6/22/2005	Dissolved Oxygen (mg/l)	4	5
Horse Creek at State Road 72	HCSW-4	7/27/2005	Dissolved Oxygen (mg/l)	4.1	5
Horse Creek at State Road 72	HCSW-4	9/24/2006	Dissolved Oxygen (mg/l)	4.1	5
Horse Creek at State Road 72	HCSW-4	7/31/2008	Dissolved Oxygen (mg/l)	3.1	5
Horse Creek at State Road 72	HCSW-4	8/26/2008	Dissolved Oxygen (mg/l)	2.2	5
Horse Creek at State Road 72	HCSW-4	9/30/2008	Dissolved Oxygen (mg/l)	4.77	5
Horse Creek at State Road 72	HCSW-4	7/8/2009	Dissolved Oxygen (mg/l)	4.2	5
Horse Creek at State Road 72	HCSW-4	8/5/2009	Dissolved Oxygen (mg/l)	3.36	5
Horse Creek at State Road 72	HCSW-4	9/2/2009	Dissolved Oxygen (mg/l)	4.89	5
Horse Creek at State Road 72	HCSW-4	10/7/2009	Dissolved Oxygen (mg/l)	4.48	5
Horse Creek at State Road 72	HCSW-4	7/12/2010	Dissolved Oxygen (mg/l)	4.31	5
Horse Creek at State Road 72	HCSW-4	4/5/2011	Dissolved Oxygen (mg/l)	4.89	5
Horse Creek at State Road 72	HCSW-4	9/7/2011	Dissolved Oxygen (mg/l)	4.29	5
Horse Creek at State Road 72	HCSW-4	7/5/2012	Dissolved Oxygen (mg/l)	2.23	5
Horse Creek at State Road 72	HCSW-4	9/5/2012	Dissolved Oxygen (mg/l)	4.12	5
Horse Creek at State Road 64	HCSW-1	4/27/2006	Color (PCU)	20	25
Horse Creek at State Road 70	HCSW-3	4/27/2006	Color (PCU)	15	25
Horse Creek at State Road 70	HCSW-3	6/29/2006	Color (PCU)	15	25
Horse Creek at Goose Pond Road	HCSW-2	1/30/2008	Total Nitrogen (mg/L)	4.8	3
Horse Creek at State Road 70	HCSW-3	9/27/2006	Total Nitrogen (mg/L)	6.7	3
Horse Creek at State Road 70	HCSW-3	6/20/2007	Total Nitrogen (mg/L)	9.68	3
Horse Creek at Goose Pond Road	HCSW-2	7/31/2008	Total Ammonia (mg/L)	0.41	0.3
Horse Creek at State Road 70	HCSW-3	7/31/2008	Total Ammonia (mg/L)	0.32	0.3
Horse Creek at State Road 70	HCSW-3	5/3/2011	Total Ammonia (mg/L)	0.31	0.3
Horse Creek at State Road 72	HCSW-4	7/31/2008	Total Ammonia (mg/L)	0.31	0.3
Horse Creek at State Road 64	HCSW-1	2/2/2010	Chlorophyll a (mg/m^3)	15.4	15
Horse Creek at Goose Pond Road	HCSW-2	4/14/2004	Chlorophyll a (mg/m^3)	16	15
Horse Creek at Goose Pond Road	HCSW-2	5/26/2004	Chlorophyll a (mg/m^3)	21	15
Horse Creek at Goose Pond Road	HCSW-2	8/30/2004	Chlorophyll a (mg/m^3)	35	15
Horse Creek at Goose Pond Road	HCSW-2	5/27/2005	Chlorophyll a (mg/m^3)	17	15
Horse Creek at Goose Pond Road	HCSW-2	11/17/2005	Chlorophyll a (mg/m^3)	17	15
Horse Creek at Goose Pond Road	HCSW-2	2/23/2006	Chlorophyll *a* (mg/m^3)	23	15
Horse Creek at Goose Pond Road	HCSW-2	3/28/2006	Chlorophyll *a* (mg/m^3)	30	15
Horse Creek at Goose Pond Road	HCSW-2	5/25/2006	Chlorophyll *a* (mg/m^3)	32	15
Horse Creek at Goose Pond Road	HCSW-2	6/29/2006	Chlorophyll *a* (mg/m^3)	45	15
Horse Creek at Goose Pond Road	HCSW-2	8/21/2006	Chlorophyll *a* (mg/m^3)	20	15

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at Goose Pond Road	HCSW-2	5/16/2007	Chlorophyll *a* (mg/m^3)	25	15
Horse Creek at Goose Pond Road	HCSW-2	6/20/2007	Chlorophyll *a* (mg/m^3)	110	15
Horse Creek at Goose Pond Road	HCSW-2	7/18/2007	Chlorophyll *a* (mg/m^3)	17	15
Horse Creek at Goose Pond Road	HCSW-2	7/31/2008	Chlorophyll *a* (mg/m^3)	22.6	15
Horse Creek at Goose Pond Road	HCSW-2	1/5/2009	Chlorophyll *a* (mg/m^3)	24.9	15
Horse Creek at Goose Pond Road	HCSW-2	4/1/2009	Chlorophyll *a* (mg/m^3)	21.7	15
Horse Creek at Goose Pond Road	HCSW-2	5/3/2011	Chlorophyll *a* (mg/m^3)	17.5	15
Horse Creek at Goose Pond Road	HCSW-2	2/2/2012	Chlorophyll *a* (mg/m^3)	75.1	15
Horse Creek at Goose Pond Road	HCSW-2	4/2/2012	Chlorophyll *a* (mg/m^3)	35.9	15
Horse Creek at Goose Pond Road	HCSW-2	5/2/2012	Chlorophyll *a* (mg/m^3)	34.1	15
Horse Creek at Goose Pond Road	HCSW-2	12/5/2012	Chlorophyll *a* (mg/m^3)	17.9	15
Horse Creek at State Road 70	HCSW-3	8/30/2004	Chlorophyll a (mg/m^3)	38	15
Horse Creek at State Road 70	HCSW-3	4/27/2006	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 70	HCSW-3	6/29/2006	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 70	HCSW-3	4/25/2007	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 70	HCSW-3	5/16/2007	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 70	HCSW-3	6/20/2007	Dissolved Calcium (mg/L)	140	100
Horse Creek at State Road 70	HCSW-3	4/2/2012	Dissolved Calcium (mg/L)	114	100
Horse Creek at State Road 72	HCSW-4	5/25/2006	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 72	HCSW-4	6/29/2006	Dissolved Calcium (mg/L)	190	100
Horse Creek at State Road 72	HCSW-4	12/13/2006	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 72	HCSW-4	6/20/2007	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 72	HCSW-4	3/27/2008	Dissolved Calcium (mg/L)	120	100
Horse Creek at State Road 72	HCSW-4	5/29/2008	Dissolved Calcium (mg/L)	110	100
Horse Creek at State Road 72	HCSW-4	2/2/2009	Dissolved Calcium (mg/L)	106	100
Horse Creek at State Road 72	HCSW-4	6/3/2009	Dissolved Calcium (mg/L)	119	100
Horse Creek at State Road 72	HCSW-4	4/2/2012	Dissolved Calcium (mg/L)	117	100
Horse Creek at State Road 72	HCSW-4	5/2/2012	Dissolved Calcium (mg/L)	109	100
Horse Creek at State Road 72	HCSW-4	6/5/2012	Dissolved Calcium (mg/L)	182	100
Horse Creek at Goose Pond Road	HCSW-2	7/27/2006	Dissolved Iron (mg/L)	1.2	1
Horse Creek at Goose Pond Road	HCSW-2	6/3/2009	Dissolved Iron (mg/L)	1.03	1
Horse Creek at State Road 72	HCSW-4	5/27/2003	Dissolved Iron (mg/L)	0.4	0.3
Horse Creek at State Road 72	HCSW-4	6/19/2003	Dissolved Iron (mg/L)	0.6	0.3
Horse Creek at State Road 72	HCSW-4	7/14/2003	Dissolved Iron (mg/L)	0.7	0.3
Horse Creek at State Road 72	HCSW-4	8/28/2003	Dissolved Iron (mg/L)	0.7	0.3
Horse Creek at State Road 72	HCSW-4	9/25/2003	Dissolved Iron (mg/L)	0.6	0.3
Horse Creek at State Road 72	HCSW-4	10/29/2003	Dissolved Iron (mg/L)	0.4	0.3
Horse Creek at State Road 72	HCSW-4	12/16/2003	Dissolved Iron (mg/L)	0.32	0.3
Horse Creek at State Road 72	HCSW-4	7/27/2004	Dissolved Iron (mg/L)	0.5	0.3
Horse Creek at State Road 72	HCSW-4	8/30/2004	Dissolved Iron (mg/L)	0.7	0.3
Horse Creek at State Road 72	HCSW-4	9/29/2004	Dissolved Iron (mg/L)	0.4	0.3
Horse Creek at State Road 72	HCSW-4	10/27/2004	Dissolved Iron (mg/L)	0.4	0.3
Horse Creek at State Road 72	HCSW-4	3/30/2005	Dissolved Iron (mg/L)	0.45	0.3
Horse Creek at State Road 72	HCSW-4	4/27/2005	Dissolved Iron (mg/L)	0.4	0.3
Horse Creek at State Road 72	HCSW-4	5/25/2005	Dissolved Iron (mg/L)	0.45	0.3
Horse Creek at State Road 72	HCSW-4	6/22/2005	Dissolved Iron (mg/L)	0.77	0.3

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at State Road 72	HCSW-4	7/27/2005	Dissolved Iron (mg/L)	0.74	0.3
Horse Creek at State Road 72	HCSW-4	8/23/2005	Dissolved Iron (mg/L)	0.58	0.3
Horse Creek at State Road 72	HCSW-4	9/29/2005	Dissolved Iron (mg/L)	0.44	0.3
Horse Creek at State Road 72	HCSW-4	11/17/2005	Dissolved Iron (mg/L)	0.49	0.3
Horse Creek at State Road 72	HCSW-4	7/27/2006	Dissolved Iron (mg/L)	0.5	0.3
Horse Creek at State Road 72	HCSW-4	8/21/2006	Dissolved Iron (mg/L)	0.7	0.3
Horse Creek at State Road 72	HCSW-4	9/27/2006	Dissolved Iron (mg/L)	1	0.3
Horse Creek at State Road 72	HCSW-4	10/19/2006	Dissolved Iron (mg/L)	0.6	0.3
Horse Creek at State Road 72	HCSW-4	7/18/2007	Dissolved Iron (mg/L)	0.42	0.3
Horse Creek at State Road 72	HCSW-4	7/31/2008	Dissolved Iron (mg/L)	0.81	0.3
Horse Creek at State Road 72	HCSW-4	8/26/2008	Dissolved Iron (mg/L)	0.96	0.3
Horse Creek at State Road 72	HCSW-4	9/30/2008	Dissolved Iron (mg/L)	0.59	0.3
Horse Creek at State Road 72	HCSW-4	10/16/2008	Dissolved Iron (mg/L)	0.64	0.3
Horse Creek at State Road 72	HCSW-4	7/8/2009	Dissolved Iron (mg/L)	0.483	0.3
Horse Creek at State Road 72	HCSW-4	8/5/2009	Dissolved Iron (mg/L)	0.567	0.3
Horse Creek at State Road 72	HCSW-4	9/2/2009	Dissolved Iron (mg/L)	0.603	0.3
Horse Creek at State Road 72	HCSW-4	10/7/2009	Dissolved Iron (mg/L)	0.527	0.3
Horse Creek at State Road 72	HCSW-4	4/6/2010	Dissolved Iron (mg/L)	0.615	0.3
Horse Creek at State Road 72	HCSW-4	7/12/2010	Dissolved Iron (mg/L)	0.719	0.3
Horse Creek at State Road 72	HCSW-4	8/3/2010	Dissolved Iron (mg/L)	0.321	0.3
Horse Creek at State Road 72	HCSW-4	9/8/2010	Dissolved Iron (mg/L)	0.421	0.3
Horse Creek at State Road 72	HCSW-4	8/16/2011	Dissolved Iron (mg/L)	0.325	0.3
Horse Creek at State Road 72	HCSW-4	9/7/2011	Dissolved Iron (mg/L)	0.506	0.3
Horse Creek at State Road 72	HCSW-4	10/24/2011	Dissolved Iron (mg/L)	0.36	0.3
Horse Creek at State Road 72	HCSW-4	7/5/2012	Dissolved Iron (mg/L)	0.779	0.3
Horse Creek at State Road 72	HCSW-4	8/2/2012	Dissolved Iron (mg/L)	0.531	0.3
Horse Creek at State Road 72	HCSW-4	9/5/2012	Dissolved Iron (mg/L)	0.604	0.3
Horse Creek at State Road 72	HCSW-4	10/10/2012	Dissolved Iron (mg/L)	0.508	0.3
Horse Creek at State Road 64	HCSW-1	4/25/2007	Alkalinity (mg/L)	120	100
Horse Creek at State Road 64	HCSW-1	5/16/2007	Alkalinity (mg/L)	170	100
Horse Creek at State Road 64	HCSW-1	6/20/2007	Alkalinity (mg/L)	140	100
Horse Creek at State Road 64	HCSW-1	1/5/2010	Alkalinity (mg/L)	109	100
Horse Creek at State Road 64	HCSW-1	10/24/2011	Alkalinity (mg/L)	102	100
Horse Creek at State Road 64	HCSW-1	11/6/2012	Alkalinity (mg/L)	100.3	100
Horse Creek at State Road 72	HCSW-4	5/25/2006	Alkalinity (mg/L)	120	100
Horse Creek at State Road 72	HCSW-4	5/4/2009	Alkalinity (mg/L)	112	100
Horse Creek at State Road 72	HCSW-4	6/8/2011	Alkalinity (mg/L)	1223	100
Horse Creek at State Road 72	HCSW-4	5/2/2012	Alkalinity (mg/L)	147.5	100
Horse Creek at State Road 72	HCSW-4	1/23/2007	Fluoride (mg/L)	2.5	1.5
Horse Creek at State Road 72	HCSW-4	2/14/2007	Fluoride (mg/L)	2.5	1.5
Horse Creek at State Road 72	HCSW-4	3/14/2007	Fluoride (mg/L))	5	1.5
Horse Creek at State Road 70	HCSW-3	3/28/2006	Sulfate (mg/L)	300	250
Horse Creek at State Road 70	HCSW-3	4/27/2006	Sulfate (mg/L)	420	250
Horse Creek at State Road 70	HCSW-3	6/29/2006	Sulfate (mg/L)	430	250
Horse Creek at State Road 70	HCSW-3	5/16/2007	Sulfate (mg/L)	360	250
Horse Creek at State Road 70	HCSW-3	6/20/2007	Sulfate (mg/L)	440	250

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at State Road 70	HCSW-3	6/26/2008	Sulfate (mg/L)	251	250
Horse Creek at State Road 70	HCSW-3	2/2/2009	Sulfate (mg/L)	280	250
Horse Creek at State Road 70	HCSW-3	4/1/2009	Sulfate (mg/L)	293	250
Horse Creek at State Road 70	HCSW-3	6/3/2009	Sulfate (mg/L)	251	250
Horse Creek at State Road 70	HCSW-3	2/2/2012	Sulfate (mg/L)	254	250
Horse Creek at State Road 70	HCSW-3	3/5/2012	Sulfate (mg/L)	287	250
Horse Creek at State Road 70	HCSW-3	4/2/2012	Sulfate (mg/L)	365	250
Horse Creek at State Road 70	HCSW-3	6/5/2012	Sulfate (mg/L)	304	250
Horse Creek at State Road 72	HCSW-4	6/29/2004	Sulfate (mg/l)	261	250
Horse Creek at State Road 72	HCSW-4	3/28/2006	Sulfate (mg/L)	300	250
Horse Creek at State Road 72	HCSW-4	5/25/2006	Sulfate (mg/L)	310	250
Horse Creek at State Road 72	HCSW-4	6/29/2006	Sulfate (mg/L)	780	250
Horse Creek at State Road 72	HCSW-4	12/13/2006	Sulfate (mg/L)	290	250
Horse Creek at State Road 72	HCSW-4	6/20/2007	Sulfate (mg/L)	320	250
Horse Creek at State Road 72	HCSW-4	3/27/2008	Sulfate (mg/L)	390	250
Horse Creek at State Road 72	HCSW-4	5/29/2008	Sulfate (mg/L)	290	250
Horse Creek at State Road 72	HCSW-4	6/26/2008	Sulfate (mg/L)	287	250
Horse Creek at State Road 72	HCSW-4	2/2/2009	Sulfate (mg/L)	290	250
Horse Creek at State Road 72	HCSW-4	6/3/2009	Sulfate (mg/L)	391	250
Horse Creek at State Road 72	HCSW-4	12/2/2009	Sulfate (mg/L)	279	250
Horse Creek at State Road 72	HCSW-4	11/3/2010	Sulfate (mg/L)	258	250
Horse Creek at State Road 72	HCSW-4	1/4/2011	Sulfate (mg/L)	262	250
Horse Creek at State Road 72	HCSW-4	7/5/2011	Sulfate (mg/L)	253	250
Horse Creek at State Road 72	HCSW-4	3/5/2012	Sulfate (mg/L)	267	250
Horse Creek at State Road 72	HCSW-4	4/2/2012	Sulfate (mg/L)	321	250
Horse Creek at State Road 72	HCSW-4	6/5/2012	Sulfate (mg/L)	665	250
Horse Creek at State Road 70	HCSW-3	4/27/2006	TDS (mg/L)	580	500
Horse Creek at State Road 70	HCSW-3	6/29/2006	TDS (mg/L)	590	500
Horse Creek at State Road 70	HCSW-3	4/25/2007	TDS (mg/L)	590	500
Horse Creek at State Road 70	HCSW-3	5/16/2007	TDS (mg/L)	530	500
Horse Creek at State Road 70	HCSW-3	6/20/2007	TDS (mg/L)	700	500
Horse Creek at State Road 70	HCSW-3	7/18/2007	TDS (mg/L)	520	500
Horse Creek at State Road 70	HCSW-3	6/26/2008	TDS (mg/L)	580	500
Horse Creek at State Road 70	HCSW-3	2/2/2009	TDS (mg/L)	520	500
Horse Creek at State Road 70	HCSW-3	4/1/2009	TDS (mg/L)	568	500
Horse Creek at State Road 70	HCSW-3	6/3/2009	TDS (mg/L)	540	500
Horse Creek at State Road 70	HCSW-3	12/2/2009	TDS (mg/L)	524	500
Horse Creek at State Road 70	HCSW-3	1/4/2011	TDS (mg/L)	513	500
Horse Creek at State Road 70	HCSW-3	12/21/2011	TDS (mg/L)	543	500
Horse Creek at State Road 70	HCSW-3	1/12/2012	TDS (mg/L)	571	500
Horse Creek at State Road 70	HCSW-3	2/2/2012	TDS (mg/L)	532	500
Horse Creek at State Road 70	HCSW-3	3/5/2012	TDS (mg/L)	603	500
Horse Creek at State Road 70	HCSW-3	4/2/2012	TDS (mg/L)	714	500
Horse Creek at State Road 70	HCSW-3	6/5/2012	TDS (mg/L)	646	500
Horse Creek at State Road 72	HCSW-4	3/28/2006	TDS (mg/L)	600	500
Horse Creek at State Road 72	HCSW-4	5/25/2006	TDS (mg/L)	560	500
Horse Creek at State Road 72	HCSW-4	6/29/2006	TDS (mg/L)	1100	500
Horse Creek at State Road 72	HCSW-4	11/9/2006	TDS (mg/L)	510	500

Sampling Location	Station ID	Date	Analyte	Concentration	Trigger Level
Horse Creek at State Road 72	HCSW-4	12/13/2006	TDS (mg/L)	550	500
Horse Creek at State Road 72	HCSW-4	6/20/2007	TDS (mg/L)	600	500
Horse Creek at State Road 72	HCSW-4	7/18/2007	TDS (mg/L)	530	500
Horse Creek at State Road 72	HCSW-4	1/30/2008	TDS (mg/L)	550	500
Horse Creek at State Road 72	HCSW-4	3/27/2008	TDS (mg/L)	660	500
Horse Creek at State Road 72	HCSW-4	5/29/2008	TDS (mg/L)	710	500
Horse Creek at State Road 72	HCSW-4	6/26/2008	TDS (mg/L)	644	500
Horse Creek at State Road 72	HCSW-4	2/2/2009	TDS (mg/L)	536	500
Horse Creek at State Road 72	HCSW-4	6/3/2009	TDS (mg/L)	692	500
Horse Creek at State Road 72	HCSW-4	12/2/2009	TDS (mg/L)	604	500
Horse Creek at State Road 72	HCSW-4	11/3/2010	TDS (mg/L)	577	500
Horse Creek at State Road 72	HCSW-4	1/4/2011	TDS (mg/L)	574	500
Horse Creek at State Road 72	HCSW-4	7/5/2011	TDS (mg/L)	660	500
Horse Creek at State Road 72	HCSW-4	12/21/2011	TDS (mg/L)	543	500
Horse Creek at State Road 72	HCSW-4	1/12/2012	TDS (mg/L)	569	500
Horse Creek at State Road 72	HCSW-4	2/2/2012	TDS (mg/L)	512	500
Horse Creek at State Road 72	HCSW-4	3/5/2012	TDS (mg/L)	585	500
Horse Creek at State Road 72	HCSW-4	4/2/2012	TDS (mg/L)	688	500
Horse Creek at State Road 72	HCSW-4	5/2/2012	TDS (mg/L)	536	500
Horse Creek at State Road 72	HCSW-4	6/5/2012	TDS (mg/L)	1,320	500
Horse Creek at State Road 64	HCSW-1	6/20/2007	Total Fatty Acids (mg/L)	1.5	0.5
Horse Creek at Goose Pond Road	HCSW-2	11/18/2004	Total Fatty Acids (mg/L)	1.1	0.5
Horse Creek at Goose Pond Road	HCSW-2	3/30/2005	Total Fatty Acids (mg/L)	0.56	0.5
Horse Creek at Goose Pond Road	HCSW-2	4/27/2005	Total Fatty Acids (mg/L)	0.53	0.5
Horse Creek at Goose Pond Road	HCSW-2	12/13/2006	Total Fatty Acids (mg/L)	1.6	0.5
Horse Creek at Goose Pond Road	HCSW-2	6/20/2007	Total Fatty Acids (mg/L)	1.6	0.5
Horse Creek at Goose Pond Road	HCSW-2	7/18/2007	Total Fatty Acids (mg/L)	0.87	0.5
Horse Creek at Goose Pond Road	HCSW-2	12/17/2007	Total Fatty Acids (mg/L)	0.88	0.5
Horse Creek at Goose Pond Road	HCSW-2	12/4/2008	Total Fatty Acids (mg/L)	0.97	0.5
Horse Creek at Goose Pond Road	HCSW-2	4/1/2009	Total Fatty Acids (mg/L)	0.68	0.5
Horse Creek at State Road 70	HCSW-3	2/24/2005	Total Fatty Acids (mg/L)	1.3	0.5
Horse Creek at State Road 72	HCSW-4	6/29/2006	Total Fatty Acids (mg/L)	1.4	0.5
Horse Creek at State Road 72	HCSW-4	12/17/2007	Total Fatty Acids (mg/L)	0.55	0.5
Horse Creek at Goose Pond Road	HCSW-2	7/27/2004	Radium (pCi/l)	5.1	5

APPENDIX G

SUMMARY OF IMPACT ASSESSMENTS FROM 2003 - 2012

Appendix G
Summary of Impact Assessments from 2003-2012

Station	Date	Exceedance	Action Taken	Conclusions
HCSW-4	7/14/2003	Dissolved Iron	A special sampling program was carried out in August 2003 where samples were collected from three locations on Horse Creek and two tributaries, but flow conditions were very high. In October 2003, eleven stations were sampled while flow was closer to normal.	Readings appear normal for the basin, the lower trigger level at this location caused the exceedance due to differences in water class. The trigger value may be set too low at this location.
HCSW-2	8/28/2003	Dissolved Oxygen	A sampling program was attempted in August 2003 in the northern portion of the stream, but flow conditions were very high. Instead six locations including tributaries were sampled at the end of October 2003.	Low DO levels persisted at HCSW-2 due to generally low streamflow levels and a greater amount of organics than the other stations. The low levels are not due to mining upstream.
HCSW-2	4/14/2004	Chlorophyll a	A special sampling program was carried out in May 2004 where samples were taken from four upstream locations in Horse Creek (due to dry conditions of tributaries).	Elevated chlorophyll a concentrations were caused by low streamflow and the physical nature of the stream channel and not mining activities.
HCSW-4	6/29/2004	Sulfate	A special sampling program was carried out where samples were taken from nearby tributaries as well as the HCSP stations during July 2004.	Nearby tributary basins have high amounts of agricultural activity (requiring irrigation) and streamflow was very low at this time which led to the elevated sulfate concentration in June 2004.
HCSW-2	7/27/2004	Total Radium	None	Blank sample results had high values, making other values suspect. No impact assessment required for July 2004, but future results should be monitored.
HCSW-1	8/30/2004	Dissolved Oxygen	None	Impact assessment deferred until the streamflows in Horse Creek are near normal for the time period the exceedance occurred (multiple hurricanes passing through region dramatically increased streamflow).

Station	Date	Exceedance	Action Taken	Conclusions
HCSW-2	8/30/2004	Dissolved Oxygen	None	Impact assessment deferred until the streamflows in Horse Creek are near normal for the time period the exceedance occurred (multiple hurricanes passing through region dramatically increased streamflow).
HCSW-3	8/30/2004	Dissolved Oxygen	None	Impact assessment deferred until the streamflows in Horse Creek are near normal for the time period the exceedance occurred (multiple hurricanes passing through region dramatically increased streamflow).
HCSW-4	8/30/2004	Dissolved Oxygen	None	Impact assessment deferred until the streamflows in Horse Creek are near normal for the time period the exceedance occurred (multiple hurricanes passing through region dramatically increased streamflow).
HCSW-2	8/30/2004	Chlorophyll a	None	Impact assessment deferred until the streamflows in Horse Creek are near normal for the time period the exceedance occurred (multiple hurricanes passing through region dramatically increased streamflow).
HCSW-3	8/30/2004	Chlorophyll a	None	Impact assessment deferred until the streamflows in Horse Creek are near normal for the time period the exceedance occurred (multiple hurricanes passing through region dramatically increased streamflow).
HCSW-2	11/18/2004	Total Fatty Acids	A special sampling program was carried out in January 2005, where three Horse Creek locations and a tributary (Brushy Creek) were sampled.	Nearby Horse Creek Prairie is likely to contribute to the elevated levels since all other stations had undetected values for fatty acids. Low streamflow and high organics in this region, not mining, were likely contributing factors.
HCSW-2	4/27/2005	Total Fatty Acids	A special sampling program was carried out in June 2005, where three Horse Creek locations and a tributary were sampled.	The exceedance is most likely caused by the surrounding habitat conditions and not impacted by mining.
HCSW-2	7/27/2006	Iron	None	Nearby Horse Creek Prairie is likely to contribute to the elevated levels since all other stations had lower iron concentrations.
HCSW-1	1/23/2007	pH	Compared measurement to SWFWMD measurements for the months of January and February.	Not an actual exceedance but equipment malfunction
HCSW-4	1/23/2007	pH	Compared measurement to SWFWMD	Not an actual exceedance but equipment malfunction

Station	Date	Exceedance	Action Taken	Conclusions
			measurements for the months of January and February.	
HCSW-1	4/25/2007	Alkalinity	Statistical analysis of HCSP alkalinity and SWFWMD measurements. When alkalinity compared to streamflow, there was a weak negative correlation between the two (high alkalinity during low flow).	No evidence that high alkalinity was caused by mining, but was rather a seasonal pattern caused by lower water levels and flow. Once those recovered during the wet season, the alkalinity values decreased.
HCSW-1	6/20/2007	Total Fatty Acids	Used conclusions from the FIPR on the rate of biodegradation and soil leaching of organic compounds in a controlled environment.	It was unlikely that fatty acids from mining process water are responsible for the elevated levels seen. Instead it represents the variation in naturally-occurring fatty acids in Horse Creek.
HCSW-2	6/20/2007	Total Fatty Acids	Used conclusions from the FIPR on the rate of biodegradation and soil leaching of organic compounds in a controlled environment.	It was unlikely that fatty acids from mining process water are responsible for the elevated levels seen. Instead it represents the variation in naturally-occurring fatty acids in Horse Creek.
HCSW-2-FD	6/20/2007	Total Nitrogen	Compared to nitrate+nitrate and TKN values from April 2003 through August at HCSP.	Elevated measurements most likely due to lab analyst or instrument error. The total nitrogen levels recorded are not corroborated by measurements taken before of after the exceedance.
HCSW-3	6/20/2007	Total Nitrogen	Compared to nitrate+nitrate and TKN values from April 2003 through August at HCSP.	Elevated measurements most likely due to lab analyst or instrument error. The total nitrogen levels recorded are not corroborated by measurements taken before of after the exceedance.
HCSW-2	7/31/2008	Ammonia	None	Elevated concentrations are either due to laboratory method change or a seasonal fluctuation in the nitrogen cycle.
HCSW-3	7/31/2008	Ammonia	None	Elevated concentrations are either due to laboratory method change or a seasonal fluctuation in the nitrogen cycle.
HCSW-4	7/31/2008	Ammonia	None	Elevated concentrations are either due to laboratory method change or a seasonal fluctuation in the nitrogen cycle.
HCSW-4	5/4/2009	Alkalinity	Statistical analysis of HCSP alkalinity and SWFWMD measurements. When alkalinity compared to rainfall, there was a strong negative correlation between the two (high alkalinity during	No evidence that high alkalinity was caused by mining, but was rather a seasonal pattern caused by lower water levels, flow, and rainfall. Once those recovered during the wet season, the alkalinity values decreased.

Station	Date	Exceedance	Action Taken	Conclusions
			low flow).	
HCSW-1	2/2/2010	Chlorophyll a	None	No connection with mining. May have been a sampling error since color, pH, and DO do not indicate a significant algal bloom causing an elevated chlorophyll a reading.
HCSW-3	5/3/2011	Ammonia	None	No connection with mining. Lab analyzing the data was getting back results that were ten-times higher than other labs. Split sampling conducted in October 2011, verified that it was lab error.
HCSW-1	11/6/2012	Alkalinity	None	Alkalinity increased during times of low flow (groundwater influence) and after summer rains which included discharge. Alkalinity increased from the beginning of the wet season through November, but then decreased again after rains and discharge stopped. High alkalinity in region.

APPENDIX H

SUMMARY OF TRENDS FROM THE HCSP 2008-2011 ANNUAL REPORTS

Appendix H
Summary of Trends from the HCSP 2009-2011 Annual Reports

Station	Year	Parameter	Trend Noted	Addressing of Trend
HCSW-1	2008	Alkalinity	increasing trend with slope of 4.58	Alkalinity was higher in the dry season and lower during times of NDPES discharge. The trend was not shown to be linked to mining discharge. The trend is more likely linked to climate factors not completely accounted for in the LOESS smoothing with streamflow.
HCSW-1	2008	Specific Conductance	increasing trend with slope of 15.31	Conductivity was higher in the dry season and lower (or equal) during times of NDPES discharge. The trend was not shown to be linked to mining discharge. The trend is more likely linked to climate factors not completely accounted for in the LOESS smoothing with streamflow.
HCSW-1	2009	Alkalinity	increasing trend with slope of 4.71	As with specific conductivity, it is likely that this trend is strongly influenced by the dry conditions in 2006 to 2007, given that most of the highest alkalinity measurements are associated with periods without NPDES discharge. The estimated slopes for both stations are small compared to the historic HCSP MDL (<= 1 mg/L) and/or the differences between primary and field duplicate samples (<= 17 mg/L).
HCSW-1	2009	Dissolved Calcium	increasing trend with slope of 1.56	As with specific conductivity, it is likely that this trend is strongly influenced by the dry conditions in 2006 to 2007, given that most of the highest calcium measurements are associated with periods without NPDES discharge. The estimated slope of the trend for HCSW-1 is small compared to historic HCSP differences between primary and field duplicate samples (<= 8.0 mg/L).
HCSW-1	2009	Chloride	slight increasing trend with slope of 0.50	As with the other dissolved ion parameters, the trend for HCSW-1 is small compared to the historic HCSP MDL (<= 4.06 mg/L) and differences between primary and field duplicate samples (<= 5.0 mg/L). The observed changes in chloride over time are probably related to the differences in rainfall over the course of the HCSP.
HCSW-1	2009	Orthophosphate	slight increasing trend with slope of 0.03	The slope is very small compared to historic HCSP values for laboratory Minimum Detection Limits (<= 0.075 mg/L) or differences between primary and field duplicate samples (<= 0.034 mg/L). Therefore, the trends at both stations are not of concern at this time, and could be related to extreme differences in rainfall and

Station	Year	Parameter	Trend Noted	Addressing of Trend
				streamflow within the sampling period.
HCSW-1	2009	Specific Conductance	increasing trend with slope of 16.73	It is likely that this trend is strongly influenced by the dry conditions and subsequent higher than average conductivity in 2006 to 2007, given that conductivity is greatly influenced by rainfall and most of the highest conductivity measurements are associated with dryer years. The estimated slope of the trend for HCSW-1 is not of concern at this time because of the substantial variability in rainfall over the course of the HCSP.
HCSW-1	2009	Total Dissolved Solids	increasing trend with slope of 9.46	As with the other dissolved ion parameters, the trend for HCSW-1 is small compared to differences between primary and field duplicate samples (<= 44 mg/L). The observed changes in TDS over time are probably related to the differences in rainfall over the course of the HCSP.
HCSW-1	2010	Alkalinity	increasing trend with slope of 4.19	2010 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2010	Dissolved Calcium	increasing trend with slope of 1.60	2010 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2010	Fluoride	slight increasing trend with slope of 0.01	2010 Impact Assessment concluded that estimated slope was too small to be ecologically significant.
HCSW-1	2010	Ammonia	slight decreasing trend with slope of -0.002	2010 Impact Assessment concluded that estimated slope was in the opposite direction of adverse change.
HCSW-1	2010	Orthophosphate	slight increasing trend with slope of 0.27	2010 Impact Assessment found that the evident trend was caused by a data bias, and extending the period of record before 2003 caused the trend to no longer be valid. Concentrations in 2008-2010 were similar to those before 2003.
HCSW-1	2010	pH	slight increasing trend with slope of 0.06	2010 Impact Assessment concluded that estimated slope was too small to be ecologically significant.
HCSW-1	2010	Specific Conductance	increasing trend with slope of 16.68	2010 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2010	Total Dissolved	increasing trend with	2010 Impact Assessment found step-change in

Station	Year	Parameter	Trend Noted	Addressing of Trend
		Solids	slope of 10.66	conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2011	Alkalinity	increasing trend with slope of 3.91	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2011	Dissolved Calcium	increasing trend with slope of 1.37	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2011	Dissolved Iron	slight decreasing trend with slope of -0.02	2011 Impact Assessment concluded that estimated slope was in the opposite direction of adverse change.
HCSW-1	2011	Fluoride	slight increasing trend with slope of 0.01	2011 Impact Assessment concluded that estimated slope was too small to be ecologically significant.
HCSW-1	2011	Sulfate	Increasing trend with slope of 2.82	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2011	Ammonia	slight decreasing trend with slope of -0.002	2011 Impact Assessment concluded that estimated slope was in the opposite direction of adverse change.
HCSW-1	2011	Orthophosphate	slight increasing trend with slope of 0.02	2011 Impact Assessment found that the evident trend was caused by a data bias, and extending the period of record before 2003 caused the trend to no longer be valid. Concentrations in 2008-2011 were similar to those before 2003.
HCSW-1	2011	pH	slight increasing trend with slope of 0.05	2011 Impact Assessment concluded that estimated slope was too small to be ecologically significant.
HCSW-1	2011	Specific Conductance	increasing trend with slope of 14.57	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-1	2011	Total Dissolved Solids	increasing trend with slope of 9.65	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.

Station	Year	Parameter	Trend Noted	Addressing of Trend
HCSW-4	2008	Dissolved Oxygen	slight decreasing trend with slope of -0.40	May be influenced by climate or other land use in southern basin.
HCSW-4	2008	Orthophosphate	slight increasing trend with slope of 0.02 (data not correlated with streamflow)	Magnitude of trend not ecologically significant. May be influenced by climate or other land use in southern basin.
HCSW-4	2009	Alkalinity	increasing trend with slope of 1.90	As with specific conductivity, it is likely that this trend is strongly influenced by the dry conditions in 2006 to 2007, given that most of the highest alkalinity measurements are associated with periods without NPDES discharge. The estimated slopes for both stations are small compared to the historic HCSP MDL (<= 1 mg/L) and/or the differences between primary and field duplicate samples (<= 17 mg/L).
HCSW-4	2009	Dissolved Oxygen	slight decreasing trend with slope of -0.42	It appears the declining trend stems from the difference between DO concentrations in 2006-2007 (dry years) compared to 2008-2009. When comparing DO overall annual and seasonal medians, DO concentrations in 2008-2009 are consistent with those in 2003-2005. Given this information and the fact that HCSW-1 does not show a significant trend, it is unlikely that mining activities are contributing to a perceived trend in dissolved oxygen concentrations at HCSW-4.
HCSW-4	2009	Orthophosphate	slight increasing trend with slope of 0.02 (data not correlated with streamflow)	The slope is very small compared to historic HCSP values for laboratory Minimum Detection Limits (<= 0.075 mg/L) or differences between primary and field duplicate samples (<= 0.034 mg/L). Therefore, the trends at both stations are not of concern at this time, and could be related to extreme differences in rainfall and streamflow within the sampling period.
HCSW-4	2010	Color	increasing trend with slope of 12.07	2010 Impact Assessment concluded that estimated slope was in the opposite direction of adverse change.
HCSW-4	2010	Orthophosphate	slight increasing trend with slope of 0.02	2010 Impact Assessment found that the evident trend was caused by a data bias, and extending the period of record before 2003 caused the trend to no longer be valid. Concentrations in 2008-2010 were similar to those before 2003.
HCSW-4	2010	Alkalinity	Increasing trend with slope of 1.62	2010 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-4	2011	Color	increasing trend with slope of 11.47	2011 Impact Assessment concluded that estimated slope was in the opposite direction of adverse change.

Station	Year	Parameter	Trend Noted	Addressing of Trend
HCSW-4	2011	Alkalinity	increasing trend with slope of 1.31	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.
HCSW-4	2011	Dissolved Iron	slight decreasing trend with slope of -0.01	2011 Impact Assessment found step-change in conductivity and dissolved ions likely caused by increased groundwater contributions from 2006-2008 drought and subsequent influence from Wingate dredge mining outfall discharge. Concentrations after 2008 remain relatively steady, with no evidence of biological impact.

APPENDIX

I

2012 WATER QUALITY TREND IMPACT ASSESSMENT

Appendix I
2012 Water Quality Trend Impact Assessment

Table of Contents

Introduction

This report was prepared as a component of the Horse Creek Stewardship Program (HCSP). As part of the HCSP, Mosaic monitors four locations on Horse Creek monthly for a number of water quality parameters and seasonally for biological indicators. At the end of each calendar year, an annual report is prepared that summarizes the collected information, including additional water quantity and quality data from public sources like the Southwest Florida Water Management District (SWFWMD) and United States Geological Service (USGS).

The HCSP plan document requires that an "impact assessment" be conducted for any trigger level exceedance or adverse water quality trends found while preparing the annual HCSP report. Impact assessments often include additional information that was not summarized in the annual report. If the impact assessment indicates or suggests that mining activities by Mosaic are the cause of the adverse exceedance or trend, then Mosaic will need to take corrective action. The intensity of the corrective action will be based on the potential for short-term or long-term ecological harm to Horse Creek and/or the integrity of the downstream potable water supply.

In the 2012 Annual Report, the Seasonal Kendall Tau procedure found several statistically significant trends in the water quality data from 2003 to 2012 (Table 1), thus triggering this impact assessment. Not all of the statistically significant trends had the potential to be ecologically significant because the direction of the trend was not harmful or the magnitude of the trend slope was within the error of the measurement device. Orthophosphate and specific conductivity were the two parameters of interest for the bulk of this impact assessment; other trends in dissolved ions were considered to be covered by the focus on specific conductivity.

In this impact assessment, we examine the statistically significant trends for indications that mining activities by Mosaic are the cause of the trend and what the potential impacts are for Horse Creek ecology and the quality of the downstream potable water supply. Our assessment consists of four parts: trend analysis of additional Horse Creek data, trend analysis of data from a non-mined stream, overview and timeline of Mosaic mining activities in the Horse Creek Basin, and an assessment of potential impacts on the biology of Horse Creek.

Analysis and Discussion

TREND ANALYSIS WITH ADDITIONAL DATA

This impact assessment was developed because the 2012 HCSP Annual Report found several statistically significant trends in water quality parameters over time. In past HCSP annual reports, the Seasonal Kendall Tau method was determined to be the most appropriate method for monotonic trend detection. The Seasonal Kendall test determines water quality trends after correcting for seasonality by only comparing values between similar seasons over time. This test will produce a test statistic and median slope, which is a measure of a monotonic trend. The Seasonal Kendall Tau test can include LOWESS smoothing for parameters that are influenced by streamflow or rainfall. The Annual Kendall Tau test is similar, but it is a nonparametric test for monotonic trends in which only annual median values are used.

The Seasonal Kendall Tau test is limited in several ways. The slope is a measure of the monotonic trend; the actual temporal variation may include step trends or trend reversals. If alternate seasons exhibit trends in opposite directions, the Seasonal Kendall test will not detect an overall trend. In addition, limited years of data will decrease the power of the test to detect trends of small magnitude. Any changes over time detected using the Seasonal Kendall test should be further examined to determine if the perceived change is caused by a data bias, if its magnitude is ecologically significant, or if the cause of the trend is related to Mosaic mining activities.

The results of the Seasonal Kendall Tau for the 2012 Annual Report are given in Table 1. Cells highlighted in yellow indicate ten parameters where a significant slope was found for at least one station, and data collected by the HCSP for these parameters from 2003 to 2012 is shown in Figures 1 - 11. Ten water quality parameters had a statistically significant trend at HCSW-1 and three parameters had a statistically significant trend at HCSW-4. Several of the trends detected during the statistical analysis have an estimated slope that 1) was not in the direction of an adverse trend[1] (color, iron, ammonia) or 2) was very small compared to limits in laboratory prediction or the observed differences between primary and duplicate field samples (pH slope of 0.05 SU/yr; ammonia slope of -0.0003 mg/L/yr; orthophosphate slope of 0.02 mg/L/yr). For the trends with higher estimated rates of change (specific conductivity and various dissolved ions), the potential trends are further discussed in this impact assessment, with a focus on the station closest to mining (HCSW-1

Specific conductivity, which has a longer period of record with more consistent data collection, is used as a surrogate[2] for the other dissolved ions (calcium, alkalinity, sulfate, and TDS) in this impact assessment. The potential trend in orthophosphate is also discussed in more detail below (even though the estimated slope is very small (0.02/mg/L/yr)) as a follow-up to the impact analysis (Appendix I) in the 2010 and 2011 annual reports (Robbins et al. 2014 a and b). Tables 2 and 3 highlight parameters where trends have been identified in the 2003-2010, 2003-2011, and 2003-2012 data sets for HCSW-1 and HCSW-4. For some parameters, there has been relatively little if any change in the slope (HCSW-1: pH, ammonia, orthophosphate, and iron; HCSW-4: iron and alkalinity), while at others there has been a gradual decrease in the slope each year as more data is added (HCSW-1: specific conductivity, calcium,

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[1] From the HCSP Plan Document, Appendix A, p. A-3 to A-4: "Since the purpose of this monitoring is to detect trends toward the trigger values, should they be present, trend analyses and other statistical tests will generally focus only upon changes toward the trigger values."

[2] From USEPA. Volunteer Stream Monitoring: A Methods Manual. Office of Water 4503F. EPA 841-B-97-003. November 1997, pg 179. "Conductivity in water is affected by the presence of inorganic dissolved solids such as chloride, nitrate, sulfate, and phosphate anions (ions that carry a negative charge) or sodium, magnesium, calcium, iron, and aluminum cations (ions that carry a positive charge)."

alkalinity, and TDS; HCSW-4: color). Fluoride at HCSW-1 and orthophosphate at HCSW-4 did not exhibit a significant increasing or decreasing trend during the 2003-2012 time period even though they had previously (Tables 2 and 3). The reduction of the trend slope and elimination of certain trends over time supports our conclusions from the 2010 and 2011 annual reports that most of these parameters experienced a step-change over this time period, rather than a continuing, adverse, monotonic trend.

Table 1. Summary of Seasonal Kendall-tau with LOWESS (F=0.5) for HCSW-1 and HCSW-4 from 2003-2012 Using HCSP Data Unless Otherwise Noted.

Parameter	HCSW-1				HCSW-4			
	tau	p-value	slope	2012 Median	tau	p-value	slope	2012 Median
pH	0.42	**0.004**	**0.05**	7.62	0.03	0.79	N/A	7.39
Dissolved Oxygen	-0.07	0.68	N/A	7.69	-0.17	0.256	N/A	7.18
Turbidity	0.08	0.61	N/A	4.32	-0.67	0.68	N/A	2.92
Color, total	0.32	**0.03**	**5.25**	120	0.50	**0.001**	**10.6**	80.0
Nitrogen, total	0.07	0.68	N/A	0.99	0.11	0.47	N/A	1.55
Nitrogen, total Kjeldahl	0.04	0.84	N/A	0.86	0.19	0.22	N/A	0.96
Nitrogen, ammonia*	-0.56	**0.05**	**-0.0003**	0.01	0.17	0.60	N/A	0.05
Nitrogen, nitrate-nitrite*	0.11	0.72	N/A	0.05	0.02	1.00	N/A	0.36
Orthophosphate	0.36	**0.01**	**0.02**	0.42	0.16	0.30	N/A	0.46
Chlorophyll a[1]	-0.13	0.40	N/A	0.72	-0.10	0.50	N/A	1.80
Specific Conductance	0.51	**0.004**	**10.6**	320	0.16	0.30	N/A	189
Calcium, dissolved	0.51	**0.0004**	**1.05**	24.0	0.20	0.18	N/A	78.0
Iron, dissolved	-0.35	**0.02**	**-0.02**	0.15	-0.35	**0.02**	**-0.01**	0.08
Alkalinity	0.42	**0.004**	**2.96**	58.4	0.53	**0.0003**	**1.66**	73.0
Chloride	0.26	0.08	N/A	18.9	0.20	0.18	N/A	31.2
Fluoride*	0.07	0.86	N/A	0.55	-0.11	0.72	N/A	0.52
Sulfate	0.32	**0.03**	**2.27**	62.5	0.13	0.41	N/A	189
Total Dissolved Solids	0.35	**0.02**	**6.64**	216	0.23	0.12	N/A	478
Radium, total[1]	-0.81	0.60	N/A	0.80	-0.09	0.56	N/A	1.60

*SWFWMD data was used from April 2003-December 2012 (some parameters missing 2007 data). Annual Mann Kendall with LOWESS was used for analysis of 2003-2012 data since sampling reduced to every other month starting October 2011.

[1]Data was not correlated with streamflow for either station; LOWESS was not used.

Table 2. Summary of Seasonal Kendall-tau with LOWESS (F=0.5) at HCSW-1 for 2003-2010, 2003-2011, and 2003-2012 Time Periods Using HCSP Data Unless Otherwise Noted.

Parameter	HCSW-1											
	2003-2010				2003-2011				2003-2012			
	tau	p-value	slope	2010 Median	tau	p-value	slope	2011 Median	tau	p-value	slope	2012 Median
pH	0.41	**0.02**	**0.06**	7.25	0.32	**0.05**	**0.05**	7.3	0.42	**0.004**	**0.05**	7.62
Color, total	0.21	0.22	N/A	155	0.24	0.13	N/A	160	0.32	**0.03**	**5.25**	120
Nitrogen, ammonia*	-0.36	**0.04**	**-0.002**	0.01	-0.32	**0.05**	**-0.002**	0.02	-0.56	**0.05**	**-0.0003**	0.01
Orthophosphate	0.5	**0.003**	**0.27**	0.45	0.5	**0.001**	**0.02**	0.31	0.36	**0.01**	**0.02**	0.42
Specific Conductance	0.57	**0.001**	**16.7**	432	0.54	**0.001**	**14.6**	323	0.51	**0.004**	**10.6**	320
Calcium, dissolved	0.51	**0.004**	**1.60**	33.1	0.54	**0.006**	**1.37**	23.1	0.51	**0.0004**	**1.05**	24
Iron, dissolved	-0.29	0.1	N/A	0.21	-0.33	**0.04**	**-0.02**	0.18	-0.35	**0.02**	**-0.02**	0.15
Alkalinity	0.57	**0.001**	**4.19**	70.8	0.43	**0.007**	**3.91**	49.2	0.42	**0.004**	**2.96**	58.4
Fluoride*	0.43	**0.01**	**0.01**	0.61	0.33	**0.04**	**0.01**	0.43	0.07	0.86	N/A	0.55
Sulfate	0.29	0.1	N/A	116	0.33	**0.04**	**2.82**	65.6	0.32	**0.03**	**2.27**	62.5
Total Dissolved Solids	0.38	**0.03**	**10.66**	343	0.43	**0.01**	**9.65**	218	0.35	**0.02**	**6.64**	216

*SWFWMD data was used from April 2003-December 2012 (some parameters missing 2007 data). Annual Mann Kendall with LOWESS was used for analysis of 2003-2012 data since sampling reduced to every other month starting October 2011.

Table 3. Summary of Seasonal Kendall-tau with LOWESS (F=0.5) at HCSW-4 from 2003-2010, 2003-2011, and 2003-2012 Time Periods Using HCSP Data Unless Otherwise Noted.

Parameter	HCSW-4											
	2003-2010				2003-2011				2003-2012			
	tau	p-value	slope	2010 Median	tau	p-value	slope	2011 Median	tau	p-value	slope	2012 Median
Color, total	0.43	**0.01**	**12.07**	150	0.48	**0.002**	**11.47**	140	0.5	**0.001**	**10.6**	80
Orthophosphate	0.38	**0.03**	**0.02**	0.41	0.22	0.16	N/A	0.4	0.16	0.3	N/A	0.46
Iron, dissolved	-0.19	0.28	N/A	0.2	-0.35	**0.03**	**-0.01**	0.13	-0.35	**0.02**	**-0.01**	0.08
Alkalinity	0.52	**0.002**	**1.62**	43	0.48	**0.002**	**1.31**	56.7	0.53	**0.0003**	**1.66**	73

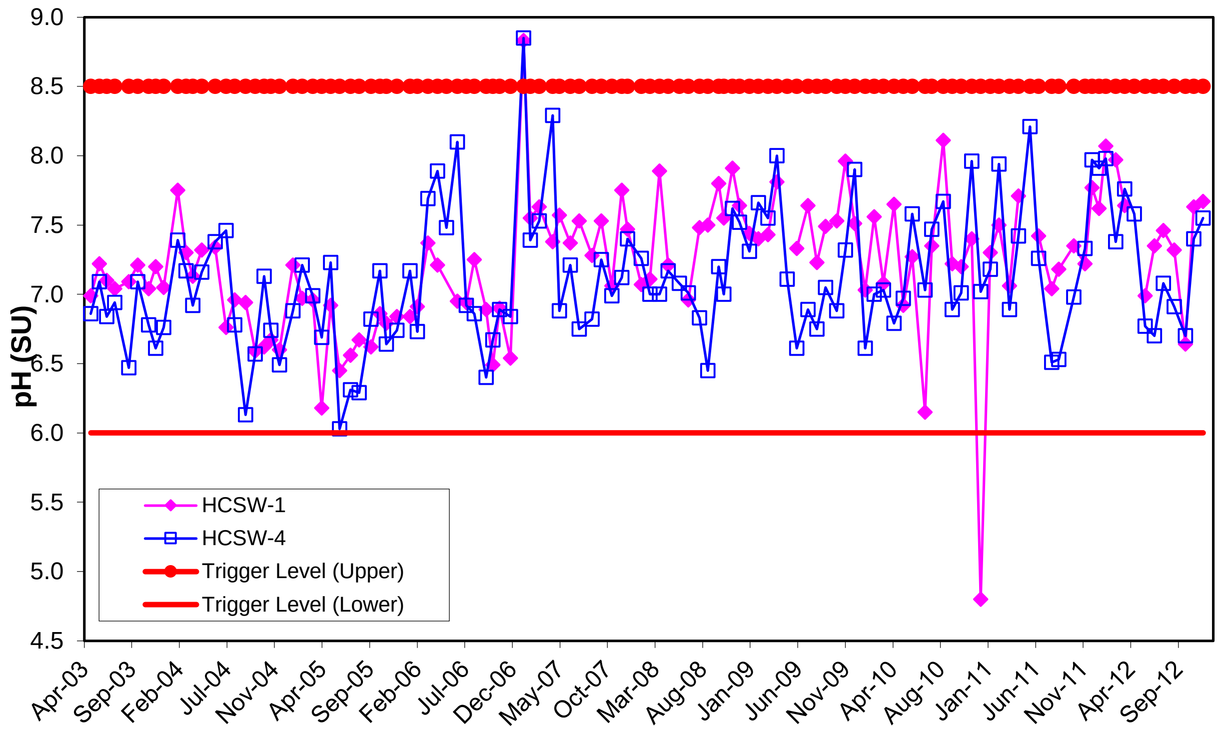

Figure 1. **Values of pH Obtained During Monthly HCSP Water Quality Sampling from 2003-2012. Minimum Detection Limit – 1 su.**

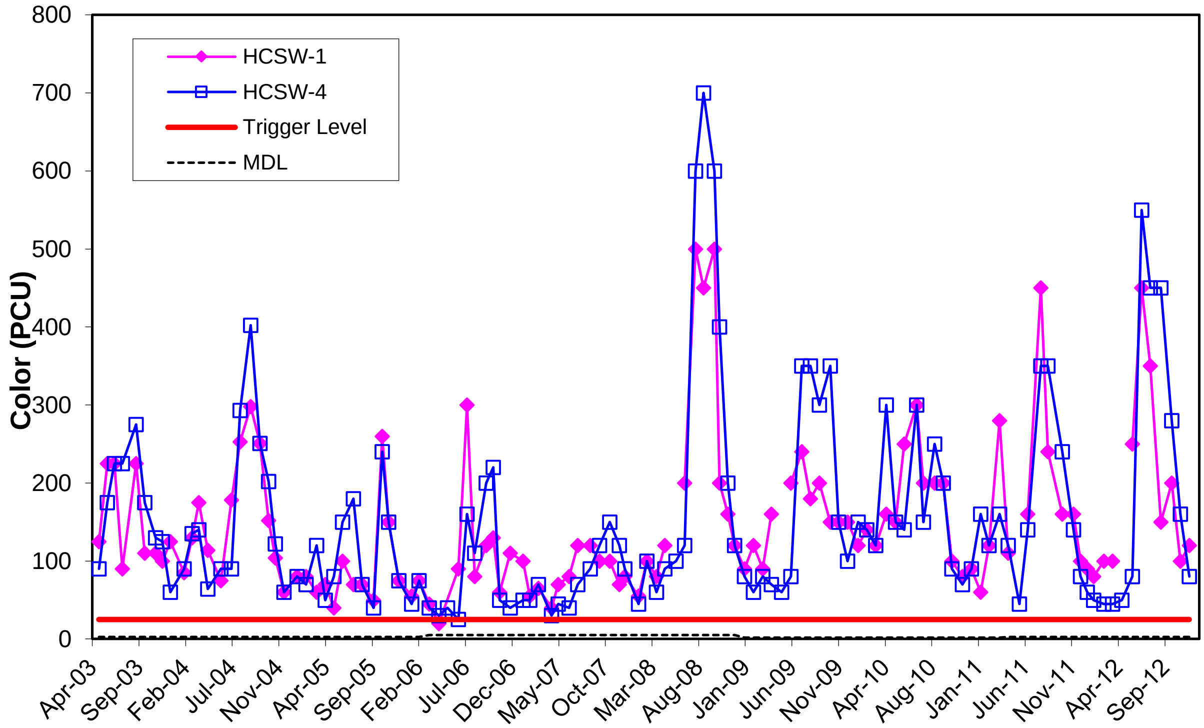

Figure 2. **Color Levels Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.**

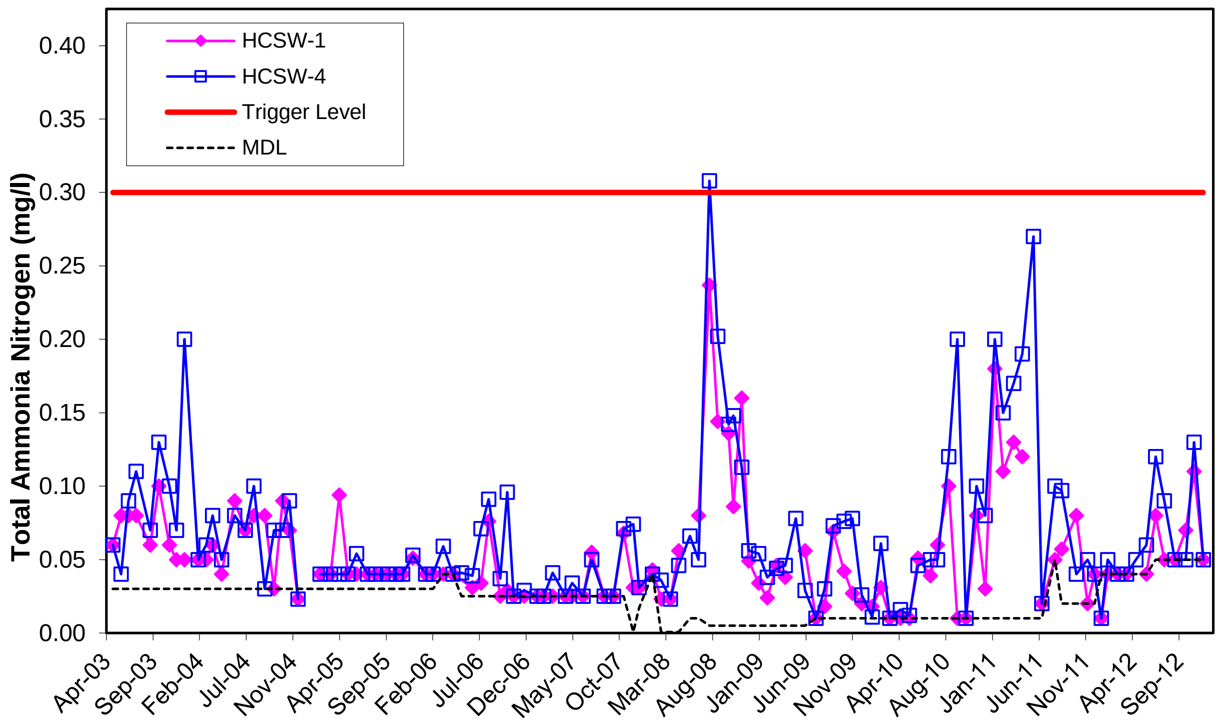

Figure 3. **Total Ammonia Nitrogen Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.**

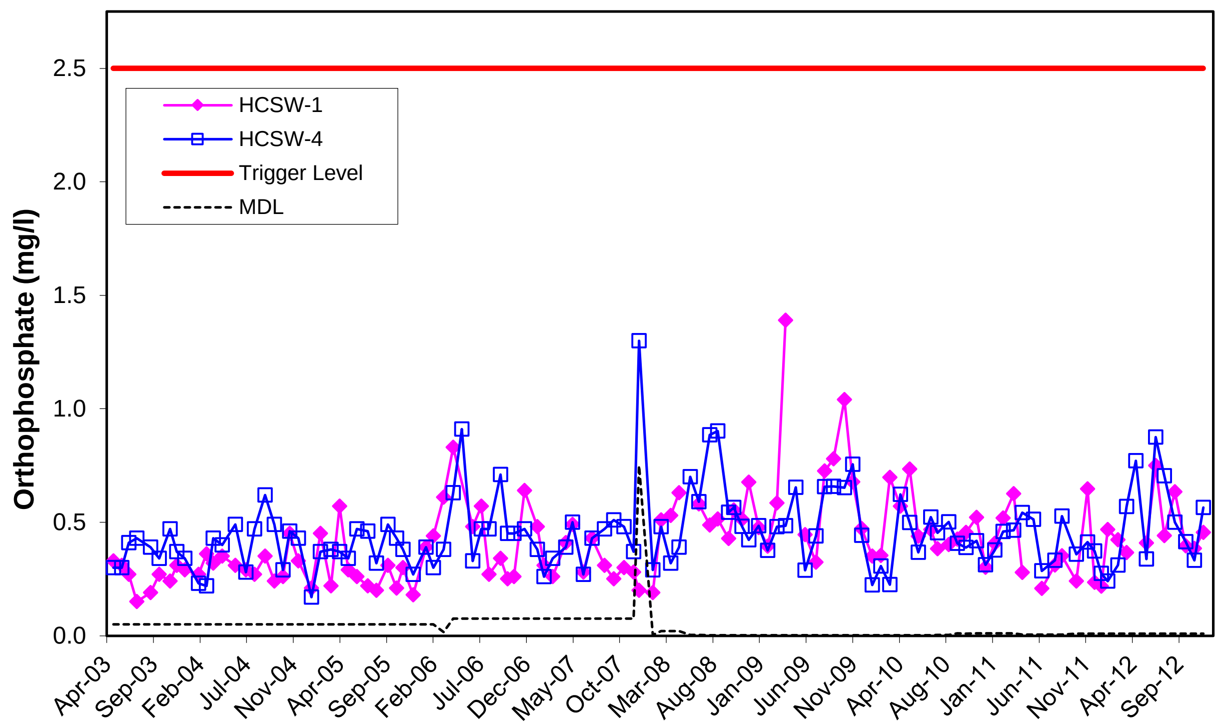

Figure 4. **Orthophosphate Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.**

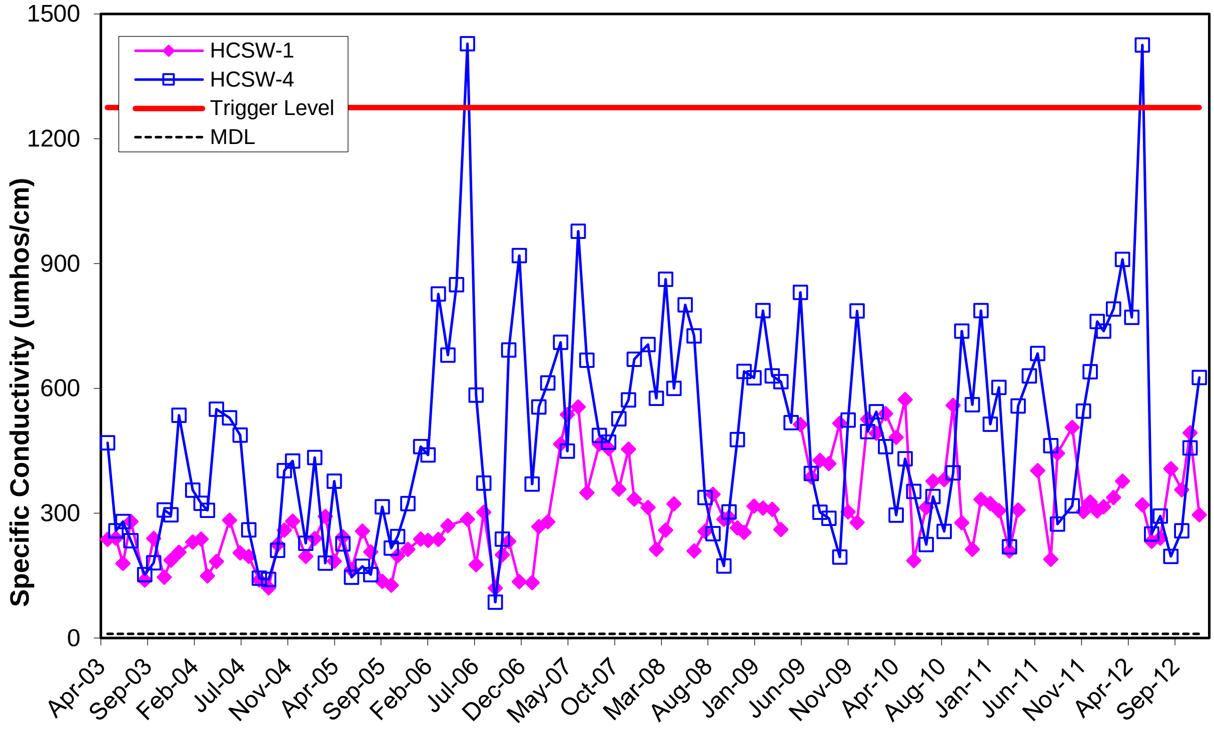

Figure 5. Levels of Specific Conductivity Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

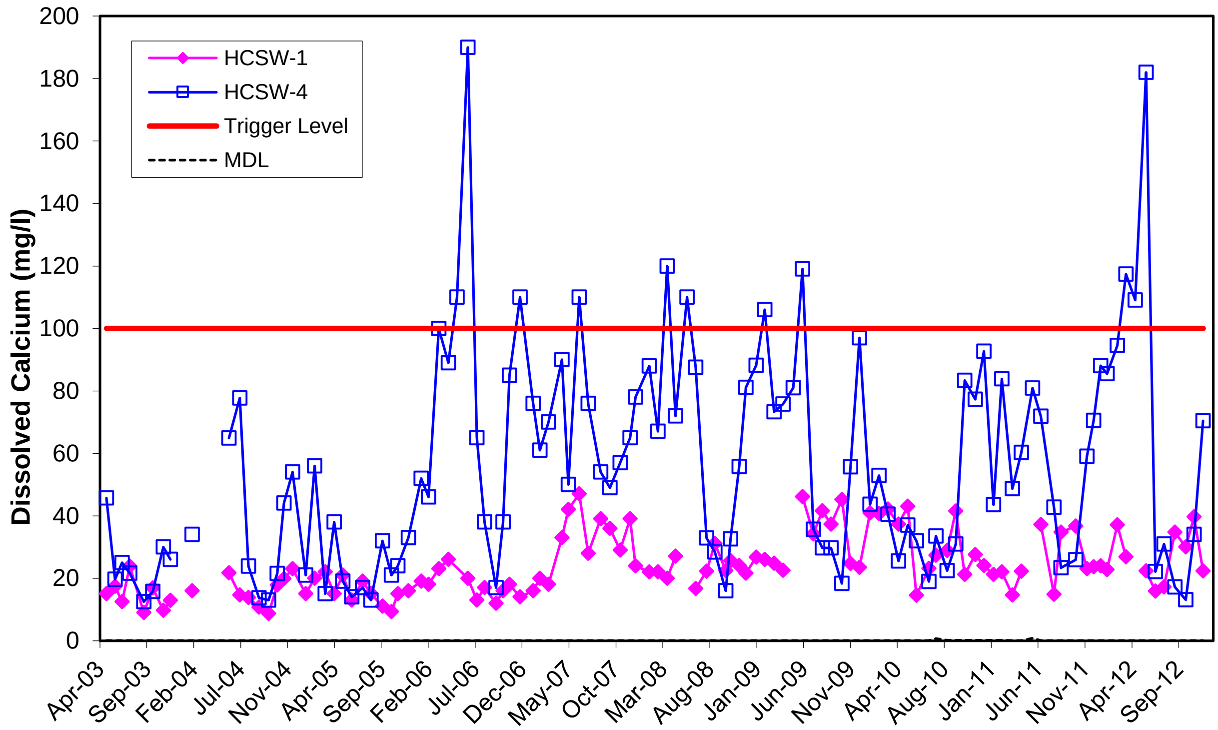

Figure 6. Dissolved Calcium Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

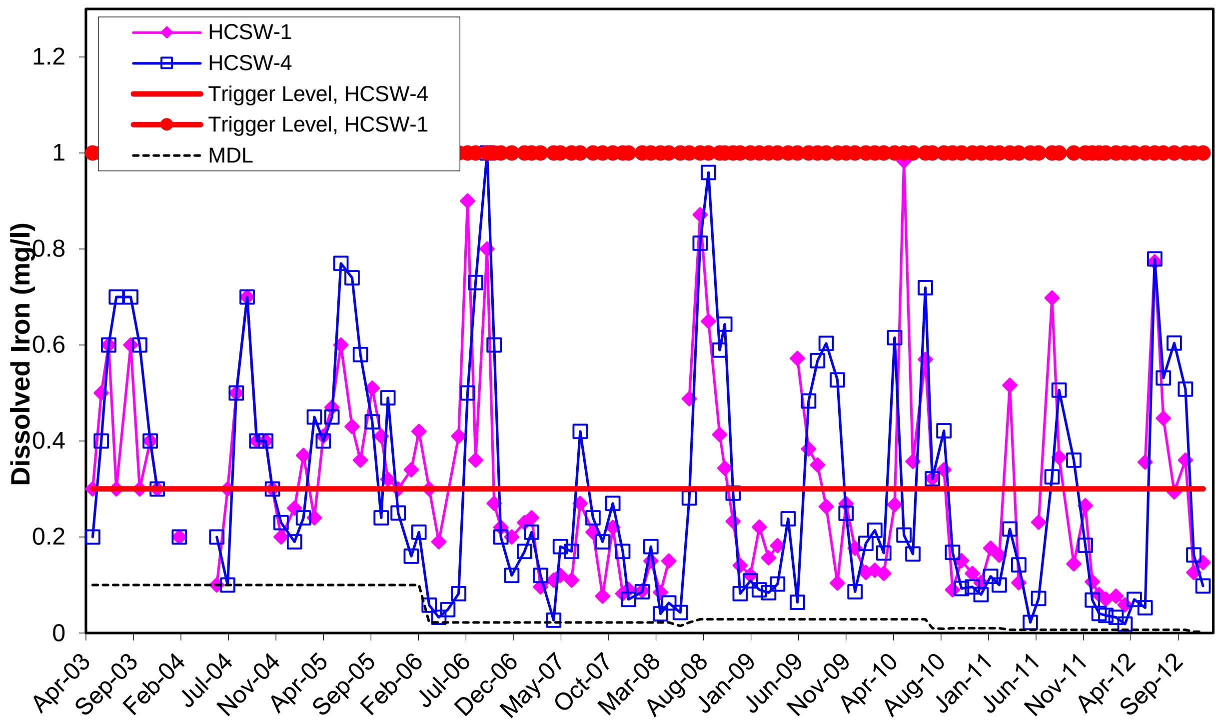

Figure 7. Dissolved Iron Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

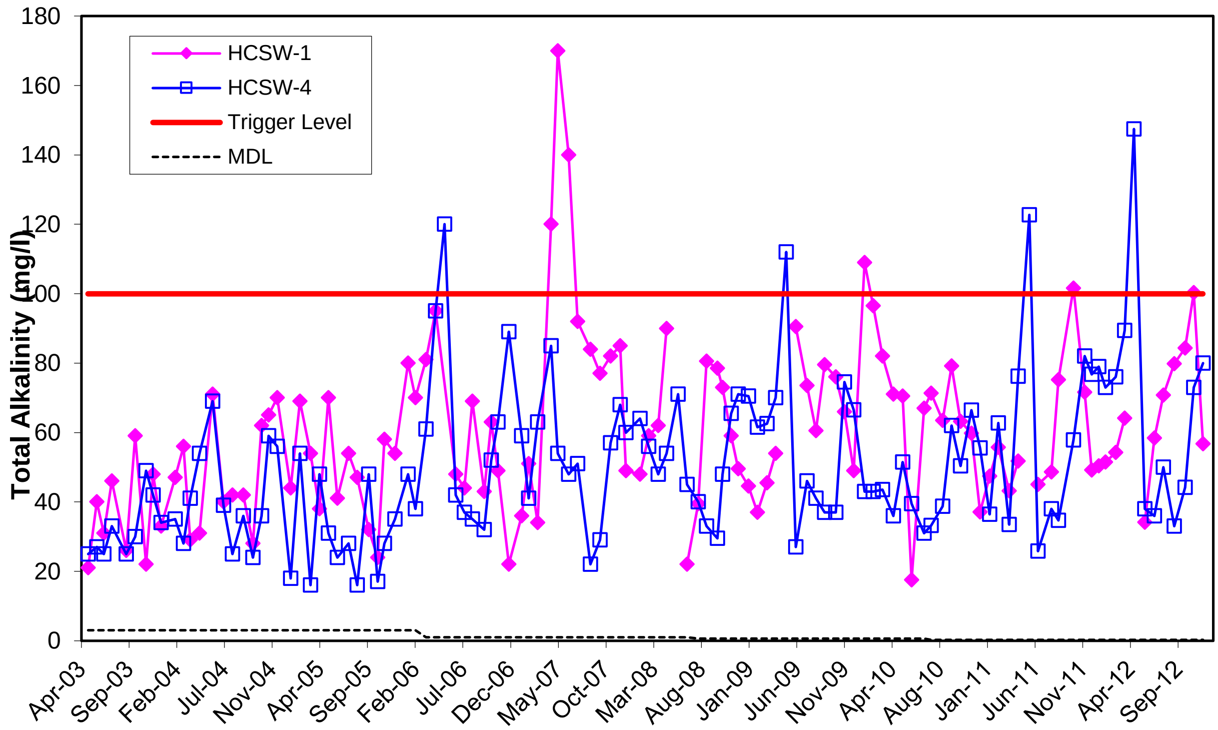

Figure 8. Total Alkalinity Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

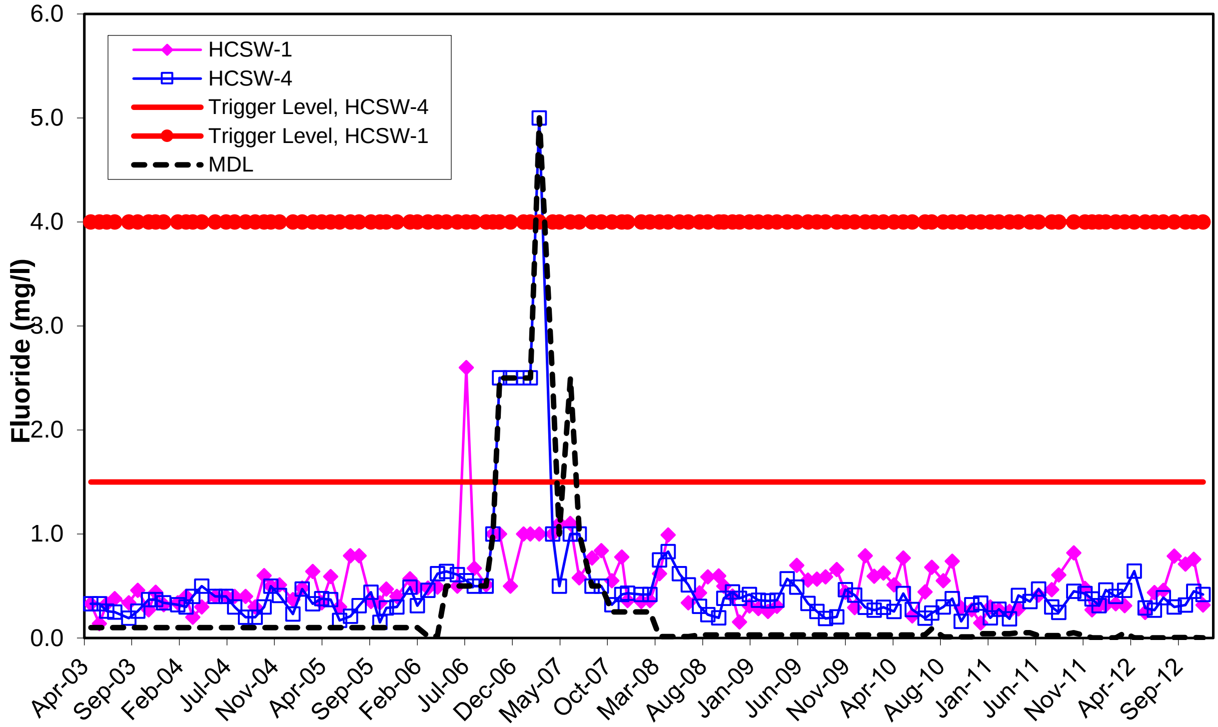

Figure 9. Fluoride Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

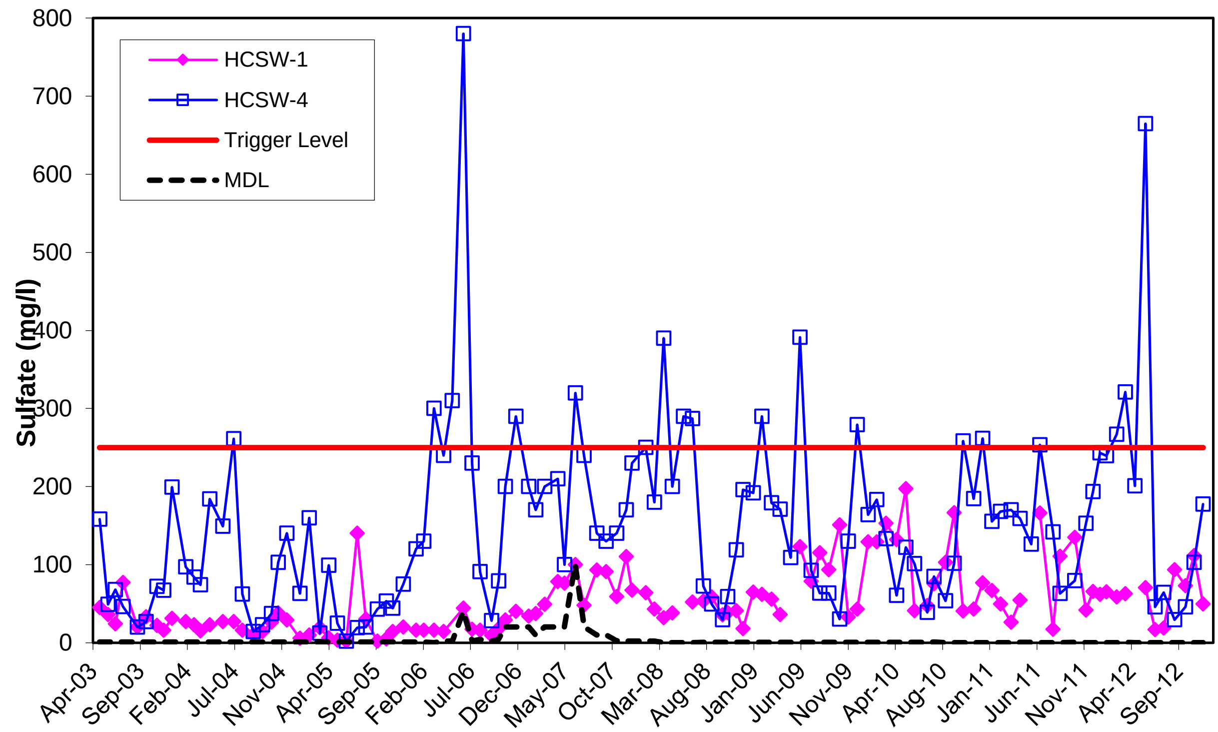

Figure 10. Sulfate Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

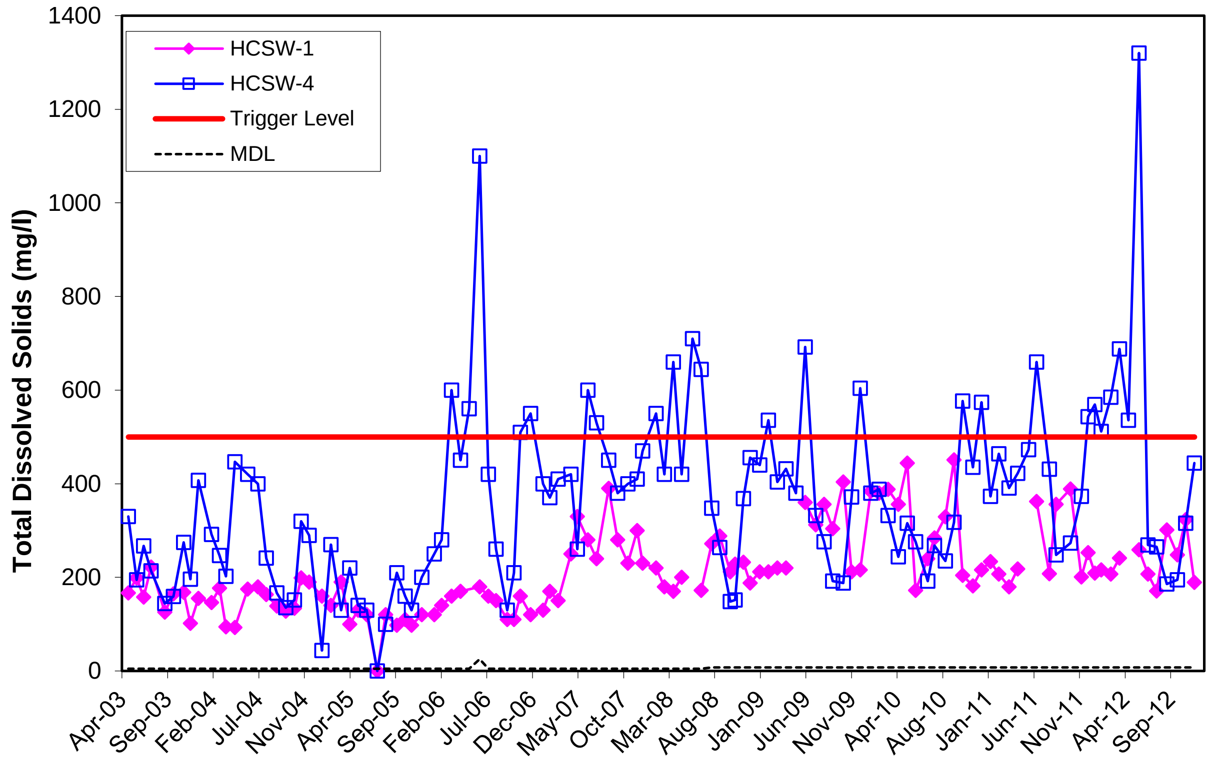

Figure 11. Total Dissolved Solids Concentrations Obtained During Monthly HCSP Water Quality Sampling from 2003-2012.

Any statistical method for trend detection is inherently biased by the time period used in the analysis. For instance, many water quality parameters may be heavily influenced by climatic conditions that are dissimilar at the beginning and end of the time period under analysis. For the 2012 HCSP Annual Report trend analysis, the Seasonal Kendall Tau covered the time period from the beginning of the HCSP (2003) through 2012. In order to investigate if the time constraint resulted in some of the observed trends, we used orthophosphate and specific conductivity data collected by SWFWMD to expand the period of record into the past.

To ensure that our assumptions did not unduly influence this new trend analysis, we tried multiple combinations of time periods, smoothing parameters, data sources, and types of Kendall Tau analysis. For the scenarios with LOWESS smoothing, we tried the following smoothing parameters: HCSW-1 USGS streamflow (same data as found in all annual reports), HCSW-1 flow minus NPDES discharge, or rainfall from one of three NOAA gauges (148, 336, or 219).

For Seasonal Kendall Tau scenarios for the SWFWMD data, 1998 had the earliest consistently collected data for orthophosphate and specific conductivity. In October 2011, SWFWMD went from sampling monthly to every other month, making the 2012 data frequency inconsistent with the rest of the period of record. Therefore, we looked at potential trends in the SWFWMD data using the Seasonal Kendall Tau from 1998-2011 and 2003-2011 (time period of the HCSP). For the Annual Kendall Tau, we were able to go back to 1992 in the SWFWMD data; the time periods used were 1992-2011, 1998-2011, and 2003-2011. For the HCSP data, all analyses were from 2003-2012.

Tables 4 and 5 present the results of the Seasonal and Annual Kendall Tau analyses of SWFWMD and HCSP orthophosphate and specific conductivity data, with statistically significant trends in bold. Orthophosphate does not show a statistically significant upward trend when considered over an expanded data time period longer than 2003-2012 (Table 4). Although orthophosphate concentrations in 2010-2012 are higher than in 2003, they are within the same range of concentrations observed in years prior to the beginning of the HCSP (Figure 12). This observation therefore suggests that the significant trend in orthophosphate found in the 2012 annual report analysis is due to a shorter data time-period used in the analysis and is likely caused by a data-bias caused by the specific conditions occurring at the beginning of the HCSP compared to those conditions occurring later in the data period. This is supported by the very small estimated slope in the 2012 trend analysis for orthophosphate at HCSW-1 (0.02 mg/L/yr), which is within the uncertainty of individual measurement error (Table 1). From the 2010 annual report to this 2012 annual report, the estimated orthophosphate slope was reduced from 0.27 mg/L/yr (2003-2010) to 0.02 mg/L/yr (2003-2012) because concentrations have remained consistent after 2008 (Table 2).

Observations from the expanded data period also show that orthophosphate concentrations increased after 2003 but are not increasing to levels higher than previously measured at the same station. Over the time period shown in Figure 12, climate conditions and mining practices have varied. For example, average monthly streamflow at HCSW-1 was much lower than normal during 1997, 1999-2000, and 2006-2008, which coincide (with some time lag) with most of the peaks in orthophosphate (Figure 12). In addition to those climatic fluctuations in water quality and quantity, the mining recirculation system that discharges to Horse Creek may be connected to active mining or limited to stormwater discharges during different time periods (see further discussion below). The levels of orthophosphate seen in 2008 to 2012 are consistent with previously recorded variation in the system and are unlikely to fluctuate to higher levels. Again, the very small estimated slope for the potential orthophosphate trend in the 2012 report (Table 1) supports the conclusion that orthophosphate conditions at HCSW-1 are fluctuating within a range of concentrations over time, but are not consistently increasing.

The statistically significant upward trend in specific conductivity in Horse Creek is apparent across multiple data time periods, data sources, and analysis methods (Table 5, Figure 13). The predicted median slope of these analyses indicated a potential increase in specific conductivity of 8 to 26 µmhos/cm/year, if we accept the assumption that specific conductivity is exhibiting a monotonic (or one directional) increasing trend. However, examining Figure 13 in more detail provides evidence of several step-changes in conductivity at HCSW-1. From 2003-2006, conductivity consistently ranged between 100-300 µmhos/cm, with several increasing and decreasing step changes occurring before 2003. From 2007-2012, concentrations have been consistently between 200-600 µmhos/cm. Given that the increase in conductivity over time is not a monotonic trend; the Seasonal Kendall Tau may not be an appropriate method to examine the magnitude of the change over time (Figure 13). The data shown in Figure 13 does show increased conductivity levels within the 2007-2012 time period that are above concentrations from 1997-2006. This period also shows that conductivity levels, though higher, are relatively stable at this time. As with orthophosphate, the effects of historical periods of low streamflow (1997, 1999-2000, and 2006-2008) can be seen in Figure 13 as elevated conductivity compared to wetter years. From the 2010 annual report to this 2012 annual report, the estimated conductivity slope was reduced from 16.7 umhos/cm/yr (2003-2010) to 10.6 umhos/cm/yr (2003-2012) because concentrations have remained consistent after 2008 (Table 2).

Table 4. Period of Record Seasonal and Annual Kendall Tau Analyses for Orthophosphate in Horse Creek Samples Collected by SWFWMD and HCSP.

LOWESS Smooth Parameter	Trended Parameter	Stat	SWFWMD – Horse Creek Near Myakka Head					HCSP – HCSW-1	
			Seasonal 1998-2011	Seasonal 2003-2011	Annual 1992-2012	Annual 1998-2012	Annual 2003-2012	Seasonal 2003-2012	Annual 2003-2012
None	Orthophosphate	p-value	0.15	**0.02**	0.33	0.37	**0.05**	**0.01**	0.21
		slope	N/A	**0.023**	N/A	N/A	**0.018**	**0.020**	N/A
Flow at HCSW-1	Orthophosphate	p-value	0.18	**0.02**	0.42	0.43	0.86	**0.01**	0.72
		slope	N/A	**0.020**	N/A	N/A	N/A	**0.019**	N/A

Table 5. Period of Record Seasonal and Annual Kendall Tau Analyses for Specific Conductivity in Horse Creek Samples Collected by SWFWMD and HCSP.

LOWESS Smooth Parameter	Trended Parameter	Stat	SWFWMD – Horse Creek Near Myakka Head					HCSP – HCSW-1	
			Seasonal 1998-2011	Seasonal 2003-2011	Annual 1992-2012	Annual 1998-2012	Annual 2003-2012	Seasonal 2003-2012	Annual 2003-2012
None	Specific Conductivity	p-value	**<0.001**	**<0.001**	**<0.001**	**0.006**	0.07	**<0.001**	**0.02**
		slope	**18.00**	**26.00**	**10.63**	**15.25**	N/A	**16.17**	**15.80**
Flow at HCSW-1	Specific Conductivity	p-value	**<0.001**	**0.0259**	**<0.001**	**0.003**	0.37	**0.004**	0.72
		slope	**11.61**	**11.31**	**8.52**	**11.92**	N/A	**10.6**	N/A

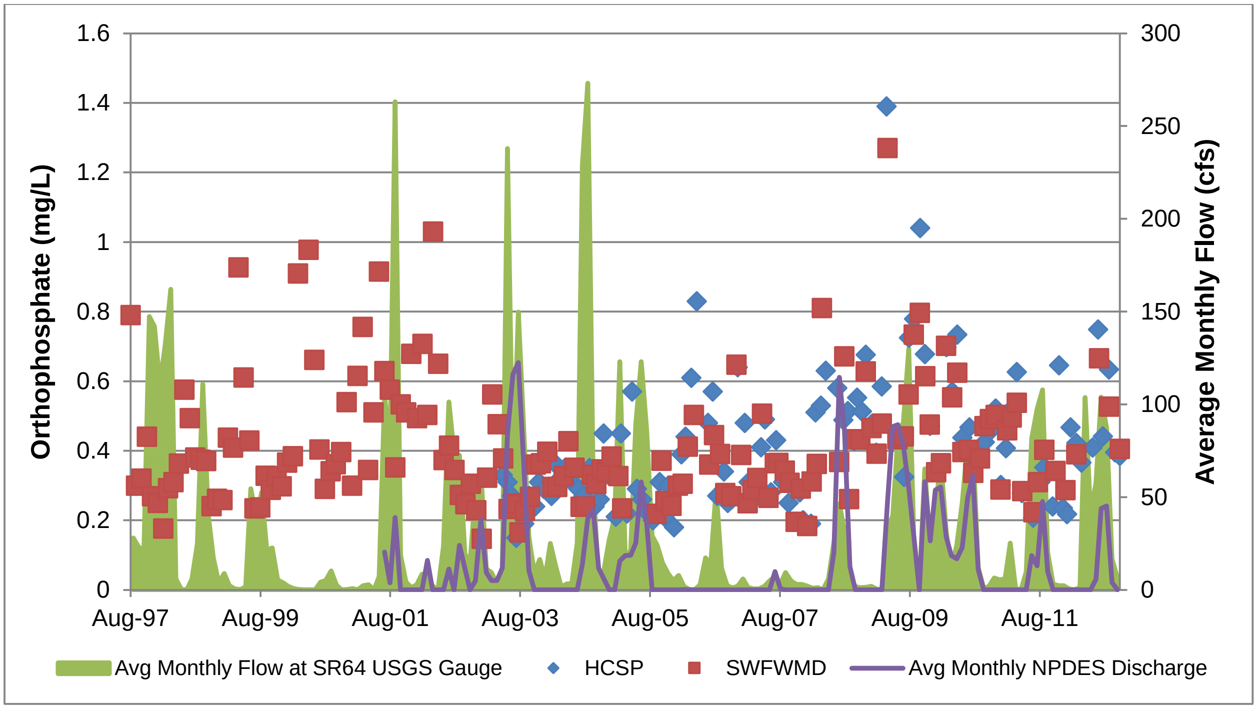

Figure 12. **HCSW-1 Orthophosphate Concentrations Obtained During Monthly HCSP and SWFWMD Water Quality Sampling With Average Monthly Flow from USGS Gauge at HCSW-1 and NPDES Discharge.**

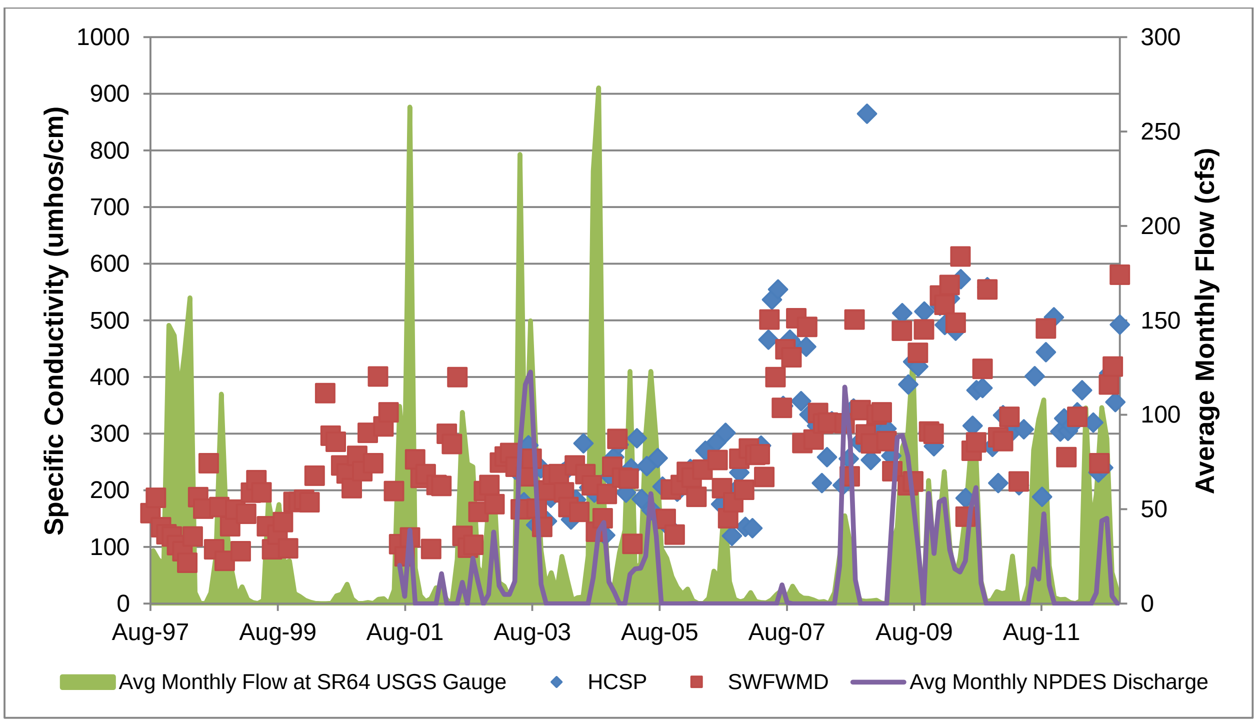

Figure 13. **HCSW-1 Specific Conductivity Concentrations Obtained During Monthly HCSP and SWFWMD Water Quality Sampling With Average Monthly Flow from USGS Gauge at HCSW-1 and NPDES Discharge.**

OTHER STREAMS

To put the Horse Creek results into perspective, we examined potential trends at Charlie Creek, a stream elsewhere in Peace River basin that has not had phosphate mining in its watershed.

Charlie Creek, like Horse Creek, is a part of the Peace River basin. Unlike Horse Creek, the Charlie Creek basin is not influenced by phosphate mining, and thus can provide some insight into potential ways that climate or other land uses may influence water quality in the Peace River system. For the Charlie Creek analysis, we used data collected by FDEP, USGS, and SWFWMD to examine potential trends in total phosphorus and specific conductivity over similar time periods as those used in our Horse Creek analysis in Table 4; total phosphorus was used instead of orthophosphate because of more consistent data collection over time. Tables 6 and 7 present the results of the Seasonal and Annual Kendall Tau analyses with and without LOWESS smoothing by Charlie Creek USGS streamflow. For the Seasonal Kendall Tau analyses, relatively consistent monthly sampling began in 1999, so we looked at trends from 1999-2012 and 2003-2012. For the Annual Kendal Tau, we were able to match the Horse Creek analysis time periods (Table 4) more exactly.

Table 6 and Figure 14 indicate that there is no trend in total phosphorus across all time periods. Specific conductivity shows a significant upward trend in flow-adjusted values from 2003-2012 (seasonally) and 1992-2012 (annually), and would likely show a trend over the entire period of record (1965-2012) if the analysis was expanded (Table 7, Figures 15 and 16). These results indicate that Charlie Creek is may be experiencing a slight increase in specific conductivity over time that is unrelated to mining. The projected rate of increase in conductivity in Charlie Creek is about 8.18 µmhos/cm/yr, which is comparable to the rate for Horse Creek (10.6 µmhos/cm/yr) over the same time period (Seasonal Kendall Tau 2003-2012, Tables 5 and 7). It is possible that whatever is influencing this increase in Charlie Creek, whether it is climate, changes in land use, agriculture irrigation run off, etc., may also be part of what is causing concentrations in Horse Creek to rise. However, the changes in conductivity in Horse Creek are following a different pattern than in Charlie Creek, so regional climatic influences are not the only causes of conductivity increases in Horse Creek.

Table 6. Period of Record Seasonal and Annual Kendall Tau Analyses for Total Phosphorus in Charlie Creek Samples Collected by FDEP, SWFWMD and USGS.

			SWFWMD and FDEP Data		USGS, SWFWMD, and FDEP Data		
LOWESS Smooth Parameter	Trended Parameter	Statistics	Seasonal 1999-2012	Seasonal 2003-2012	Annual 1992-2012	Annual 1998-2012	Annual 2003-2012
None	**Total Phosphorus**	p-value	0.20	0.28	0.72	0.96	0.59
		slope	N/A	N/A	N/A	N/A	N/A
USGS Flow	**Total Phosphorus**	p-value	0.34	0.61	0.61	0.43	0.72
		slope	N/A	N/A	N/A	N/A	N/A

Table 7. Period of Record Seasonal and Annual Kendall Tau Analyses for Specific Conductivity in Charlie Creek Samples Collected by FDEP, SWFWMD and USGS.

			SWFWMD and FDEP Data		USGS, SWFWMD, and FDEP Data		
LOWESS Smooth Parameter	Trended Parameter	Statistics	Seasonal 1999-2012	Seasonal 2003-2012	Annual 1992-2012	Annual 1998-2012	Annual 2003-2012
None	**Specific Conductivity**	p-value	0.19	**0.02**	0.07	**0.05**	**0.03**
		slope	N/A	**16.7**	N/A	**9.33**	**16.7**
USGS Flow	**Specific Conductivity**	p-value	0.08	**0.01**	**0.01**	0.09	0.59
		slope	N/A	**8.18**	**3.21**	N/A	-0.28

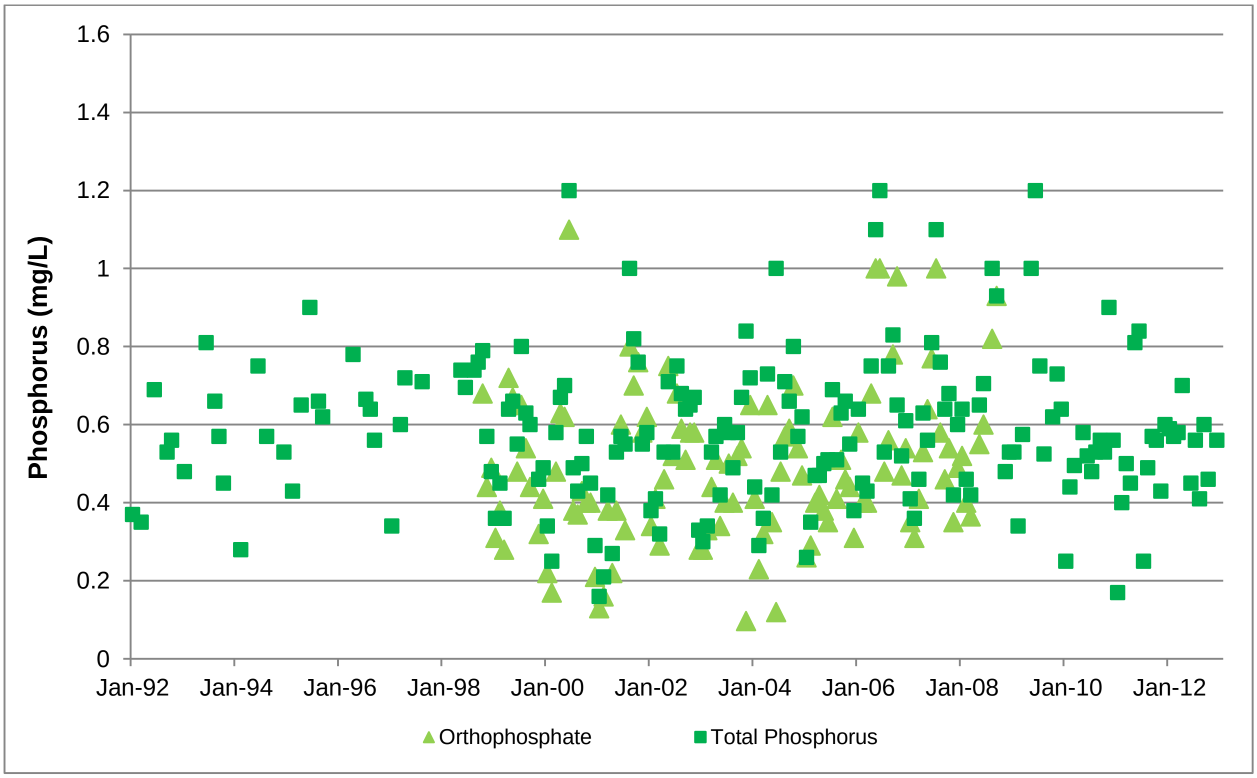

Figure 14. Charlie Creek Monthly Average Total Phosphorus and Orthophosphate Collected by USGS, SWFWMD, and FDEP Water Quality Sampling from 1992-2012.

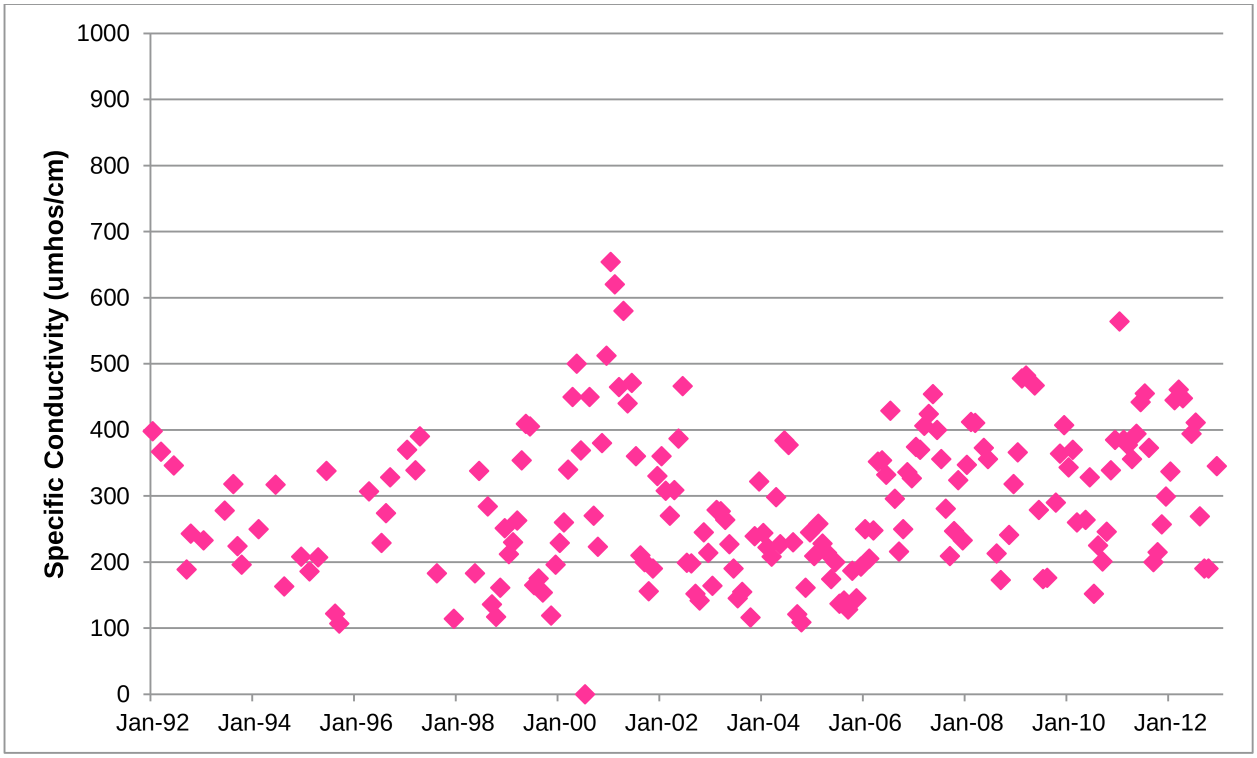

Figure 15. Charlie Creek Monthly Average Specific Conductivity Collected by USGS, SWFWMD, and FDEP Water Quality Sampling from 1992-2012.

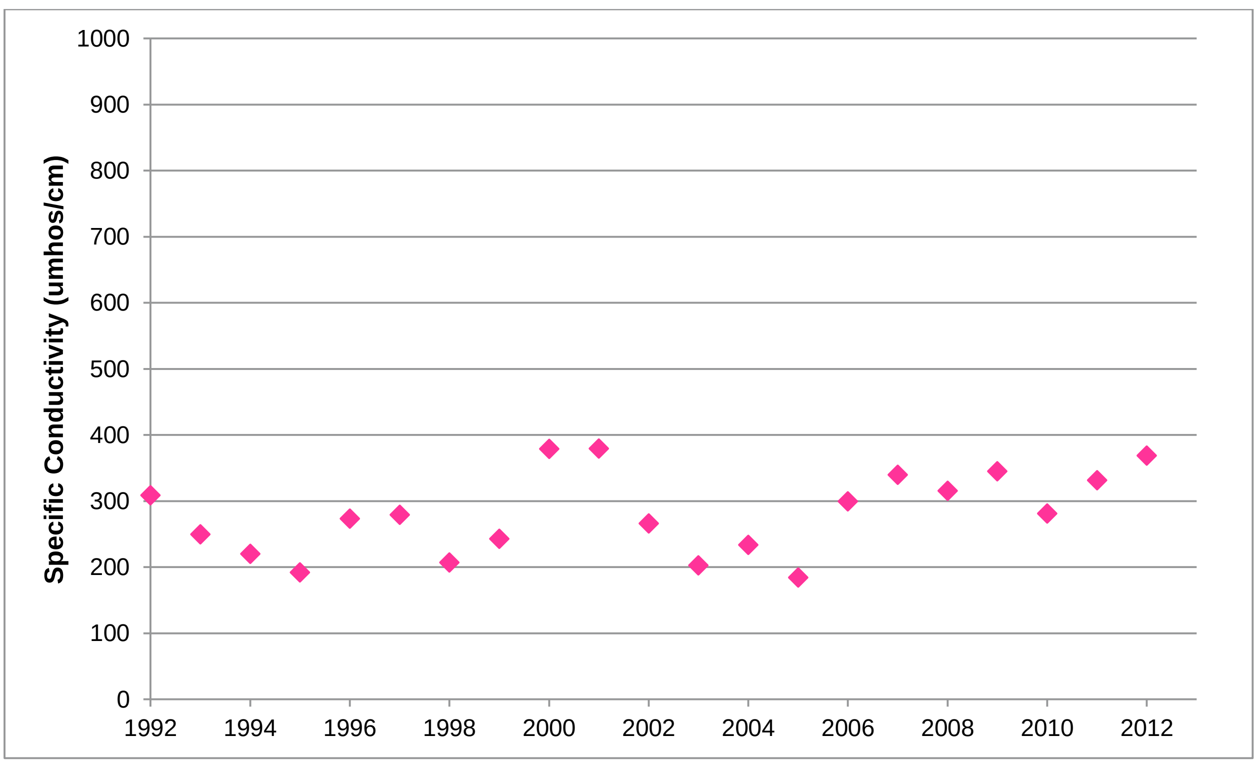

Figure 16. Charlie Creek Annual Average Specific Conductivity Collected by USGS, SWFWMD, and FDEP Water Quality Sampling from 1992-2012.

MINING MILESTONES

Additional trend analyses of specific conductivity data over a longer time period confirm an increasing trend at HCSW-1 since 1992 (Table 5); however, Charlie Creek, which is not influenced by phosphate mining, also shows signs of a long-term, slow upward trend in specific conductivity. In order to evaluate whether Mosaic's mining activities have influenced the increasing trend in specific conductivity at HCSW-1, we examined the history of mining changes in the Horse Creek basin with respect to water quality.

Since mining began in the Horse Creek basin in the late 1980s, mining practices have varied in several important ways. For several years prior to 2006, the NPDES outfalls that discharge into Horse Creek were connected to active clay settling areas that received clays from strip mining conducted in the Four Corners or Fort Green mines. In June 2006, the last clays from Fort Green beneficiation plant were sent to Clay Settling Areas FGH3 and FGH4, which discharge to Horse Creek via FTG-003 and FTG-004. After 2006, the outfalls were not used to release process water for several years because the clay settling areas were not being used to store new clay; in addition, extremely dry conditions during this time period resulted in very little stormwater discharge into Horse Creek via the FTG-003 and FTG-004 outfalls (Figure 17). In October 2008, clays mined by dredge from the Wingate Mine began to be transported to facilities and settling areas (FM1) in the Horse Creek basin for processing and storage.

Figures 18 and 19 show the changes in specific conductivity and orthophosphate at HCSW-1 as mining operations and mine water management have changed. When the Horse Creek outfalls were receiving water from clays collected by strip mining in Fort Green/Four Corners, the specific conductivity ranged between 100 and 400 µmhos/cm (Figure 18). From 2006 to 2008, when the outfalls were discharging only small quantities of surface water, specific conductivity started to increase to between 200 and 500 µmhos/cm; this increase was likely caused by an increased proportion of natural baseflow (groundwater influence) at HCSW-1 during those unusually dry years. When the Horse Creek outfalls began to discharge water from clays collected by Wingate dredge mining, the conductivity remained at these higher levels and increased slightly (200-600 µmhos/cm); dredge mining (used at Wingate mine) relies more on groundwater sources than the previous Ft Green / Four Corners strip mining, which is the likely explanation for the increased levels of conductivity when compared to pre-2006 concentrations. Groundwater influence, whether through climatic or mining activities, is the most likely cause of the statistically significant trend (or more accurately, step-change) in specific conductivity in Horse Creek at station HCSW-1. The conductivity increases in recent years were influenced by drought-period baseflow contributions and more groundwater use by mining activities, so the increase that occurred around 2006 is a step-change instead of a monotonic trend. This means that the slope from the Seasonal Kendall Tau analysis is not an accurate predictor of changes over time at HCSW-1; specific conductivity is likely to stay relatively constant at HCSW-1 given planned mining practices and the stable range of concentrations shown since Wingate mine began affecting the Horse Creek NPDES discharge in 2008.

Including pre-HCSP Horse Creek data in our trend analyses indicated that the apparent orthophosphate trend was very small and was not evident when a longer period of record was used in the analysis; thus, orthophosphate is not a parameter of concern or needing corrective action as of the 2012 reporting year. Concentrations during the pre-HCSP (1997-2001) time period are similar to those measured after Wingate clays began affecting the Horse Creek NPDES outfalls (post-2008), thus reinforcing the conclusion that current mining practices are not causing orthophosphate to increase over the period of record at HCSW-1. Although changes in mining practices may contribute to changes in orthophosphate concentrations over time, the largest peaks in orthophosphate concentrations coincide with periods of lower than average streamflow, with some lag time (1997, 1999-2000, 2006-2008). During those periods of low streamflow, NPDES discharge was extremely limited, which means that elevated orthophosphate concentrations did not originate in NPDES discharge (Figure 12). In summary, orthophosphate does not show a statistically significant upward trend when considered over an expanded data time period longer than 2003-2011 or 2012 (Table 4), and the estimated slope from 2003-2012 was very small (0.02 mg/L/yr). Although orthophosphate concentrations in 2008-2012 are higher than in 2003, they are within the same range of concentrations observed in years prior to the beginning of the HCSP (Figures 12 and 19). This observation suggests that the significant trend in orthophosphate found in the 2012 annual report analysis is due to a shorter data time-period used in the analysis and is likely caused by a data-bias caused by the specific conditions occurring at the beginning of the HCSP compared to those conditions occurring later in the data period. Given the historical variation in this parameter prior to the beginning of the HCSP, current orthophosphate conditions are not causing a degradation of historical stream quality and no further analysis or corrective action for this parameter is needed (Figure 19).

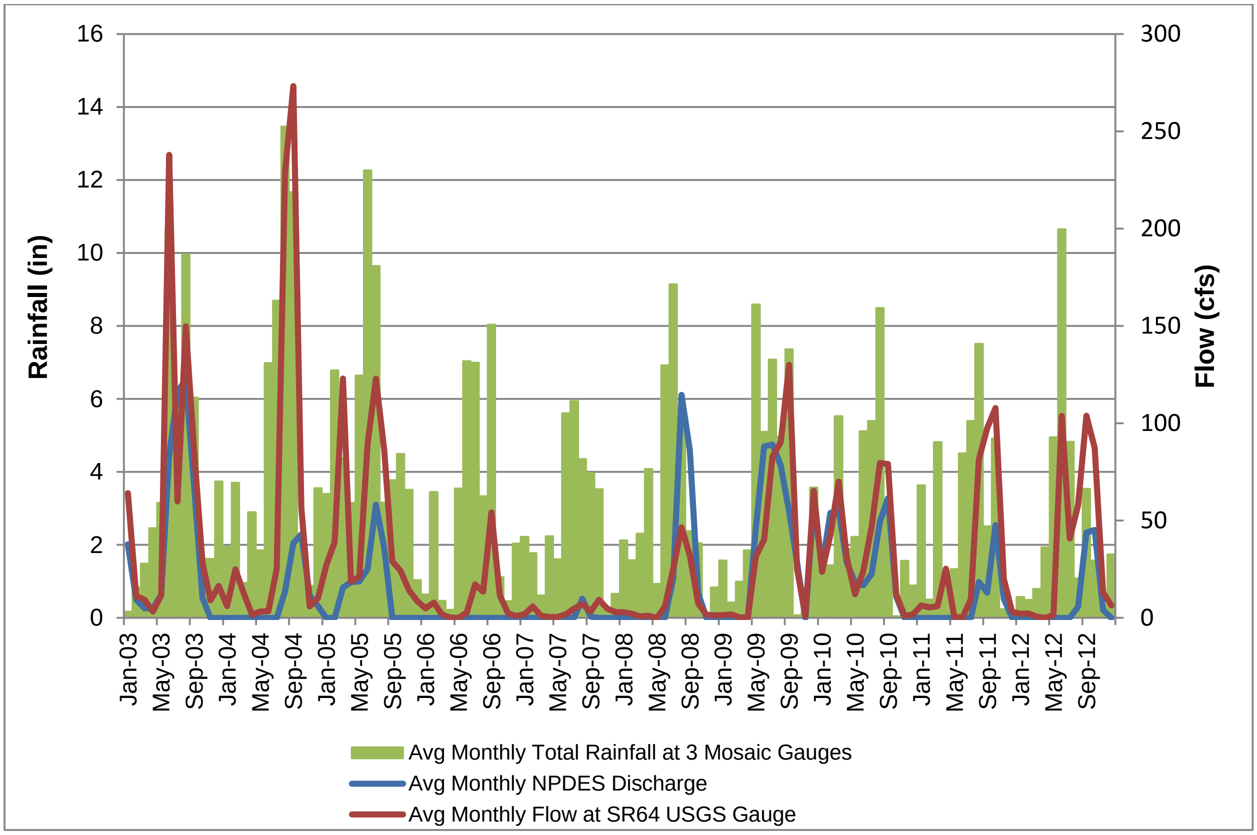

Figure 17. Mosaic NPDES Discharge, Rainfall, and USGS streamflow for HCSW-1 from 2003-2012.

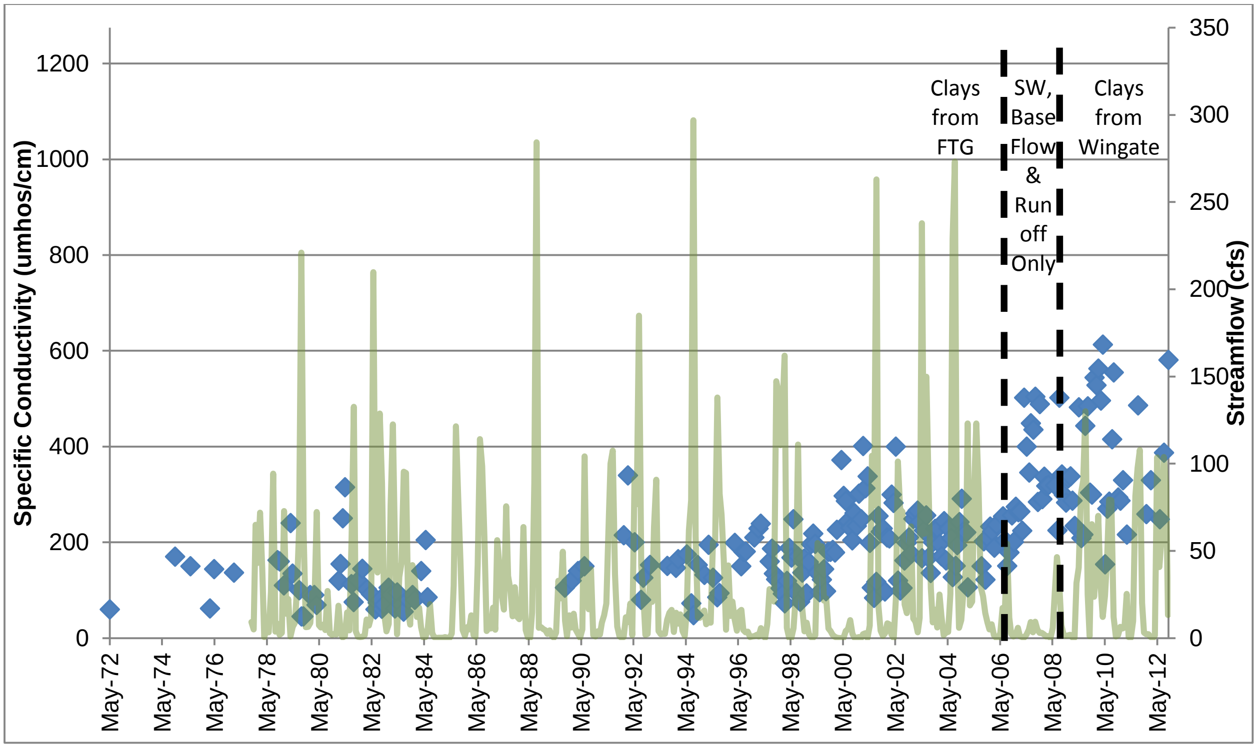

Figure 18. **Specific Conductivity Collected by FDEP, USGS, and SWFWMD Water Quality Sampling from 1972-2012 at HCSW-1, with USGS Streamflow for HCSW-1.**

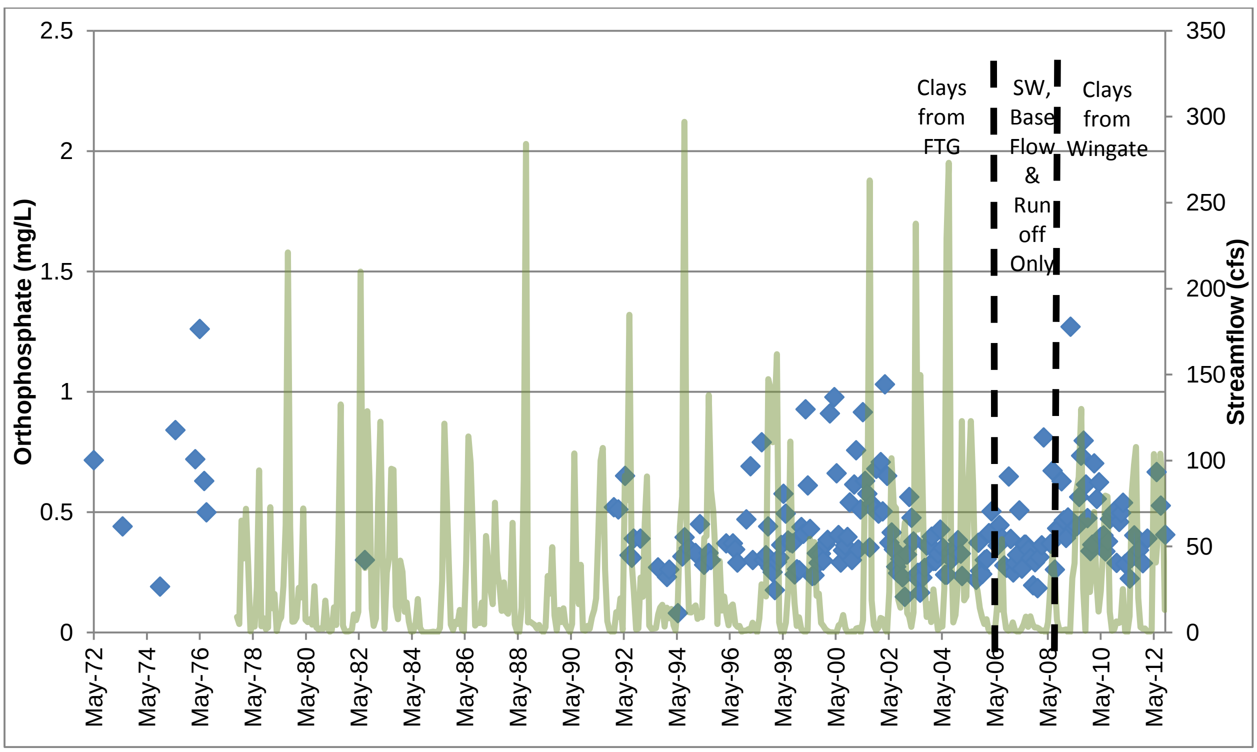

Figure 19. **Orthophosphate Collected by FDEP, USGS, and SWFWMD Water Quality Sampling from 1972-2012 at HCSW-1, with USGS Streamflow for HCSW-1.**

WATER QUALITY STANDARDS AND BIOLOGICAL INTEGRITY

Although there is a statistically significant increasing trend in specific conductivity and some specific dissolved ions in Horse Creek (as well as in Charlie Creek), the magnitude of the trend is not of concern when compared to state drinking water or Class III surface water standards. For fluoride, alkalinity, and specific conductivity (with significant trends in the 2012 Annual Report), all have been well below the applicable Florida Surface Water Class III Standards and Drinking Water Standards through 2012 (Table 8).

Iron exhibited a negative potential trend in 2012 (Table 1), indicating that any potential change is in the opposite direction of the HCSP trigger levels and water quality standards. When compared to water quality standards, HCSW-1 and HCSW-3 have met the Class III standard for iron through 2012. HCSW-2 has had two exceedances of the Class III iron standard (1.2 mg/L in 2006 and 1.03 mg/L in 2009), which were attributed to upstream influence of Horse Creek Prairie by the 2006 impact assessment. HCSW-4, the only station to be assigned a HCSP trigger level equivalent to the iron drinking water standard, frequently exceeds the drinking water standard but not the Class III standard; in a 2003 impact assessment, it was determined that this station historically exceeds the trigger of 0.3 mg/L of iron, without mining influence. Therefore, the four parameters listed in Table 8 do not pose a concern in regards to state water quality standards at this time.

Table 8. HCSP Water Quality Concentrations Compared to Florida Drinking Water and Class III Surface Water Standards for Dissolved Ions.

Parameter	HCSP through 2012	Drinking Water Standard	Class III Standard
Fluoride	2.6 mg/L (Max)	Not > 4.0 mg/L	Not > 10.0 mg/L
Alkalinity	17.5 mg/L (Min)		Not < 20 mg/L (opposite direction as HCSP trigger value)
Specific Conductance	865 µmhos/cm (Max)		Not > 1275 µmhos/cm
Iron	1.2 mg/L (Max)	Not > 0.3 mg/L	Not > 1.0 mg/L. (0.3 mg/L trigger value for HCSP determined to be not consistent with historic values at HCSW-4 in 2003 impact assessment)

In the 2012 Annual Report analysis, orthophosphate exhibited a statistically significant (but very small) trend, although this trend is no longer significant when considered over a longer period of record. In this impact assessment, we examine the compliance of HCSW-1 with recently revised nutrient standards (that were not implemented at the time of data collection). Under the recently approved state numeric nutrient standards, to avoid being listed as impaired for nutrients, a stream must pass a combination of biological and/or numerical criteria. According to 62-302 and 62-303 F.A.C. biological and nutrient data collected during the HCSP at HCSW-1 indicates no nutrient impairment. Table 9 lists some of the ways that HCSW-1 passes nutrient criteria that would otherwise put it on the Planning, Study, or Verified Lists under the Impaired Waters Rule.

According to the FDEP NNC Implementation Document and 62-302.531(2)(c), streams without site-specific criteria with have achieved the nutrient criteria from 62-302.530(47)(b), F.A.C. if:

- There is no imbalance in flora and fauna based on chlorophyll a levels, algal mats or bloom, nuisance macrophyte growth, or changes in algal species composition; AND EITHER
- The average score of two temporally independent (90 days) SCIs is >= 40, with neither of the two most recent SCIs < 35, OR

- The Nutrient Thresholds (0.49 mg/L TP and 1.65 mg/L TN for Horse Creek) expressed as annual geometric means are not exceeded more than once in a 3 year period.

62-302.531(2)(c), F.A.C. does not have specific Biocriteria to define flora and fauna imbalance. FDEP's approach is to determine if floral components at a stream are within the 90th percentile of the EPA reference stream distribution. These floral components include Rapid Periphyton Survey (RPS), community composition of dominant algal taxa, Linear Vegetation Survey (LVS), and chlorophyll a data. These floral metrics are used in a weight-of-evidence approach, although an indication of imbalance for any one of them means that FDEP would conclude that the stream does not meet the NNC.

In addition to the floral metrics, SCI, and nutrient thresholds listed above for attainment of NNC under 62-302.351(2)(c), streams may be added to the Planning List (62-303.351 F.A.C.), Study List (62-303.390 F.A.C.), and/or Verified List (62-303.450 F.A.C.) under the Impaired Waters Rule for failing the Biological Health Assessment (62-303.330 F.A.C.), having algal mats or blooms, chlorophyll a exceedances of 20 ug/L for more than one annual geometric mean in three year period, or statistically significant trends in total phosphorus, total nitrogen, or chlorophyll.

As of December 2012, HCSW-1 likely meets the numeric nutrient criteria set in 62-302.351(2)(c). Even though the annual geometric mean total phosphorus is greater than the regional threshold of 0.49 mg/L, Horse Creek at HCSW-1 likely meets the criteria because it shows no imbalance of flora and fauna (chlorophyll, RPS, and LVS) and has healthy benthic macroinvertebrate conditions (SCI). (However, the passing scores for RPS and LVS assessment in 2012 are inconclusive because there were not two independent samples.) HCSW-1 shows no evidence of persistent algal blooms, has annual geometric mean concentrations of chlorophyll < 3.2 µg/L, and has one passing Rapid Periphyton Surveys (RPS) and Linear Vegetation Surveys (LVS) in 2012. The HCSW-1 average of SCI scores is > 40, with neither of the two most recent scores < 35. In addition, there are no statistically significant increasing trends for total phosphorus, once the confounding factor of streamflow is removed. HCSW-1 also meets the SCI portion of the Biological Heath Assessment in 62-303.330 with the two most recent SCI scores > 35 and within 20 points of the historic maximum (if the historic maximum is above 64).

Table 9. Selected Criteria for Class III Surface Water Nutrient Standards Compared to HCSW-1 Results Through 2012.

Parameter	Criteria for Passing	HCSW-1 Results
Numeric Nutrient Criteria: Floral Metrics (62-302.531(2)(c), F.A.C., FDEP NNC Implementation)	RPS rank 4-6 <= 25%; if 20%-25%, no dominant algal species are nutrient enrichment indicators LVS CofC score <= 2.5 and FLEPPC exotic taxa <=25% Annual geomean chlorophyll a <= 3.2 ug/L; or not exceeding 20 ug/L more than once in 3 year period with site-specific evaluation	RPS and LVS sampling began in 2012 with only one independent sample collected; 2012 sample was passing but inconclusive for NNC without second sample. No algal mats or blooms. All chlorophyll annual geomean < 3.2 ug/L. No trend in chlorophyll.
Numeric Nutrient Criteria: SCI (62-302.531(2)(c), F.A.C.)	Avg SCI > 40 for at least 2 independent samples, with neither 2 most recent < 35	Avg SCI score > 40 with recent 2 > 35.
Numeric Nutrient Criteria: Nutrient Thresholds (62-302.531(2)(c), F.A.C.)	Annual Geometric Mean TP < 0.49 mg/L and TN < 1.65 mg/L, not exceeded more than once in 3 year period	TN < 1.65 mg/L, but TP > 0.49 mg/L. Passing NNC by SCI if more passing floral metrics collected in 2013.
Biological Health Assessment (62-303.330, F.A.C.) used for Planning and Verified Lists	2 recent SCI > 35 –AND- not 20 pts < Historic Max	Recent SCI scores > 35 and within 20 pts of Historic Max (65);

Parameter	Criteria for Passing	HCSW-1 Results
Chlorophyll for Planning List (62-303.351(3) and (4) F.A.C)	No algal mats or blooms and 2 of 3 Annual Geometric Mean Chlorophyll < 20 ug/L	No algal mats or blooms. All chlorophyll < 20 ug/L
Trends for Planning List, Study List, and Impaired List [62-303.351(5); 62-303.390(2)(a); 62-303.450(4) F.A.C.]	Statistically significant trend in annual geometric mean TP, TN, Chlorophyll using one-sided Mann's trend test with 95%. Planning list – 10 years of data. Study List – remove confounding variables and predicted impairment within 10 years. Verified list – trend on study list and predicted impairment within 5 years	No significant trends for TN and chlorophyll a from 2003-2012. OP (HCSP) and TP (SWFWMD) had statistically significant trend for 2003-2012 (p <0.04), but trend no significant when confounding variable of streamflow was accounted for. Could be listed on planning, but not study or verified list. No impairment.

Horse Creek fish populations at HCSW-1[3] and HCSW-4 show no evidence of declines during the HCSP study period through 2012 (Figures 20 and 21), despite the step-change in conductivity concentrations. Freshwater fish, or those species that are confined to freshwater, are part of the Cyprinidae, Catostomidae, Ictaluridae, Centrarchidae, and Percidae families (Peterson and Meador 1994). In general, a fish species is only as tolerant of changes in conductivity or salinity as their most sensitive life stage. The tolerance to salinity/conductivity at each life stage varies with the species. The main stress caused by salinity changes is the demand of maintaining an osmotic balance (Nordilie and Mirandi 1996). However, salinity is not the only factor influencing the survival of freshwater species. Several factors, including habitat complexity, predation, and prey availability, also influence growth and consequently survival (Peterson and Meador 1994). Other considerations, such as water temperature and suitable habitat (woody debris or macrophytes) may affect the taxa richness and abundance of freshwater fish in Horse Creek.

Freshwater fish can be found in a range of conductivities, but tend to have a preferred range based on the species. Of the common species (Seminole killifish, shiner species, and brook silverside) and game species (bass, bluegill, redear sunfish, and spotted sunfish) that have been collected within the lower portion of Horse Creek, the ideal range of conductivities was from 200-500 µmhos/cm (Call et.al. 2011). More than 90% of the HCSP conductivity measurements at HCSW-1 are below 500 µmhos/cm from 2003-2012, suggesting that conditions at HCSW-1 are well within the preferred range of most freshwater species. Over that time period, there was no correlation between the specific conductivity and number of freshwater fish species collected at HCSW-1 (Figure 21). At HCSW-4, specific conductivity concentrations were often above 500 µmhos/cm and conductivity concentrations found at HCSW-1, but a diverse suite of freshwater fish species were collected with no correlation with conductivity (Figure 21).

[3] Fish richness and diversity at HCSW-1 in 2010 was affected by higher than usual streamflow and gauge height during sampling that resulted in few habitat refuges for fish at the HCSW-1 sampling location, as well as record cold temperatures that were responsible for increases in regional fish mortality.

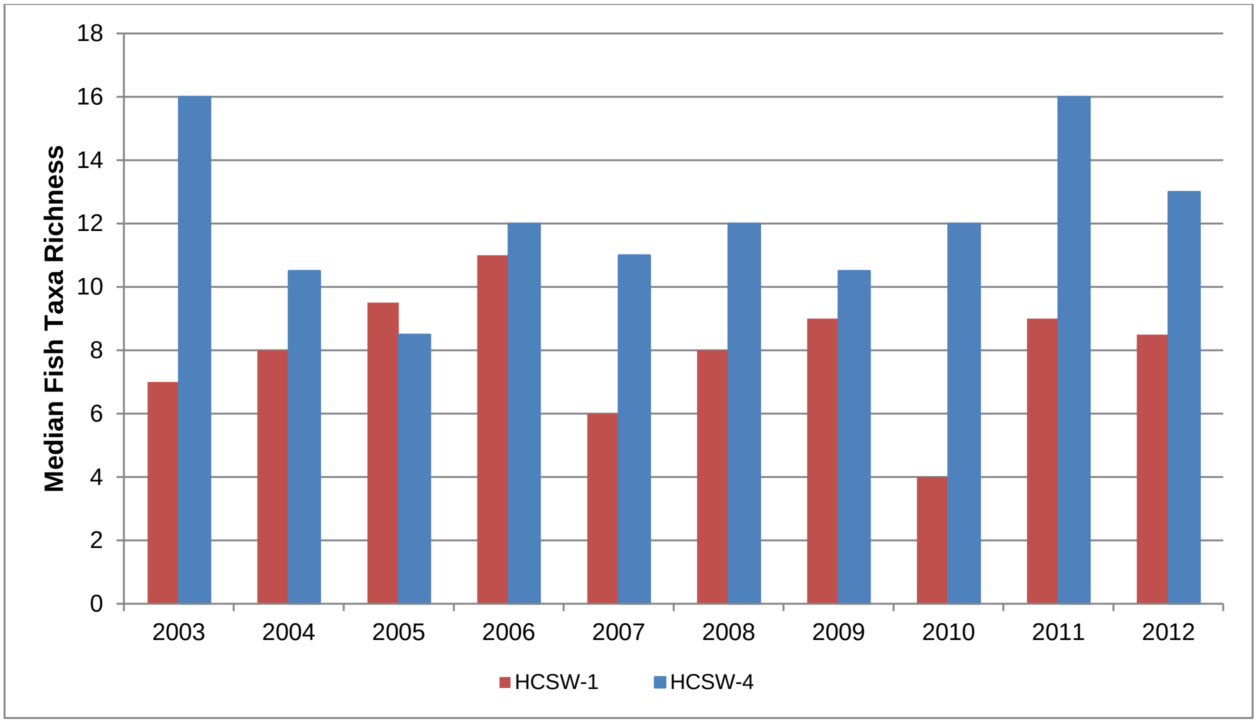

Figure 20. **Annual Median Fish Taxa Richness at HCSW-1 and HCSW-4 Collected During HCSP.**

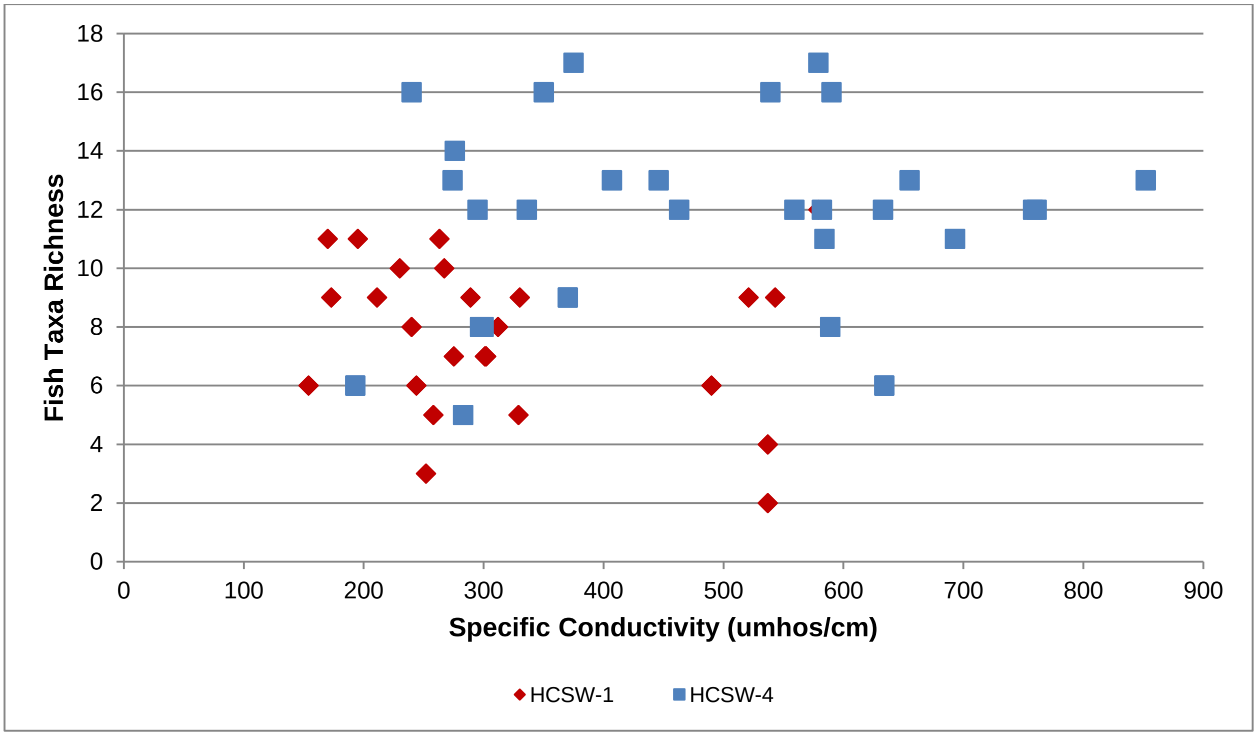

Figure 21. **Specific Conductivity versus Number of Freshwater Fish Species at HCSW-1 and HCSW-4 Collected During HCSP from 2003-2012.**

Overall, Horse Creek macroinvertebrate communities at HCSW-1 and HCSW-4 are healthy according to the Florida SCI, with no evidence of declines over time during the HCSP study period (Figure 22). Comparing the SCI scores for HCSW-1, HCSW-3, and HCSW-4[4] to the three-month average[5] conductivity shows no directional relationship or step-change in SCI over the range of conductivity seen in Horse Creek (Figure 23).

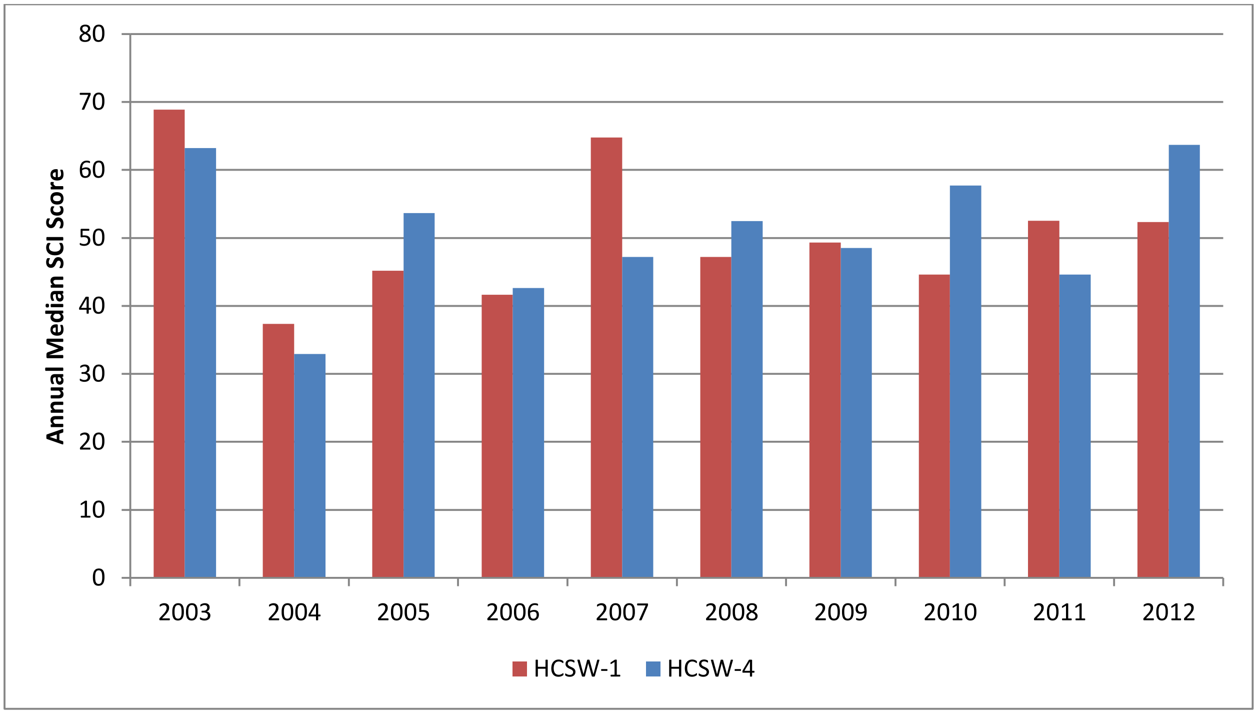

Figure 22. Annual Median SCI Score at HCSW-1 and HCSW-4 Collected During HCSP[6].

[4] HCSW-2 was omitted from this analysis because physical constraints at that station (low flow and dissolved oxygen) have a greater effect on the invertebrate community than any potential influence by conductivity.

[5] We tried several conductivity values for this analysis: conductivity at sampling, three-month average, three-month maximum, six-month average, and six-month maximum. The three-month average conductivity was representative of all of the potential relationships. Three months (or 90 days) is the minimum streamflow requirement for conducting SCI within a stream.

[6] Changes in SCI SOP around 2007 may mean that 2008-2012 SCI scores are not directly comparable to those collected previously.

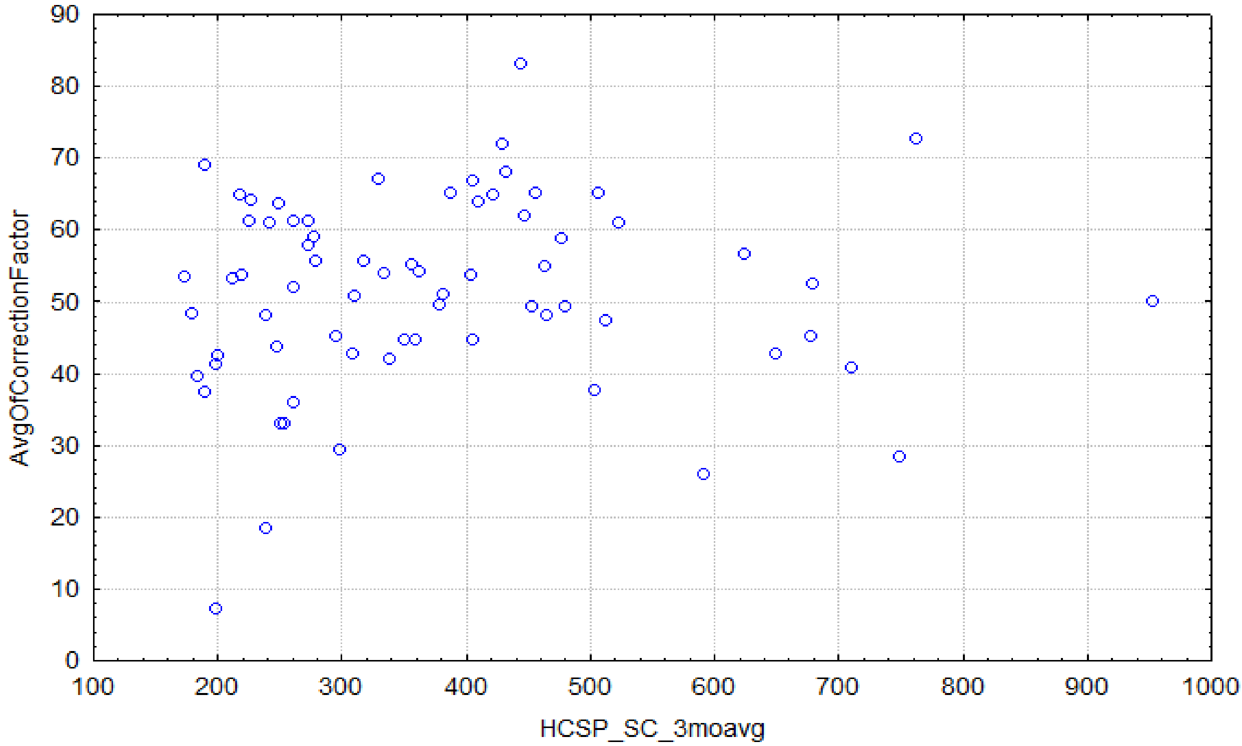

Figure 23. Specific Conductivity (3 month average) versus SCI Score at HCSW-1, HCSW-3, and HCSW-4 Collected During the HCSP from 2003-2012.

Conclusions

In the 2012 Annual Report, the Seasonal Kendall Tau procedure found several statistically significant trends in the water quality data from 2003 to 2012 (Table 1), thus triggering this impact assessment. Seven water quality parameters had a statistically significant trend at HCSW-1 only and three parameters at both HCSW-1 and HSW-4. Several of the trends detected during the statistical analysis have an estimated slope that 1) was not in the direction of an adverse trend (color and iron) or 2) was very small compared to limits in laboratory prediction or the observed differences between primary and duplicate field samples (pH, ammonia, fluoride, orthophosphate). For the trends with higher estimated rates of change (specific conductivity and various dissolved ions), the potential trends are further discussed in this impact assessment, with a focus on the station closest to mining (HCSW-1). Specific conductivity, which has a longer period of record with more consistent data collection, is used as a surrogate for the other dissolved ions (calcium, alkalinity, sulfate, and TDS) in this impact assessment. Orthophosphate was also discussed further in this impact assessment as a follow-up to the trend impact assessment of the 2010 Annual Report (Robbins et al. 2014).

Expanding the period of record for the trend analysis of orthophosphate at HCSW-1 showed that the orthophosphate trend was an artifact of the time period used in the original analysis, and concentrations of orthophosphate in recent years are similar to those measured prior to the start of the HCSP. This is supported by the very small slope estimate found in the original 2010 trend analysis (Table 1). Therefore, there is no increasing trend in orthophosphate in Horse Creek that would necessitate any corrective action related to mining activities. In addition, data collected at HCSW-1 for nutrients and biological health indicates that Horse Creek meets water quality standards (both the existing narrative standard, and the recently approved Florida numeric nutrient criteria), because there is no evidence of an imbalance of flora and fauna.

For specific conductivity, a significant increasing trend was detected at HCSW-1 even when the period of record was expanded, but the inclusion of data through 2012 makes it clear that the most visible increase is a single step-change rather than a monotonic, continuing increase. In addition, at least part of the increase may be caused by regional influences, as Charlie Creek also shows a slow increase in conductivity over time. In addition, specific conductivity at Horse Creek and West Fork Horse Creek stations upstream of the NPDES outfalls has also been increasing over the same time period (data found in Mosaic DRI reports to Manatee County). Specific conductivity at HCSW-1 began to rise during a very dry period in 2006-2008 when Mosaic had very little NPDES discharge into Horse Creek, and the stream was dominated by baseflow (surficial aquifer) contributions. After the Horse Creek outfalls began to receive water decanting from Wingate Mine clays, the specific conductivity remained at the higher levels, presumably because of the greater proportion of groundwater involved in the dredge mining. If groundwater influence is the main cause of increased specific conductivity at HCSW-1, then concentrations have reached a threshold and should not continue to increase. This theory has been validated considering that the range and median specific conductivity from 2009 to 2012 has been very consistent with no expectation that it will increase from the recently observed range in the future.

Dissolved ion concentrations at HCSW-1 in recent years still meet all applicable state drinking water and Class III surface water standards, except where otherwise noted in previous monthly impact assessments. In addition, the current specific conductivity concentrations at HCSW-1 are within the preferred range of Horse Creek freshwater fish species, and the diversity and composition of fish populations have remained steady over the HCSP study period with no correlation with specific conductivity. SCI Scores, an indicator of benthic Macroinvertebrate community health, have remained steady over the HCSP study period and show no relationship with specific conductivity.

The HCSP plan document requires that an “impact assessment” be conducted for any trigger level exceedance or adverse water quality trends found while preparing the annual HCSP report. If the impact assessment indicates or suggests that mining activities by Mosaic are the cause of the adverse exceedance or trend, then Mosaic needs to take corrective action. The intensity of the corrective action will be based on the potential for short-term or long-term ecological harm to Horse Creek and/or the integrity of the downstream potable water supply. In this impact assessment, we have concluded that some of the trends found in the 2012 Annual Report (pH, ammonia, color, iron, and orthophosphate) are not adverse and require no corrective action. The orthophosphate trend was found to be a result of data bias and is not a valid trend that would require corrective action. The change in specific conductivity is a step-change instead of a monotonic trend with a very low probability of further increases over time. It has been potentially influenced by increases in groundwater contributions during drought conditions, changes in mining operations and mine water management, and possibly other regional climatic or hydrological factors. The biological effects of this step-increase in conductivity should be minimal, given that more than 90% of the concentrations at HCSW-1, the station closest to mining, are within the preferred conductivity range of freshwater fish, and all recorded conductivities are within their tolerance. Invertebrate SCI scores also show no effect of conductivity changes at HCSW-1. At this time, we do not recommend corrective action beyond continued monitoring. This will ensure that existing biological quality is preserved in upper Horse Creek.

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APPENDIX J

COMMENTS ON HCSP SCI DATA

Appendix J
Comments on HCSP SCI Data

Beginning with the 2010 annual report, we have reevaluated the HCSP SCI data with strict interpretation of FDEP SOP guidance, including the upper and lower limits of the SOP target number of individuals, the SOP target of 90 days of previous flow, and the SOP target of less than a 0.5 m water level increase in the previous 30 days. As a result of this evaluation, some SCI scores have been removed from the analysis (in red italics). In addition, some SCI results with more than the target number of individuals were randomly resampled to provide an unbiased result.

Date	HCSW-1			HCSW-2			HCSW-3			HCSW-4		
	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments
4/25/2003	*134*	*62.99*	Less than 90 days of flow	134	52.41	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	142	51.11	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	147	61.84	
7/29/2003	*141*	*54.69*	Greater than 0.5m water level increase over previous 30 days	*139*	*13.59*	Greater than 0.5m water level increase over previous 30 days	*151*	*27.17*	Less than SOP target number of individuals; Greater than 0.5m water level increase over previous 30 days	*146*	*61.50*	Less than SOP target number of individuals; Greater than 0.5m water level increase over previous 30 days
11/20/2003	133	65.44		121	36.81	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	131	61.23		135	61.30	
4/22/2004	138	37.36		134	26.78	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	138	*34.37*	Less than SOP target number of individuals	141	*56.94*	Less than SOP target number of individuals
11/3/2004	NA	*58.19*	Less than SOP target number of individuals	117	5.17		99	*23.62*	Less than SOP target number of individuals	111	32.95	
2/15/2005	131	48.12	Randomly subsampled because original	117	62.19	Randomly subsampled because original sort had more	112	50.72	Randomly subsampled because original sort had more than the	113	54.34	Randomly subsampled because original sort had more than the

Date	HCSW-1			HCSW-2			HCSW-3			HCSW-4		
	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments
			sort had more than the recommended number of individuals from FDEP SOP			than the recommended number of individuals from FDEP SOP			recommended number of individuals from FDEP SOP			recommended number of individuals from FDEP SOP
4/20/2005	126	18.45	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	117	39.95		124	59.04		121	67.09	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP
9/15/2005[1]	129	42.40	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	124	21.00	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	121	53.32	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	114	53.06	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP
12/15/2005	130	48.31	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	114	37.29	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	115	41.35		115	36.07	
4/6/2006	110	43.70	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	98	25.14		103	59.65		105	45.65	
7/27/2006	*115*	*59.45*	Less than 90 days of flow	*106*	*26.66*	Less than 90 days of flow; Less than SOP target number of individuals	*118*	*32.34*	Less than 90 days of flow	127	49.81	

Date	HCSW-1			HCSW-2			HCSW-3			HCSW-4		
	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments
11/28/2006[2]	115	39.54		93	33.73		121	42.52		113	42.17	
3/28/2007	115	65.42	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	100	32.29	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	117	55.04	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	113	50.08	
8/9/2007	123	65.06		--	--	No flow - no samples collected	121	28.54	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	130	40.95	
11/27/2007	116	64.89		108	22.07		116	65.05	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	124	60.96	
4/24/2008	*101*	*47.34*	Less than 90 days of flow	109	22.72		114	47.97		104	52.43	
9/12/2008	122	45.12	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	104	8.56		121	7.12		119	33.07	
11/19/2008	115	48.06	Randomly subsampled because original sort had more than the recommended number of individuals from FDEP SOP	84	24.74		109	29.27		108	55.81	

Date	HCSW-1			HCSW-2			HCSW-3			HCSW-4		
	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments	HA Score	SCI Score	Comments
4/22/2009	--	--	No flow - no samples collected	--	--	No flow - no samples collected	--	--	No flow - no samples collected	105	45.13	
10/22/2009	124	49.36		123	21.73		106	53.78		114	52.07	
4/20/2010	126	36.85		115	29.42		103	58.65		110	67.92	
9/28/2010	128	55.17		102	10.63		99	64.92		109	58.08	
11/4/2010 (or 11/11/10)	119	44.63		105	32.38		100	64.19		105	54.87	
4/18/2011	127	55.71		102	20.28		103	66.72		113	83.38	
8/9/2011	--	--	Less than 90 days of flow	--	--	Less than 90 days of flow	112	--	Less than 90 days of flow	122	25.81	
10/26/2011	110	49.48		--	--	Stream flooded - no samples collected	109	61.13		116	44.62	
3/30/2012	--	--	Low water levels - no samples collected	--	--	Dry - no samples collected	--	--	Very little to no flow - no samples collected	121	72.86	
10/26/2012	126	53.80		--	--	High water levels - no samples collected	118	61.00		97	63.58	
12/12/2012	120	51.39		--	--	No flow - no samples collected	104	72.09		103	62.07	

[1] Sorting method change in FDEP SOP

[2] Sorting and calculation method change in FDEP SOP; two vial average

APPENDIX

K

SUMMARY OF MILESTONES DURING THE HCSP

Appendix K
Summary of Milestones During the HCSP

K.1 Events Timeline

April 2003 – HCSP began.

August 2004 – Hurricane Charley moves up the Horse Creek Basin.

September 2004 – Hurricane Frances moves up the Horse Creek Basin.

September 2004 – Hurricane Jeanne moves up the Horse Creek Basin.

August 2005 – Invertebrate sorting methodology change in FDEP SCI SOP. Target number of individuals between 100 and 120 per sample.

October 2005 – USGS rain gauge discontinued at HCSW-1. Began using SWFWMD rain gauge 494 for annual reports.

June 2006 – The last clays from Fort Green beneficiation plant were sent to clay settling areas (CSAs) FGH3 and FGH4 which discharge to Horse Creek via FTG-003 and FTG-004.

November 2006 – Invertebrate sorting methodology change in FDEP SCI SOP. Two vials with a target number of individuals of 140-160 per sample are required. The average SCI score of the two vials is used for reporting purposes.

2006 – 2008 – Time period with lower than average streamflow and rainfall for the Horse Creek Basin.

July 2006 - September 2008 – Very little NPDES discharge (stormwater and baseflow only) from FTG-003 and FTG-004 due to extremely dry conditions.

October 2008 – Clays mined via dredge from the Wingate Mine began to be transported to facilities and FM1 in the Horse Creek basin for processing and storage. NPDES discharge was comprised mostly of groundwater from the Wingate mining process.

March 2009 – Added CSA FM-1 to existing monitoring program.

September 2009 – discontinue monitoring FL-PRO, fatty acids, and total amines at all four Horse Creek locations. Sampling began in Brushy Creek (BCSW-1) minus trigger levels and impact assessments.

Winter 2009/2010 – Florida experienced one of the coldest winters on record (December-February the 10th coldest period in Tampa since records started in 1890). In Hillsborough County, overnight lows in early January were at or below freezing for 12 consecutive nights. Cold temperatures led to large fish kill in the area as a result.

December 2010 – Coldest December for the Tampa Bay area in recorded history (the daily average [53.2°C] was 10°C lower for the month than normal). Several areas throughout west-central and southwest Florida also set record lows.

October 2011 – SWFWMD reduced sampling frequency at HCSW-1 and HCSW-4 to every other month from monthly sampling.

November 2011 – SWFWMD rain gaugue 494 discontinued. Began using NOAA gauge.

K.2 Lab Changes Timeline

April 2003 – November 2004: Various labs

December 2004 – May 2008: STL/Test America (all but Radiologicals)

April 2006 – July 2008: KNL Labs (Radiologicals only)

July 2008 – July 2010: Benchmark Analytical (still analyzing color and chlorophyll-a)

August 2008 – Present: Florida Radiochemistry (Radiologicals only)

August 2010 – Present: Mosaic's Laboratory

K.3 Major MDL Changes

January 2006 – July 2008: Nitrate+Nitrite highly variable

April 2003 – December 2011: Ammonia (around 0.03mg/L through October 2007, variable through July 2008, stable through July 2011, then variable)

December 2007: Orthophosphate abnormally high value (0.75mg/L)

April 2003 – December 2011: Dissolved iron started at 0.1mg/L, reduced in March 2006 to 0.022mg/L, stable from August 2010 at around 0.01mg/L

March 2006 – February 2008: Chloride numerous changes ranging from 0.022-30mg/L; stable since March 2008

March 2006 – February 2008: Fluoride numerous changes ranging from 0.017-5mg/L; relatively stable since March 2008

March 2006 – February 2008: Sulfate numerous changes; stable since March 2008